# INVESTIGATION OF THE ATMOSPHERIC OZONE FORMATION POTENTIAL OF DIMETHYL SULFOXIDE 

Report to the<br>Gaylord Chemical Corporation<br>by<br>William P. L. Carter, Dongmin Luo, and Irina L. Malkina

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College of Engineering
Center for Environmental Research and Technology
University of California
Riverside, California 92521


#### Abstract

A series of environmental chamber experiments and computer model calculations were carried out to assess the atmospheric ozone formation potential of dimethyl sulfoxide (DMSO). The experiments consisted of measuring ozone formation, NO oxidation and DMSO consumption rates in irradiations of DMSO - $\mathrm{NO}_{\mathrm{x}}$ mixtures and determining the effects of DMSO on $\mathrm{O}_{3}$ formation, NO oxidation and integrated OH radical levels when added to various simulated photochemical smog systems. The results indicated that DMSO is highly reactive towards ozone formation under all conditions examined. High yields of formaldehyde were observed, and approximately half of the sulfur in DMSO reacts to form products that are not detected by a total gas phase sulfur analyzer (i.e., not $\mathrm{SO}_{2}$ ). In addition, an upper limit rate constant of $3 \times 10^{-20} \mathrm{~cm}^{3}$ molec ${ }^{-1} \mathrm{~s}^{-1}$ was determined for the reaction of DMSO with $\mathrm{O}_{3}$. The information available from previous studies is not sufficient to determine the mechanism for DMSO's atmospheric reactions, and a number of alternative mechanisms were examined for consistency with the data obtained in this study. The best results are obtained using a mechanism where $75 \%$ of the reaction of DMSO with OH results in the formation of $\mathrm{SO}_{2}$ and two formaldehyde molecules after conversions of two molecules of NO to $\mathrm{NO}_{2}$; with the remaining $25 \%$ involving the formation of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{CH}_{3}\left(\mathrm{DMSO}_{2}\right)$ and $\mathrm{HO}_{2}$. Although this mechanism underpredicts the effects of DMSO on NO oxidation and $\mathrm{O}_{3}$ formation in some experiments, it generally gives good simulations of the experiments most closely representing polluted urban atmospheres. This mechanism predicted that DMSO emissions form about twice as much ozone on a mass basis then emissions of the mixture of reactive VOCs representing emissions from all sources.


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## INTRODUCTION

Ozone in photochemical smog is formed from the gas-phase reactions of volatile organic compounds (VOCs) and oxides of nitrogen $\left(\mathrm{NO}_{\mathrm{x}}\right)$ in sunlight. Although Houston and Los Angeles have one of the worst ozone problems in the United States, other areas of the country also have episodes where ozone exceeds the federal air quality standard. Ozone control strategies in the past have focused primarily on VOC controls, though the importance of $\mathrm{NO}_{x}$ control has become recognized in recent years. VOC and $\mathrm{NO}_{\mathrm{x}}$ controls have differing effects on ozone formation. $\mathrm{NO}_{\mathrm{x}}$ is required for ozone formation, and if the levels of $\mathrm{NO}_{\mathrm{x}}$ are low compared to the levels of reactive VOCs, then changing VOC emissions will have relatively little effect on ozone. Since $\mathrm{NO}_{\mathrm{x}}$ is removed from the atmosphere more rapidly than VOCs, ozone in areas far downwind from the primary sources tend to be more $\mathrm{NO}_{\mathrm{x}}$ limited, and thus less responsive to VOC controls. VOC controls tend to reduce the rate that $\mathrm{O}_{3}$ is formed when $\mathrm{NO}_{\mathrm{x}}$ is present, so VOC controls are the most beneficial in reducing $\mathrm{O}_{3}$ in the urban source areas, where $\mathrm{NO}_{\mathrm{x}}$ is relatively plentiful, and where $\mathrm{O}_{3}$ yields are determined primarily by how rapidly it is being formed. Because of this, any comprehensive ozone control strategy should involve reduction of emissions of both $\mathrm{NO}_{\mathrm{x}}$ and VOCs.

Many different types of VOCs are emitted into the atmosphere, each reacting at different rates and having different mechanisms for their reactions. Because of this, they can differ significantly in their effects on ozone formation, or their "reactivity". Some compounds, such as CFCs, do not react in the lower atmosphere at all, and thus make no contribution to ground-level ozone formation. Others, such as methane, react and contribute to ozone formation, but react so slowly that their practical effect on ozone formation in urban atmospheres is negligible. Obviously, it does not make sense to regulate such compounds as ozone precursors. In recognition of this, the EPA has exempted certain compounds from such regulations on the basis of having "negligible" effects on ozone formation. Although the EPA has no formal policy on what constitutes "negligible" reactivity, in practice it has used the ozone formation potential of ethane as the standard in this regard. This is because ethane is the most reactive of the compounds that the EPA has exempted to date. Therefore, the ozone formation potential of a compound relative to ethane is of particular interest when assessing whether it might be a likely candidate for exemption from regulation as an ozone precursor.

Many VOCs that would not be judged to have "negligible" reactivity under the current criterion might still have much lower ozone formation potential than average, and substituting emissions of highly reactive VOCs with such moderate-to-low reactivity VOCs would be expected to result in air quality improvements. Although the current EPA policies do not encourage such substitutions, it has been proposed to implement reactivity-based policies on a voluntary basis in consumer product regulations in California (CARB, 1999), and the EPA is currently re-evaluating its reactivity-based VOC policies (Dimitriades, 1999, RRWG, 1999). Mc.Bride et al (1977) showed that adopting reactivity-based VOC control policies could result in significant cost savings in ozone reduction strategies, though a number of
difficult policy and enforcement issues need to be resolved (RRWG, 1999). Although regulatory approaches that appropriately deal with differences in VOC reactivity are still evolving, it is clear that producers of solvent VOCs will need to know how their VOCs might be classified under any such system, so they can appropriately adapt to reactivity-based policies once they are implemented. This requires an ability to reliably estimated the ozone impacts of the VOCs of interest.

Dimethyl sulfoxide ( $\mathrm{DMSO}, \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{3}$ ) is an important solvent compound that is manufactured by Gaylord Chemical Company. Since the atmospheric ozone impact of DMSO has not previously been assessed, Gaylord contracted us to carry out a preliminary evaluation of its likely range of ozone impacts, and the possibility that it may have sufficiently low ozone impact that it might appropriately be exempted as an ozone precursor. The results of this assessment (Carter, 1997) concluded that DMSO reacts in the atmosphere too rapidly to be exempted on the basis of low reaction rate, but that the mechanism for its atmospheric reactions is highly uncertain. In particular, the possibility existed that DMSO might react in a way that actually inhibits ozone formation, depending on how some of the sulfurcontaining intermediate radicals react under atmospheric conditions. If this were the case, it would not be appropriate to regulate emissions of DMSO as an ozone precursor.

Because of this uncertainty, Gaylord Chemical contracted the College of Engineering Center for Environmental Research and Technology (CE-CERT) to obtain the data needed to better quantify the ozone formation potential of DMSO, and to determine whether it might inhibit ozone formation under any atmospheric conditions. This involved conducting environmental chamber experiments to determine the effects of DMSO on ozone formation under representative atmospheric conditions, developing a mechanism for the atmospheric reactions of DMSO that is consistent with these data and results of previous kinetic and mechanistic studies, and then using this mechanism to obtain quantitative estimates for the ozone formation potential of DMSO under a range of atmospheric conditions. The results of this program are documented in this report.

## EXPERIMENTAL AND DATA ANALYSIS METHODS

## Overall Experimental Approach

Most of the experiments for this program consisted of conducting environmental chamber experiments where DMSO reacted under simulated atmospheric conditions, to provide data to test whether chemical mechanisms could correctly predict the effects of DMSO's reactions on ozone formation and other measures of reactivity. Two general types of experiments with DMSO were carried out: DMSO - $\mathrm{NO}_{\mathrm{x}}$ - air irradiations and incremental reactivity experiments with DMSO. These are discussed below. In addition, several experiments were carried out to determine the upper limit for the rate constant of DMSO with $\mathrm{O}_{3}$, to determine if this needed to be considered in models of DMSO's atmospheric reactivity.

The DMSO - $\mathrm{NO}_{\mathrm{x}}$ - air experiments were carried out to provide data for mechanism evaluation under simpler chemical conditions. These consisted of irradiations of DMSO in the presence of $\mathrm{NO}_{\mathrm{x}}$ in air without other reactants. Such experiments do not represent realistic atmospheric conditions because of the lack of other pollutants that are present in real atmospheres, and they do not provide useful data for compounds that are radical inhibitors (Carter et al, 1982, Carter and Lurmann, 1991). However, for sufficiently reactive compounds they can provide useful data for mechanism testing complications and uncertainties involved with modeling the reactions of the other organics present in more realistic experiments. These were included in this project once the results of the reactivity experiments, discussed below, indicated that DMSO was apparently sufficiently reactive that such experiments should provide useful data.

Most of the chamber experiments for this program consisted of measurements of "incremental reactivities" of DMSO under various conditions. These involve two types of irradiations of model photochemical smog mixtures. The first is a "base case" experiment where a mixture of reactive organic gases (ROGs) representing those present in polluted atmospheres (the "ROG surrogate") is irradiated in the presence of oxides of nitrogen $\left(\mathrm{NO}_{\mathrm{x}}\right)$ in air. The second is the "test" experiment that consists of repeating the base case irradiation except that the VOC whose reactivity is being assessed is added. The differences between the results of these experiments provide a measure of the atmospheric impact of the test compound, and the difference relative to the amount added is a measure of its reactivity.

To provide data concerning the reactivities of the test compound under varying atmospheric conditions, three types of reactivity experiments were carried out:

Mini-Surrogate Experiments. This base case employed a simplified ROG surrogate and relatively low $\mathrm{ROG} / \mathrm{NO}_{\mathrm{x}}$ ratios. Low $\mathrm{ROG} / \mathrm{NO}_{\mathrm{x}}$ ratios represent "maximum incremental reactivity" (MIR) conditions, which are most sensitive to VOC effects. This is useful because it provides a sensitive test for the model, and also because it is most important that the model correctly predict a VOC's reactivity under
conditions where the atmosphere is most sensitive to the VOCs. The ROG mini-surrogate mixture employed consisted of ethene, n-hexane, and m-xylene. This surrogate was employed in our previous studies (Carter et al, 1993; 1995a-c, 1997, 2000), and was found to provide a more sensitive test of the mechanism than the more complex surrogates which more closely represent atmospheric conditions (Carter et al, 1995b). This high sensitivity to mechanistic differences makes the mini-surrogate experiments most useful for mechanism evaluation.

Full Surrogate Experiments. This base case employed a more complex ROG surrogate under somewhat higher, though still relatively low, $\mathrm{ROG} / \mathrm{NO}_{\mathrm{x}}$ conditions. While less sensitive to the mechanism employed, experiments with a more representative ROG surrogate are needed to evaluate the mechanism under conditions that more closely resembling the atmosphere. The ROG surrogate employed was the same as the 8 -component "lumped molecule" surrogate employed in our previous study (Carter et al. 1995b), and consists of n-butane, n-octane, ethene, propene, trans-2-butene, toluene, m-xylene, and formaldehyde. Calculations have indicated that use of this 8 -component mixture will give essentially the same results in incremental reactivity experiments as actual ambient mixtures (Carter et al. 1995b).

Full Surrogate, low $\mathrm{NO}_{\underline{x}}$ Experiments. This base case employing the same 8-component "lumped molecule" surrogate as the full surrogate experiments described above, except that lower $\mathrm{NO}_{\mathrm{x}}$ levels (higher $\mathrm{ROG} / \mathrm{NO}_{\mathrm{x}}$ ratios) were employed to represent $\mathrm{NO}_{\mathrm{x}}$-limited conditions. Such experiments are necessary to assess the ability of the model to properly simulate reactivities under conditions where $\mathrm{NO}_{\mathrm{x}}$ is low. The initial ROG and $\mathrm{NO}_{\mathrm{x}}$ reactant concentrations were comparable to those employed in our previous studies (Carter et al. 1995b).

An appropriate set of control and characterization experiments necessary for assuring data quality and characterizing the conditions of the runs for mechanism evaluation were also carried out. These are discussed where relevant in the results or modeling methods sections.

## Environmental Chamber Experiments

## Chamber Employed

The experiments were carried out using the CE-CERT "Dividable Teflon Chamber" (DTC) with a blacklight light source. This consists of two $\sim 6000$-liter 2-mil heat-sealed FEP Teflon reaction bags located adjacent to each other and fitted inside an $8^{\prime} \times 8^{\prime} \times 8^{\prime}$ framework that has two diametrically opposed banks of 32 Sylvania 40-W BL black lights that serve as the light source. The lighting system in the DTC was found to provide so much intensity that only half the lights were used for irradiation. The air conditioner for the chamber room was turned on before and during the experiments. Four air blowers which are located in the bottom of the chamber were used to help cool the chamber as well as mix the contents of the chamber. The CE-CERT is described in more detail elsewhere (Carter et al, 1995c).

The DTC is designed to allow simultaneous irradiations of experiments with and without added test reactants under the same reaction conditions. Since the chamber is actually two adjacent FEP Teflon
reaction bags, two mixtures can be simultaneously irradiated using the same light source and with the same temperature control system. These two reaction bags are referred to as the two "sides" of the chambers (Side A and Side B) in the subsequent discussion. The sides are interconnected with two ports, each with a box fan, which rapidly exchange their contents to assure that base case reactants have equal concentrations in both sides. In addition, a fan is located in each of the reaction bags to rapidly mix the reactants within each chamber. The ports connecting the two reactors can then be closed to allow separate injections on each side, and separate monitoring of each side.

The blacklight light source has the advantage of being relatively inexpensive to operate and provides a reasonably good simulation of natural sunlight in the region of the spectrum that is important in affecting most photolysis reactions of importance for non-aromatic VOCs (Carter et al, 1995d). This is therefore appropriate for studies of reactivities of compounds that are not photoreactive or believed to form significant yields of photoreactive products whose action spectra are not well characterized. This is believed to be the case for DMSO.

## Procedures

The reaction bags were flushed with dry air produced by an AADCO air purification system for 14 hours ( $6 \mathrm{pm}-8 \mathrm{am}$ ) on the nights before experiments. The continuous monitors were connected prior to reactant injection and the data system began logging data from the continuous monitoring systems. The reactants were injected as described below. The common reactants were injected in both sides simultaneously using a 2 feet long Pyrex tube (with the outlet connected to a " Y "-shape glass tube that was connected to side A and B respectively in the injection line and were well mixed before the chamber was divided. The contents of each side were blown into the other using two box fans located between them. Mixing fans were used to mix the reactants in the chamber during the injection period, but these were turned off prior to the irradiation. The sides were then separated by closing the ports that connected them, after turning all the fans off to allow their pressures to equalize. After that, reactants for specific sides (the test compound in the case of reactivity experiments) were injected, mixed, and analyzed. The lights were then turned on and the irradiation proceeded for 6 hours. After the run, the contents of the chamber were emptied by allowing the bags to collapse, and then the chamber was flushed with purified air. The contents of the reactors were vented into a fume hood.

The procedures for injecting the various types of reactants were as follows. The NO and $\mathrm{NO}_{2}$ were prepared for injection using a high vacuum rack. Known pressures of NO, measured with MKS Baratron capacitance manometers, were expanded into Pyrex bulbs with known volumes, which were then filled with nitrogen (for NO ) or oxygen (for $\mathrm{NO}_{2}$ ). The contents of the bulbs were then flushed into the chamber with nitrogen. The gaseous reactants were prepared for injection either using a high vacuum rack or a gas-tight syringes whose amounts were calculated. The gas reactants in a gas-tight syringe were usually diluted to $100-\mathrm{ml}$ with nitrogen in a syringe. The volatile liquid reactants were injected, using a micro syringe, into a 2 ft long Pyrex injection tube surrounded with heat tape and equipped with one port for the injection of the liquid and four ports to attach bulbs with gas reactants Then one end of the
injection tube was attached to the " $Y$ "-shape glass tube (equipped with stopcocks) that was connected to both sides of the chamber and the other to a nitrogen source. To introduce all the reactants into the chamber simultaneously gas and liquid reactants were injected at the same time. The stopcocks were then opened, and the contents of the bulbs were flushed into the chamber with a combination of nitrogen and heating (injection tube was surrounded with heat tape) for approximately 5 minutes. Formaldehyde was prepared in a vacuum rack system by heating paraformaldehyde in an evacuated bulb until the pressure corresponded to the desired amount of formaldehyde. The bulb was then closed and detached from the vacuum system and its contents were flushed into the chamber with nitrogen through the injection port.

In case of first three experiments DMSO was prepared using a high vacuum rack, using a similar injection procedure as described for formaldehyde, above. This was found to give less than irreproducible amounts of DMSO in the gas phase. Because of this, both the amounts of DMSO injected into the chamber during the experiments and the calibration factors for the DMSO analyses by gas chromatography were uncertain. In the subsequent experiments the desired quantity of the liquid DMSO was injected with a micro syringe into preheated to 110 C Pyrex injection tube. The tube was then flushed into the chamber with nitrogen at 4 liters/minute for about 10 minutes. This was found to give more satisfactory results. The DMSO injections and calibrations during this subsequent period were also verified by using an independent determination using a total gas-phase sulfur analyzer, as discussed below.

Because of the uncertainties in the DMSO injection and analysis methods and the lack of verification of the initial DMSO measurements for the first three experiments with DMSO, the results of these experiments were not used for mechanism evaluation.

## Upper Limit $\mathrm{O}_{3}$ Rate Constant Determination Experiments

The upper limit $\mathrm{O}_{3}+$ DMSO rate constant experiments were based on monitoring the rates of consumption (or lack thereof) of DMSO in the presence of excess $\mathrm{O}_{3}$. These experiments were carried out using a "pillow-shaped" ~ 330 liters 2-mil heat-sealed FEP Teflon reaction bag covered with black material. The temperature was monitored by a thermocouple, and was $294 \pm 1^{\circ} \mathrm{K}$ for all experiments.

Several different procedures were used as discussed in the Results section, but in the most useful experiments approximately 50 ppm of $\mathrm{O}_{3}$ was injected first, and then approximately 100 ppm of cyclohexane was added to serve as a sink for any OH radicals that may be formed. The ozone was made by flushing purified dry air through the quartz tube (ozone generator) into the chamber and monitored by the analyzer until the desired amount of ozone formed. Approximately $0.2-0.4 \mathrm{ppm}$ of DMSO was then injected and its concentration was monitored for several hours in the dark. The injection procedures for the DMSO the alkane were as employed in most of the chamber experiments, as described above. After the run the reaction bag was emptied by allowing it to collapse and then filled with the purified air. This procedure was repeated three times.

Ozone was monitored using a Monitor Labs model M-8410 chemiluminescence ozone analyzer, which works on principle of chemiluminescence from the reaction between ozone and ethylene. This instrument was calibrated at low ozone concentrations, and the accuracy of the data at high concentrations of ozone is uncertain. However, it is unlikely that the instrument would inaccurate more than $10 \%$.

The cyclohexane was added to scavenge the OH because if OH were formed from the reactions of $\mathrm{O}_{3}$ with DMSO or background materials or the walls of the chamber, then it would cause consumption of DMSO due to reaction with OH rather than $\mathrm{O}_{3}$.

## Analytical Methods

Ozone and nitrogen oxides $\left(\mathrm{NO}_{\mathrm{x}}\right)$ were continuously monitored using commercially available continuous analyzers with Teflon sample lines inserted directly into the chambers. The sampling lines from each side of the chamber were connected to solenoids that switched from side to side every 10 minutes, so the instruments alternately collected data from each side. Ozone was monitored using a Dasibi 1003-AH UV photometric ozone analyzer and NO and total oxides of nitrogen (including $\mathrm{HNO}_{3}$ and organic nitrates) were monitored using a Teco Model 42 chemiluminescent $\mathrm{NO} / \mathrm{NO}_{\mathrm{x}}$ monitor. The output of these instruments, along with that from the temperature sensors and the formaldehyde instrument, were attached to a computer data acquisition system, which recorded the data at 10 minutes intervals for ozone, $\mathrm{NO}_{\mathrm{x}}$ and temperature (and at 15 minutes for formaldehyde), using 30 second averaging times. This yielded a sampling interval of 20 minutes for taking data from each side.

The Teco instrument and Dasibi CO analyzer were calibrated with a certified NO and CO source and CSI gas-phase dilution system. It was done prior to chamber experiment for each run. The $\mathrm{NO}_{2}$ converter efficiency check was carried out in regular intervals. The Dasibi ozone analyzer was calibrated against transfer standard ozone analyzer using transfer standard method in a interval of three months and was check with CSI ozone generator for each experiment to assure that the instrument worked properly. The details were discussed elsewhere (Carter et al, 1995c)

Organic reactants other than formaldehyde were measured by gas chromatography with FID detection as described elsewhere (Carter et al. 1993; 1995c). GC samples were taken for analysis at intervals from 20 minutes to 30 minutes either using 100 ml gas-tight glass syringes or by collecting the 100 ml sample from the chamber onto Tenax-GC solid adsorbent cartridge. The Tenax sampling method was used for DMSO but the syringe sampling method was used for the primary analysis method for the other organic reactants monitored by GC in these experiments. These samples were taken from ports directly connected to the chamber after injection and before irradiation and at regular intervals after irradiation was started. The sampling method employed for injecting the sample onto the GC column depended on the volatility or "stickiness" of the compound. For analysis of the more volatile species, which included the organic compounds in the ROG surrogates used in this study, the contents of the syringe were flushed through a 10 ml and 5 ml stainless steel or $1 / 8$ ' Teflon tube loop and subsequently injected onto the column by turning a gas sample valve.

Formaldehyde was monitored using an adaptation of the diffusion scrubber method developed by Dasgupta et al (1988, 1990), as described by Carter et al (1995c). It was calibrated using a formaldehyde diffusion tube whose weight loss was monitored over time. The system cycled between zero, calibrate, and sample modes to correct for zero and span drifts.

For most experiments with DMSO a Meloy 285-SA photometric sulfur analyzer was used to verify the initial DMSO injections and analyze for total gas-phase sulfur during the experiments.. The Meloy instrument was calibrated prior the chamber experiments with the certified $\mathrm{SO}_{2}$ source. This instrument is a total gas-phase sulfur analyzer, and thus responds to DMSO and probably other gas-phase sulfur-containing species as well as $\mathrm{SO}_{2}$. The response of this instrument to the expected DMSO oxidation product dimethyl sulfone ( $\mathrm{DMSO}_{2}, \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{CH}_{3}$ ) is not known. Note that this instrument does not respond to particle phase sulfur (e.g., sulfate aerosol) because a particle filter is used in the sampling inlet.

The calibrations for the GC analyses for most compounds were carried out by sampling from chambers or vessels of known volume into which known amounts of the reactants were injected, as described previously (Carter et al, 1995c).

## Characterization Methods

## Temperature

Three temperature thermocouples were used to monitor the chamber temperature, two of which were located in the sampling line of continuous analyzers to monitor the temperature in each side. The third one was located in the outlet of the air conditioning system used to control the chamber temperature. The temperature range in these experiments was typically 25-30 C.

## Blacklight Light Source

The light intensity in the DTC chamber was monitored by periodic $\mathrm{NO}_{2}$ actinometry experiments utilizing the quartz tube method of Zafonte et al (1977), with the data analysis method modified as discussed by Carter et al. (1995c). The results of these experiments were tracked over time, and although there was a gradual decrease in light intensity over time during most of the operational lifetime of this chamber, the light intensity appeared to be relatively constant during the period of these experiments. Averages of results of actinometry experiments carried out during this period indicated an $\mathrm{NO}_{2}$ photolysis rate of $0.161 \mathrm{~min}^{-1}$. This was used when modeling all the experiments for this program.

The spectrum of the blacklight light source is periodically measured using a LiCor LI-1200 spectra radiometer, and found to be essentially the same as the general blacklight spectrum recommended by Carter et al (1995c) for use in modeling blacklight chamber experiments.

## Dilution

The dilution of the DTC chamber due to sampling is expected to be small because the flexible reaction bags can collapse as samples are withdrawn for analysis. Also, the chamber was designed to operate under slightly positive pressure, so any small leaks would result in reducing the bag volume rather than diluting the contents of the chamber. Information concerning dilution in an experiment can be obtained from relative rates of decay of added VOCs which react with OH radicals with differing rate constants (Carter et al. 1993; 1995c). Most experiments had a more reactive compounds such as m-xylene and n -octane present either as a reactant or added in trace amounts to monitor OH radical levels. Trace amounts ( $\sim 0.1 \mathrm{ppm}$ ) of n -butane were also added to experiments if needed to provide a less reactive compound for monitoring dilution. In addition, specific dilution check experiments such as CO irradiations were carried out. Based on these results, the dilution rate was found to be negligible in this chamber during this period, being less than $0.3 \%$ per hour in all runs, and usually less than $0.1 \%$ per hour.

## Reactivity Data Analysis Methods

As indicated above, most of the experiments for this program consisted of simultaneous irradiation of a "base case" reactive organic gas (ROG) surrogate - $\mathrm{NO}_{\mathrm{x}}$ mixture in one of the dual reaction chambers, together with an irradiation, in the other reactor, of the same mixture with the test compound (DMSO in this case) added. The results are analyzed to yield two measures of VOC reactivity: the effect of the added VOC on the amount of NO reacted plus the amount of ozone formed, and integrated OH radical levels. These are discussed in more detail below.

The first measure of reactivity is the effect of the VOC on the change in the quantity $\left[\mathrm{O}_{3}\right]-[\mathrm{NO}]$, or $\left(\left[\mathrm{O}_{3}\right]_{t}-[\mathrm{NO}]_{t}\right)-\left(\left[\mathrm{O}_{3}\right]_{0}-[\mathrm{NO}]_{0}\right)$, which is referred to as $\Delta([\mathrm{O} 3]-[\mathrm{NO}])$ in the subsequent discussion. As discussed elsewhere (e.g., Johnson, 1983; Carter and Atkinson, 1987; Carter and Lurmann, 1990, 1991, Carter et al, 1993, 1995a), this gives a direct measure of the amount of conversion of NO to $\mathrm{NO}_{2}$ by peroxy radicals formed in the photooxidation reactions, which is the process that is directly responsible for ozone formation in the atmosphere. (Johnson calls it "smog produced" or "SP".) The incremental reactivity of the VOC relative to this quantity, which is calculated for each hour of the experiment, is given by

$$
\begin{equation*}
\operatorname{IR}\left[\Delta\left(\left[\mathrm{O}_{3}\right]-[\mathrm{NO})_{\mathrm{t}}^{\mathrm{VOC}}\right]=\frac{\Delta\left(\left[\mathrm{O}_{3}\right]-[\mathrm{NO})_{\mathrm{t}}^{\mathrm{Test}}-\Delta\left(\left[\mathrm{O}_{3}\right]-[\mathrm{NO})_{\mathrm{t}}^{\text {Base }}\right.\right.}{[\mathrm{VOC}]_{0}}\right. \tag{I}
\end{equation*}
$$

where $\Delta\left([\mathrm{O} 3]-[\mathrm{NO})_{\mathrm{t}}{ }^{\text {Test }}\right.$ is the $\Delta\left(\left[\mathrm{O}_{3}\right]-[\mathrm{NO}]\right)$ measured at time t from the experiment where the test VOC was added, $\Delta\left([\mathrm{O} 3]-[\mathrm{NO})_{\mathrm{t}}{ }^{\text {Base }}\right.$ is the corresponding value from the corresponding base case run, and $[\mathrm{VOC}]_{0}$ is the amount of test VOC added. An estimated uncertainty for $\operatorname{IR}\left[\Delta\left(\left[\mathrm{O}_{3}\right]-[\mathrm{NO}]\right)\right]$ is derived based on assuming an $\sim 3 \%$ uncertainty or imprecision in the measured $\Delta\left(\left[\mathrm{O}_{3}\right]-[\mathrm{NO}]\right)$ values. This is consistent with the results of the side equivalency test, where equivalent base case mixtures are irradiated on each side of the chamber.

Note that reactivity relative to $\Delta([\mathrm{O} 3]-[\mathrm{NO}])$ is essentially the same as reactivity relative to $\mathrm{O}_{3}$ in experiments where $\mathrm{O}_{3}$ levels are high, because under such conditions $[\mathrm{NO}]_{\mathrm{t}}^{\text {base }} \approx[\mathrm{NO}]_{\mathrm{t}}^{\text {test }} \approx 0$, so a change in $\Delta([\mathrm{O} 3]-[\mathrm{NO}])$ caused by the test compound is due to the change in $\mathrm{O}_{3}$ alone. However, $\Delta([\mathrm{O} 3]-[\mathrm{NO}])$ reactivity has the advantage that it provides a useful measure of the effect of the VOC on processes responsible for $\mathrm{O}_{3}$ formation even in experiments where $\mathrm{O}_{3}$ formation is suppressed by relatively high NO levels.

The second measure of reactivity is the effect of the VOC on integrated hydroxyl $(\mathrm{OH})$ radical concentrations in the experiment, which is abbreviated as "IntOH" in the subsequent discussion. This is an important factor affecting reactivity because radical levels affect how rapidly all VOCs present, including the base ROG components, react to form ozone. If a compound is present in the experiment that reacts primarily with OH radicals, then the IntOH at time $t$ can be estimated from

$$
\begin{equation*}
\mathrm{IntOH}_{\mathrm{t}}=\frac{\ln \left([\text { tracer }]_{0} /[\text { tracer }]_{\mathrm{t}}\right)-\mathrm{Dt}}{\mathrm{kOH}^{\text {tracer }}} \tag{II}
\end{equation*}
$$

where $[\operatorname{tracer}]_{0}$ and $[\text { tracer }]_{\mathrm{t}}$ are the initial and time=t concentrations of the tracer compound, $\mathrm{kOH}^{\text {tracer }}$ its OH rate constant, and D is the dilution rate in the experiments. The latter was found to be small and was neglected in our analysis. The concentration of tracer at each hourly interval was determined by linear interpolation of the experimentally measured values. M-xylene was used as the OH tracer in these experiments because it is a surrogate component present in all experiments, its OH rate constant is known (the value used was $2.36 \times 10^{-11} \mathrm{~cm}^{3}$ molec $^{-1} \mathrm{~s}^{-1}$ [Atkinson, 1989]), and it reacts relatively rapidly.

The effect of the VOC on OH radicals can thus be measured by its IntOH incremental reactivity, which is defined as

$$
\begin{equation*}
\mathrm{IR}[\mathrm{IntOH}]_{\mathrm{t}}=\frac{\mathrm{IntOH}_{\mathrm{t}}^{\text {Test }}-\mathrm{IntOH}_{\mathrm{t}}^{\text {Base }}}{[\mathrm{VOC}]_{0}} \tag{III}
\end{equation*}
$$

where $\operatorname{IntOH}$ est and $\mathrm{IntOH}^{\text {Base }}$ are the IntOH values measured at time t in the added VOC and the base case experiment, respectively. The results are reported in units of $10^{6} \mathrm{~min}$. The uncertainties in IntOH and $\operatorname{IR}[\mathrm{IntOH}]$ are estimated based on assuming an $\sim 2 \%$ imprecision in the measurements of the m -xylene concentrations. This is consistent with the observed precision of results of replicate analyses of this compound.

## CHEMICAL MECHANISMS

## General Atmospheric Photooxidation Mechanism

The chemical mechanism used in the environmental chamber and atmospheric model simulations in this study is the "SAPRC-99" mechanism, which is documented in detail by Carter (2000). This mechanism represents a complete update of the SAPRC-90 mechanism of Carter (1990), and incorporates recent reactivity data from a wide variety of VOCs, though not (up to this work) including DMSO. This includes assignments for $\sim 400$ types of VOCs, and can be used to estimate reactivities for $\sim 550$ VOC categories. A condensed version, developed for use in regional models, is used to represent base case emissions in the atmospheric reactivity simulations discussed in this report. The mechanism was evaluated against the results of almost 1700 environmental chamber experiments carried out at the University of California at Riverside, including experiments to test ozone reactivity predictions for over 80 types of VOCs.

A listing of the portions of the mechanism that was used in the model simulations discussed in this report is given in Appendix A. These consist of the "base mechanism" representing the reactions of the inorganics and common organic products, the reactions of the specific VOCs used in the environmental chamber experiments, and the reactions of the lumped model species used when representing base case VOCs in the ambient reactivity simulations. The listing in Appendix A does not include the reactions of DMSO, which are not part of the SAPRC-99 mechanism as documented by Carter (2000), and which had to be added for the purpose of this study. The reactions of DMSO and how they were represented in the model calculations discussed in this work are discussed in the following section.

## Atmospheric Reactions of Dimethyl Sulfoxide

The possible gas-phase consumption reactions that need to be considered when assessing atmospheric impacts of VOCs are the reactions of the compound with OH radicals, with $\mathrm{O}_{3}$, with $\mathrm{NO}_{3}$ radicals, and by direct photolysis. DMSO does not have a measurable absorption cross section at wavelengths less than 250 nm (Hynes and Wine, 1995), so it should not undergo significant direct photolysis in the atmosphere. Information concerning the other reaction pathways, and the mechanisms used in the model simulations in this work, are discussed below.

## Reaction with OH Radicals

The room temperature rate constant for the reaction of OH radicals with DMSO has been measured in several laboratories, and the available data are summarized in Table 1. There some differences between the measurements, with the data of Barnes et al (1989) and Falbe-Hansen et al (2000) indicating a rate constant of around $6 \times 10^{-11} \mathrm{~cm}^{3} \mathrm{molec}^{-1} \mathrm{~s}^{-1}$, while data of Hynes and Wine (1996) and

Table 1. Summary of measured room temperature rate constants for the reactions of DMSO with OH and $\mathrm{NO}_{3}$ radicals and $\mathrm{O}_{3}$. Error ranges indicate $2 \sigma$ overall uncertainties.

| Reaction | Rate Constant $\left(\mathrm{cm}^{3}\right.$ molec $\left.^{-1} \mathrm{~s}^{-1}\right)$ | Reference |
| :--- | :---: | :--- |
| OH | $(5.9 \pm 1.5) \times 10^{-11}$ | Falbe-Hansen et al (2000) |
|  | $(6.2 \pm 2.2) \times 10^{-11}$ | Barnes et al (1989) |
|  | $(10 \pm 3) \times 10^{-11}$ | Hynes and Wine (1996) |
|  | $(8.7 \pm 1.6) \times 10^{-11}$ | Urbanski et al (1998) |
| $\mathrm{NO}_{3}$ | $(5.0 \pm 3.8) \times 10^{-13}$ | Falbe-Hansen et al (2000) |
|  | $(1.7 \pm 0.3) \times 10^{-13}$ | Barnes et al (1989) |
| $\mathrm{O}_{3}$ | $<3 \times 10^{-20}$ | This work |
|  | $<1 \times 10^{-19}$ | Falbe-Hansen et al (2000) |
|  | $<5 \times 10^{-19}$ | Barnes et al (1989) |

Urbanski et al (1998) indicate rate constants around $9 \times 10^{-11} \mathrm{~cm}^{3} \mathrm{molec}^{-1} \mathrm{~s}^{-1}$. However, these differences are within the stated uncertainties of most of the measurements, and possibility reflects difficulties in handling this relatively low volatility compound. For this work, we use

$$
\mathrm{k}_{\mathrm{OH}+\mathrm{DMDO}}=7.5 \times 10^{-11} \mathrm{~cm}^{3} \mathrm{molec}^{-1} \mathrm{~s}^{-1}
$$

which is approximately the average of these determinations. The uncertainty is approximately $\pm 30 \%$. This is a relatively high rate constant, indicating an atmospheric lifetime of less than one day (Falbe-Hansen et al, 2000).

There is considerable uncertainty concerning the details of the mechanism of the reaction of DMSO with OH radicals, and there are inconsistencies in the data in the literature. The available product data for the reactions of DMSO with OH radicals are summarized on Table 2. It can be seen that there is considerable variability with reaction conditions and in some cases differences between experiments carried out under comparable conditions. Possible mechanisms accounting for these products are discussed below.

OH radicals can react with DMSO either by adding to the sulfur forming a vibrationally excited adduct (reaction 1 ), or by abstraction from the methyl group (reaction 2 ):

$$
\begin{gather*}
\mathrm{OH}+\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})(\mathrm{OH}) \mathrm{CH}_{3}{ }^{*}  \tag{1}\\
\mathrm{OH}+\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \cdot+\mathrm{H}_{2} \mathrm{O} \tag{2}
\end{gather*}
$$

Under low pressure conditions, the major fate of the adduct might be either decomposition back to $\mathrm{OH}+$ DMSO or formation of methyl radicals and methane sulfinic acid (MSIA).

$$
\begin{equation*}
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})(\mathrm{OH}) \mathrm{CH}_{3}{ }^{*} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OH}+\mathrm{CH}_{3} . \tag{3}
\end{equation*}
$$

Table 2. Summary of the available product yield data concerning the reactions of OH radicals with DMSO. All experiments were carried out at approximately ambient temperature ( $\sim 298 \mathrm{~K}$ ).

| Reference | Barnes et al (1989) | Sørensen et al (1996) | Becker and Patroescu (1996) |  |  |  | $\begin{aligned} & \text { Urbanski } \\ & \text { et al } \\ & (1999) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |
| Pressure | Atm. | Atm. | Atm. | Atm. | Atm | Atm. | Low |
| Reactants other than DMSO | $\mathrm{NO}_{2}$, Air | $\underset{\text { Air }}{\mathrm{CH}_{3} \mathrm{ONO}}$ | $\begin{aligned} & \mathrm{CH}_{3} \mathrm{ONO}, \\ & \mathrm{NO}, \text { Air } \end{aligned}$ | $\begin{gathered} \mathrm{H}_{2} \mathrm{O}_{2} \\ \mathrm{NO}_{2}, \mathrm{Air} \end{gathered}$ | $\mathrm{H}_{2} \mathrm{O}_{2}$, Air | $\mathrm{H}_{2} \mathrm{O}_{2}, \mathrm{~N}_{2}$ | $\begin{gathered} \mathrm{H}_{2} \mathrm{O}_{2}, \mathrm{~N}_{2}, \\ \mathrm{CH}_{4} \end{gathered}$ |
| Photolysis $\lambda$ (nm) | $\geq 300$ | $\geq 300$ | $\geq 300$ | 254 max | 254 max | 254 max | 248 |
| $\underset{\left(\mathrm{DMSO}_{2}\right)}{\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{CH}_{3}}$ | ~30 | $22 \pm 10$ | $5 \pm 1$ | $30 \pm 8$ | $29 \pm 9$ | $\sim 5$ | - |
| $\underset{(\text { MSA })}{\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OH}}$ | - | $<0.5$ | $1.2 \pm 0.5$ | $19 \pm 6$ | $6 \pm 2$ | $\sim 0$ | - |
| $\underset{\text { (MSIA) }}{\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OH}}$ | - | $<0.3$ | - | - | - | - | high [d] |
| $\underset{\text { (MSPN) }}{\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OONO}_{2}}$ | Obs. [a,b] | $10 \pm 5$ | $58 \pm 10$ | Obs. [c] | - | - | - |
| $\mathrm{SO}_{2}$ | $60 \pm 10$ | $20 \pm 15$ | $15 \pm 4$ | $46 \pm 4$ | $40 \pm 4$ | $\sim 60$ | - |
| $\mathrm{SO}_{4}{ }^{-}$ | - | <0.1 | - | - | - | - | - |
| HCHO | Obs. | - | - | $52 \pm 3$ | $30 \pm 7$ | $\sim 26$ | - |
| $\mathrm{CH}_{3} \mathrm{OH}$ | - | - | - | $17 \pm 6$ | $12 \pm 2$ | $\sim 8$ | - |
| $\mathrm{CH}_{3} \mathrm{OOH}$ | - | - | - | $27 \pm 18$ | $32 \pm 6$ | $\sim 35$ | - |
| $\mathrm{HC}(\mathrm{O}) \mathrm{OH}$ | - | - | - | $7 \pm 3$ | $10 \pm 5$ | $\sim 12$ | - |
| CO | Obs. | - | - | $40 \pm 14$ | $34 \pm 3$ | $\sim 72$ | - |
| $\mathrm{CH}_{3} \mathrm{ONO}_{2}$ | Obs. | - | - | - | - | - | - |
| $\mathrm{CH}_{3}$. | - | - | - | - | - | - | $98 \pm 12$ |
| Sulfur balance | ~90 | $53 \pm 30$ | $79 \pm 16$ | $94 \pm 18$ | $76 \pm 15$ | $\sim 65$ | - |
| Carbon balance | - | - | - | $111 \pm 33$ | $91 \pm 22$ | $\sim 80$ | - |

[a] Obs. = Observed but not quantified.
[b] Structure given as $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OONO}_{2}$ by Barnes et al (1989), but based on discussion in Becker and Patroescu (1996) it is believed that this is the same product that is identified in subsequent work in this laboratory as $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OONO}_{2}$
[c] Up to 5\% Sulfur at 30\% DMSO consumption; subsequently decayed.
[d] Not observed directly, but this product is predicted to be formed in high yield as the co-product from $\mathrm{CH}_{3}$. See text.

The data of Urbanski et al (1999) suggest that this may be the major process, since high yields of methyl radicals are observed in experiments carried out under low pressures. Further evidence that addition dominates over abstraction comes from the Hynes and Wine (1996), who observed no apparent kinetic isotope effect in the reaction of OH with $\mathrm{CD}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CD}_{3}$. A measurable kinetic isotope effect would be expected if abstraction (Reaction 2) were important.

Under higher pressure conditions in the presence of $\mathrm{O}_{2}$, the adduct would be expected to be stabilized and react with $\mathrm{O}_{2}$ to form $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{CH}_{3}\left(\mathrm{DMSO}_{2}\right)$.

$$
\begin{gather*}
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})(\mathrm{OH}) \mathrm{CH}_{3} *+\mathrm{M} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})(\mathrm{OH}) \mathrm{CH}_{3}+\mathrm{M}  \tag{4}\\
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})(\mathrm{OH}) \mathrm{CH}_{3}+\mathrm{O}_{2} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{CH}_{3}+\mathrm{HO}_{2} . \tag{5}
\end{gather*}
$$

The observation of $\mathrm{DMSO}_{2}$ in $\sim 20-30 \%$ yields in most experiments carried out in 1 atm of air (Table 2) suggests that this reaction is important but not dominant under atmospheric conditions. (The reason for the low $\mathrm{DMSO}_{2}$ yield in the $\mathrm{CH}_{3} \mathrm{ONO} / \mathrm{NO}$ experiment of Becker and Patroescu (1996) is unknown, but based on the consistent data from the other studies in different laboratories it is assumed to be anomalous.) The facts that the $\mathrm{DMSO}_{2}$ yield decreases with reduced $\mathrm{O}_{2}$ (Becker and Patroescu, 1996) but appears to be independent of $\mathrm{NO}_{\mathrm{x}}$ are consistent with this mechanism.

The facts that the $\mathrm{DMSO}_{2}$ yields under atmospheric conditions are no greater than $\sim 30 \%$ and that $\mathrm{SO}_{2}$ and other products are observed indicate that either decomposition of the excited adduct (Reaction 3) is still important at atmospheric pressure or that, contrary to the data of Urbanski et al (1999) and Hynes and Wine (1996), abstraction (Reaction 2) is occurring to a significant extent. If Reaction (3) is assumed to be the major competing process, then the expected products would be high yields of MSIA and the various products formed from $\mathrm{CH}_{3}$, which would include primarily HCHO in the presence of $\mathrm{NO}_{\mathrm{x}}$, and $\mathrm{HCHO}, \mathrm{CH}_{3} \mathrm{OH}$, and $\mathrm{CH}_{3} \mathrm{OOH}$ in the absence of $\mathrm{NO}_{\mathrm{x}}$. MSIA is not observed in the high yields predicted by this mechanism, but it is expected to have relatively weak O-H bonds (Yin et al, 1990) and thus is likely to react relatively rapidly with OH radicals via:

$$
\begin{equation*}
\mathrm{OH}+\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OH} \rightarrow \mathrm{H}_{2} \mathrm{O}+\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \tag{6}
\end{equation*}
$$

The $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O}$. could either decompose, which would account for the observed formation of $\mathrm{SO}_{2}$, or react with $\mathrm{O}_{2}$, which could account for the eventual formation of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OONO}_{2}$ (MSPN) and $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OH}$ (MSA).

$$
\begin{gather*}
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \cdot \mathrm{CH}_{3} \cdot+\mathrm{SO}_{2}  \tag{7}\\
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \cdot+\mathrm{O}_{2}+\mathrm{M} \Leftrightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OO} \cdot+\mathrm{M}  \tag{-8}\\
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OO} \cdot+\mathrm{NO}_{2} \Leftrightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OONO}_{2}(\mathrm{MSPN})  \tag{-9}\\
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OO} \cdot+\mathrm{NO} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{O} \cdot+\mathrm{NO}_{2}  \tag{9}\\
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OO} \cdot+\mathrm{RO}_{2} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{O} \cdot+\mathrm{RO} \cdot+\mathrm{O}_{2}
\end{gather*}
$$

$$
\begin{equation*}
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{O} \cdot+\mathrm{R}-\mathrm{H} \text { or walls } \rightarrow \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OH}(\mathrm{MSA})(+\mathrm{R} \cdot ?) \tag{12}
\end{equation*}
$$

The variable yields of $\mathrm{SO}_{2}$, in the various studies may be due to the possibility that the reaction of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O}$. with $\mathrm{O}_{2}$ may be reversible, making the competition between Reaction (7) and formation of MSPN or MSA being dependent on reaction conditions. In addition, the variable yields of MSPN and other products could also be due to the thermal instability of MSPN and the fact that its formation and decomposition may also depend on reaction conditions. The rate of decomposition of MSPN is uncertain; if it is as stable as acyl peroxynitrates (e.g, $\mathrm{PAN}, \mathrm{CH}_{3} \mathrm{C}(\mathrm{O}) \mathrm{OONO}_{2}$ ) it may be relatively stable under atmospheric conditions, but if it decomposes as rapidly as methyl peroxynitrate $\left(\mathrm{CH}_{3} \mathrm{OONO}_{2}\right)$, its formation would not be significant at ambient temperatures.

The rate constant for Reaction (7) has been measured to be about $510 \pm 150 \mathrm{~s}^{-1}$ at 298 K , which means that to be competitive under atmospheric conditions the net effective rate constant for reaction with $\mathrm{O}_{2}$ (Reaction 8) would have to be less than $\sim 1 \times 10^{-16} \mathrm{~cm}^{3}$ molec $^{-1} \mathrm{~s}^{-1}$. No information could be found concerning the kinetics of this reaction with $\mathrm{O}_{2}$, or its reverse.

Although this appears to be a reasonable explanation of the available literature data, as discussed later in this report models based on this mechanism give predictions that are inconsistent with the results of the environmental chamber experiments carried out for this program. In addition, it is difficult to reconcile this explanation with the fact that, as indicated in Table 2, only low yields of MSIA are reported in the DMSO +OH product studies carried out under approximate atmospheric conditions. Therefore, alternative explanations of these data need to be considered.

An alternative explanation to account for the observation of the other products besides $\mathrm{DMSO}_{2}$ would be to assume that, contrary to the conclusions drawn based on the data of Urbanski et al (1999) and Hynes and Wine (1999), the abstraction reaction of OH with DMSO (Reaction 2) is indeed significant. The subsequent reactions of the $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \cdot$ radical formed are expected to be as follows, where (for simplicity) only the reactions expected to be important in the presence of $\mathrm{NO}_{\mathrm{x}}$ are shown:

$$
\begin{gather*}
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \cdot+\mathrm{O}_{2} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \mathrm{OO} .  \tag{13}\\
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \mathrm{OO}+\mathrm{NO} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \mathrm{O} \cdot+\mathrm{NO}_{2}  \tag{14}\\
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \mathrm{O} \cdot \rightarrow \mathrm{HCHO}+\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) . \tag{15}
\end{gather*}
$$

In the presence of $\mathrm{NO}_{\mathrm{x}}$ the $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})$. radicals would be expected to be converted to $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O}$-, which could then react via Reactions (7-12), above, forming the same products as would result from the MSIA + OH mechanism. This could occur either by reaction with $\mathrm{O}_{2}$ followed by reaction converting NO to $\mathrm{NO}_{2}$,

$$
\begin{align*}
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \cdot+\mathrm{O}_{2} & \Leftrightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OO} .  \tag{-16}\\
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OO}+\mathrm{NO} & \rightarrow \mathrm{NO}_{2}+\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} . \tag{16}
\end{align*}
$$

or by reaction with $\mathrm{NO}_{2}$ or $\mathrm{O}_{3}$.

$$
\begin{align*}
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \cdot+\mathrm{NO}_{2} & \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O}+\mathrm{NO}  \tag{18}\\
\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \cdot+\mathrm{O}_{3} & \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \cdot+\mathrm{O}_{2} \tag{19}
\end{align*}
$$

The reactions of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \cdot$ with $\mathrm{NO}_{2}$ and $\mathrm{O}_{3}$ have been studied, and their rate constants recommended by the IUPAC evaluation (Atkinson et al, 1997) are $1.2 \times 10^{-11}$ and $6 \times 10^{-13} \mathrm{~cm}^{3} \mathrm{molec}^{-1} \mathrm{~s}^{-1}$, respectively. However, the rate and equilibrium constants for the reactions of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \cdot$ with $\mathrm{O}_{2}$ is unknown, so the relative importance of Reactions (16) and (17) compared to Reactions (18) or (19) are unknown.

There are clearly significant uncertainties concerning the relative importances of many of these competing processes involved in the reactions of OH radicals with DMSO, so a number of alternative mechanisms were examined to determine which sets of assumptions are most consistent with the environmental chamber data obtained in this program. Although a large number of possibilities can be considered, six representative alternative mechanisms, designated Mechanisms A - F, were considered, based making several alternative assumptions concerning the various uncertain processes discussed above. These are summarized on Table 3. As indicated there, differing assumptions were made concerning the relative importance of addition vs. abstraction (Reaction 1 vs. 2) in the initial reaction of OH with DMSO, and the reactions of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \cdot \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \cdot$, and $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{O} \cdot$, and $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) 2 \mathrm{OONO}_{2}$ where applicable. On the other hand, all five mechanisms are based on the assumption that the overall yield of $\mathrm{DMSO}_{2}$ under atmospheric conditions is $25 \%$, and that relatively high yields of $\mathrm{SO}_{2}$ are formed under atmospheric conditions. These assumptions appear to be indicated by the available laboratory data, as shown in Table 2.

Clearly, other alternative assumptions concerning the uncertain reactions can be made, and the rate constants and branching rations used in some of the alternative mechanisms are somewhat arbitrary. However, as discussed in the Results section, the examination of these alternatives turned out to be sufficient to indicate the types of mechanisms that are or are not consistent with the chamber data obtained in this work. For example, the results showed that mechanisms assuming nonnegligible radical inhibition processes (e.g., Mechanisms "D" and "E") performed very poorly in simulating our data, so other alternative mechanisms involving possible radical inhibition processes need not be considered.

## Reaction with $\mathrm{NO}_{3}$ Radicals

The room temperature rate constant for the reaction of $\mathrm{NO}_{3}$ radicals with DMSO have been measured by Barnes et al (1989) and Falbe-Hansen et al (2000), and the results are summarized in Table 1. These values differ by about a factor of 3, but Falbe-Hansen et al (2000) indicate that they consider this agreement to be within the uncertainty of the measurement because of the difficulties in handling this compound. For modeling purposes, we use the geometric mean of these two determinations, which is

$$
\mathrm{K}_{\mathrm{NO} 3+\mathrm{DMDO}}=3 \times 10^{-13} \mathrm{~cm}^{3} \mathrm{molec}^{-1} \mathrm{~s}^{-1}
$$

Under atmospheric conditions, reaction with $\mathrm{NO}_{3}$ is expected to be a relatively minor fate for DMSO compared to reaction with OH radicals (Falbe-Hansen et al, 2000), so the uncertainty in this rate constant

Table 3. Alternative mechanisms for the reactions of OH with DMSO that were considered in the model simulations of the environmental chamber experiments for this project.

| Mechanistic Assumptions for Atmospheric Conditions | Mechanism [a] |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | A | B | C | D | E | F |
| Initial OH + DMSO Reaction: |  |  |  |  |  |  |
| Reacts only by addition, as suggested by the data of Urbanski et al (1999) and Hynes and Wine (1999). (Reaction 1 dominates over Reaction 2). | X | - | - | - | - | - |
| Reacts $25 \%$ of the time by addition (Reaction 1) and $75 \%$ of the time by abstraction (Reaction 2). | - | X | X | X | X | X |
| $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})(\mathrm{OH}) \mathrm{CH}_{3}$ Reactions |  |  |  |  |  |  |
| Stabilized and reacts with $\mathrm{O}_{2}$ to form $\mathrm{DMSO}_{2} 25 \%$ of the time, and decomposes to $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OH}$ and $\mathrm{CH}_{3} .75 \%$ of the time. | X | - | - | - | - | - |
| Primarily stabilized and reacts with $\mathrm{O}_{2}$ to form $\mathrm{DMSO}_{2}$. | - | X | X | X | X | X |
| $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OH}$ Reactions |  |  |  |  |  |  |
| Reacts with OH radicals (via Reaction 6) with a very high rate constant of $1.0 \times 10^{-10} \mathrm{~cm}^{3}$ molec $^{-1} \mathrm{~s}^{-1}$. | X | N/A | N/A | N/A | N/A | N/A |
| $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \cdot$ Reactions |  |  |  |  |  |  |
| Either reacts only slowly with $\mathrm{O}_{2}$ (or the equilibrium constant is such that the decomposition of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OO}$ - is fast). The major fate for $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})$. is therefore reaction with $\mathrm{NO}_{2}, \mathrm{O}_{3}$, or peroxy radicals. | N/A | X | - | - | - | - |
| The rate and equilibrium constant for the reaction with $\mathrm{O}_{2}$ are sufficiently high that reaction with $\mathrm{O}_{2}$ (Reaction 16) will be the major net fate, and decomposition of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OO} \cdot$ is negligible compared to competing reactions. | N/A | - | X | X | X | X |
| $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \cdot$ Reactions |  |  |  |  |  |  |
| Assumed to primarily decompose to $\mathrm{CH}_{3} \cdot$ and $\mathrm{SO}_{2}$. | X | X | X | - | - | - |
| Approximately half decomposes to $\mathrm{CH}_{3} \cdot$ and $\mathrm{SO}_{2}$, with the other half reacting with $\mathrm{O}_{2}$ to form $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OO} \cdot$ (Reaction 8), which reacts with NO and $\mathrm{NO}_{2}$ and peroxy radicals (Reactions 9-11) with rate constants that are the same as those for analogous reactions of acyl peroxy radicals (e.g., $\mathrm{RC}(\mathrm{O}) \mathrm{OO} \cdot)$. | - | - | - | X | X | X |
| $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OONO}_{2}$ Reactions |  |  |  |  |  |  |
| Decomposes with the same A factor and activation energy as higher PAN $\left(\mathrm{RC}(\mathrm{O}) \mathrm{OONO}_{2}\right.$ ) analogues (relatively stable). | N/A | N/A | N/A | - | X | - |
| Decomposes with a similar activation energy as methyl peroxynitrate (relatively unstable). | N/A | N/A | N/A | X | - | X |
| $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{O} \cdot$ Reactions |  |  |  |  |  |  |
| Reacts to form $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OH}$ (MSA) via a processes that does not regenerate radicals. | N/A | N/A | N/A | X | - | - |
| Reacts to form $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OH}$ (MSA) via a processes that regenerates radicals. This is represented by $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{O} \cdot \rightarrow \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OH}+\mathrm{OH}$. | N/A | N/A | N/A | - | X | X |

[a] $\mathrm{X}=$ Assumed; $-=$ Not assumed; N/A $=$ Irrelevant given the other assumptions used in this mechanism.
is probably not a major uncertainty in terms of atmospheric simulations. However, the reaction with $\mathrm{NO}_{3}$ was found to be a non-negligible process in the environmental chamber experiments carried out for this study, which employs a light source which have relatively low intensities in the visible parts of the spectrum that most affect the photolysis rates of $\mathrm{NO}_{3}$. Therefore, this uncertainty may have some effect on the predictions of the mechanism developed in this study. This is discussed further in the Results section.

The products of the reaction of DMSO with $\mathrm{NO}_{3}$ have been studied by Barnes et al (1989) and Falbe-Hansen et al (2000), and the only product they observed (other than $\mathrm{HNO}_{3}$ ) was $\mathrm{DMSO}_{2} . \mathrm{SO}_{2}$ in particular was not formed. This suggests that the $\mathrm{DMSO}+\mathrm{NO}_{3}$ mechanism proceeds via

$$
\begin{align*}
& \mathrm{CH}_{2} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{3}+\mathrm{NO}_{3} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})\left(\mathrm{ONO}_{2}\right) \mathrm{CH}_{3}  \tag{19}\\
& \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})\left(\mathrm{ONO}_{2}\right) \mathrm{CH}_{3} \rightarrow \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{CH}_{3}+\mathrm{NO}_{2} \tag{20}
\end{align*}
$$

and this is what is assumed in this work. However, Falbe-Hansen et al (2000) observed variable yields of $\mathrm{DMSO}_{2}$ (from $10-94 \%$ ), suggesting that a more complex mechanism may be occurring. The effects of assuming that other processes may be occurring are discussed further in the Results section.

## Reaction with $\mathrm{O}_{3}$

As indicated in Table 2, attempts to measure the rate constant for the reaction of DMSO have been made by Barnes et al (1989), Falbe-Hansen et al (2000), and in this work, and only upper limit rate constants have been obtained. The lowest upper limit is that obtained in this work (see below), where the rate constant was found to be less than $10^{-20} \mathrm{~cm}^{3} \mathrm{molec}^{-1} \mathrm{~s}^{-1}$. Therefore, we assume that this reaction is negligible in the model simulations in this study.

## Representation of DMSO in the Model Simulations

The atmospheric reactions of DMSO used in the model simulations in this study were derived based on the considerations discussed in the previous sections, with each of the six alternative mechanisms indicated in Table 3 being used in the simulations of the chamber experiments. The listing of these mechanisms in terms of SAPRC-99 model species is given in Table 4. Footnotes to the table document the reactions and rate constants used, where appropriate, and indicate the terminology employed. The listings for the rest of the base SAPRC-99 mechanism and the mechanisms for the other VOC species used in the model simulations are given in Appendix A of this report.

Table 4. Reactions and rate constants used to represent the alternate DMSO mechanism in the SAPRC-99 model calculations.


## Footnotes for Table 4:

[a] Except as indicated, the rate constants are given by $\mathrm{k}(\mathrm{T})=\mathrm{A} \cdot \mathrm{e}^{-\mathrm{Ea} / \mathrm{RT}}$, where the units of k and A are $\mathrm{cm}^{3}$ molec $^{-1} \mathrm{~s}^{-1}$, Ea are kcal $\mathrm{mol}^{-1}$, T is ${ }^{0} \mathrm{~K}$, and $\mathrm{R}=0.0019872 \mathrm{kcal} \mathrm{mol}^{-1} \mathrm{deg}^{-1}$. Exceptions are: Fast: The reaction is assumed to be the only fate of the reactant, for which the steady state approximation is used. Same k as Rxn label: The rate constant is the same as that for the reaction on Table A-2 in Appendix A with the indicated label.
[b] Footnotes documenting the reactions are as follows. See text for additional discussion.

1. Rate constant is near middle of range of experimental values shown on Table 1. Temperature dependence is assumed to be small and is ignored.
2. See text and Table 3 for a discussion of the alternative assumptions concerning these reactions.
3. Overall reactions of the $\mathrm{OH}-\mathrm{DMSO}$ adduct with $\mathrm{O}_{2}$ or (for mechanism A ) or by decomposition are represented as overall net processes in the presence of $\mathrm{O}_{2}$.
4. Rate constant is geometric mean of experimental values shown on Table 1. Temperature dependence is ignored. Mechanism based on assumed $100 \% \mathrm{DMSO}_{2}$ formation as discussed in the text.
5. Incorporates an assumed rapid decomposition of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \cdot$ to $\mathrm{CH}_{3} \cdot$, and reaction of $\mathrm{CH}_{3}$ with $\mathrm{O}_{2}$.
6. $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \cdot$ is assumed to react with $\mathrm{O}_{2}$ to form $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \mathrm{OO}$, then react in the presence of NO to form $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} \mathrm{O}$, which then decomposes to $\mathrm{HCHO}+\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})$. The formation of $\mathrm{HCHO}+$ $\mathrm{CH} 3 \mathrm{SO}+\mathrm{R} 2 \mathrm{O} 2$. (the NO to $\mathrm{NO}_{2}$ conversion operator) represents this net process.
7. Rate constant recommended by (Atkinson et al, 1997). Formation of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O}$. is assumed.
8. Speculative reaction with arbitrarily estimated rate constant to represent fate of $\mathrm{CH}_{3} \mathrm{SO}$ under conditions where both $\mathrm{NO}_{2}$ and $\mathrm{O}_{3}$ are low. Not expected to be important under the conditions of the model simulations carried out using this mechanism.
9. The $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})$ is assumed to react with $\mathrm{O}_{2}$ to form $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OO} \cdot$, which then reacts with NO to form $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \cdot$, which then decomposes to $\mathrm{CH}_{3} \cdot+\mathrm{SO}_{2}$. Therefore, the $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})$. formed in Mechanism " B " is replaced by $\mathrm{R} 2 \mathrm{O} 2+\mathrm{C}-\mathrm{O} 2 .+\mathrm{SO} 2$ to represent this overall process.
10. One half of the $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \cdot$ is assumed to decompose to $\mathrm{CH}_{3}+\mathrm{SO}_{2}$ and therefore is represented as indicated for Mechanism "C", while the other half is assumed to add $\mathrm{O}_{2}$ to form $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OO} \cdot$, which is represented explicitly.
11. Assumed to react with the same rate constant and an analogous mechanism as the lumped higher acyl peroxy radical RCO-O2.
12. Assumed to decompose with the same A factor as the decomposition of the lumped higher acyl peroxynitrate species PAN2, but with the same activation energy as recommended by Atkinson et al (1997) for the decomposition of $\mathrm{CH}_{3} \mathrm{ONO}_{2}$. This predicts that MSIA decomposes sufficiently rapidly that it does not build up in concentration under atmospheric conditions.
13. Assumed to react on the walls to form $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OH}$ without generation of radicals.
14. Assumed to decompose with the same A factor and activation energy as the lumped higher acyl peroxynitrate species PAN2. This is sufficiently slow that build-up of MSIA will be nonnegligible.
15. Assumed to react with other species in the gas phase to generate radicals, e.g., via $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{O} \cdot+\mathrm{R}-\mathrm{H}$ $\rightarrow \mathrm{CH}_{2} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OH}+\mathrm{R} \cdot$. For simplicity, radical generation is represented by $\mathrm{HO}_{2}$, and loss of $\mathrm{R}-\mathrm{H}$ is not represented.
[c] Format of reaction listing: "=" separates reactants from products; "\#number" indicates stoichiometric coefficient, "\#coefficient $\{$ product list \}" means that the stoichiometric coefficient is applied to all the products listed.
[d] A listing of the names and meanings of the model species in the base SAPRC-99 mechanism is given in Table A-1 in Appendix A. The following abbreviations are used for DMSO species: DMSO = Dimethyl sulfoxide $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{3}$; DMSO2 $=$ Dimethyl sulfone $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{CH}_{3}$; MSIA $=$ Methane sulfinic acid $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OH} ; \mathrm{MSPN}=$ Methane sulfonic peroxynitrate $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OONO}_{2} ;$ MSA $=$ Methane sulfonic acid $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OH}$.

## MODELING METHODS

## Environmental Chamber Simulations

The ability of the chemical mechanisms to appropriately simulate the atmospheric impacts of DMSO was evaluated by conducting model simulations of the environmental chamber experiments carried out for this study. This requires including in the model appropriate representations of chamberdependent effects such as wall reactions and characteristics of the light source. The methods used are based on those discussed in detail by Carter and Lurmann (1990, 1991), updated as discussed by Carter et al. (1995c; 1997). The photolysis rates were derived from results of $\mathrm{NO}_{2}$ actinometry experiments and measurements of the relative spectra of the light source. The thermal rate constants were calculated using the temperatures measured during the experiments, with the small variations in temperature with time during the experiment being taken into account. The computer programs and modeling methods employed are discussed in more detail elsewhere (Carter et al, 1995C). The specific values of the chamberdependent parameters used in the model simulations of the experiments for this study are given in Table A-4 in Appendix A.

In the case of DMSO, model simulations were carried out using all six of the alternative mechanisms shown in Table 4.

## Atmospheric Reactivity Simulations

To estimate its effects on ozone formation under conditions more representative of polluted urban atmospheres, incremental reactivities, defined as the change in $\mathrm{O}_{3}$ caused by adding small amounts of a compound to the emissions, were calculated for DMSO, as well as for several other representative compounds. The scenarios employed were those used by Carter (1994a, 2000) to develop various reactivity scales to quantify impacts of VOCs on ozone formation in various environments. These were based on a series of single-day EKMA box model scenarios (EPA, 1984) derived by the EPA to represent 39 different urban ozone exceedence areas around the United States (Baugues, 1990). It was found that $\mathrm{NO}_{\mathrm{x}}$ levels are the most important factor affecting differences in relative ozone impacts among VOCs, and that the ranges of relative reactivities in the various scales can be reasonably well represented by ranges in relative reactivities in three "averaged conditions" scenarios representing three different $\mathrm{NO}_{x}$ conditions. These scenarios were derived by averaging the inputs to the 39 EPA scenarios, except for the $\mathrm{NO}_{\mathrm{x}}$ emissions. In the "maximum reactivity" scenario, the $\mathrm{NO}_{\mathrm{x}}$ inputs were adjusted such that the final $\mathrm{O}_{3}$ level is most sensitive to changes in VOC emissions; in the "maximum ozone" scenario the $\mathrm{NO}_{\mathrm{x}}$ inputs were adjusted to yield the highest maximum $\mathrm{O}_{3}$ concentration; and in the "equal benefit" scenario the $\mathrm{NO}_{\mathrm{x}}$ inputs were adjusted such that relative changes in VOC and $\mathrm{NO}_{x}$ emissions had equal effect on ozone formation. As discussed by Carter (1994a), there represent respectively the high, medium and low ranges
of $\mathrm{NO}_{\mathrm{x}}$ conditions which are of relevance when assessing VOC control strategies for reducing ozone. This is discussed further in the "Atmospheric Reactivity Calculations" section of this report.

The DMSO mechanism used in the atmospheric reactivity simulations was Mechanism "C", which as discussed below was found to gave the best simulations of the environmental chamber data obtained in this work.

## RESULTS AND DISCUSSION

## Upper Limit $\mathrm{O}_{3}$ Rate Constant Measurements

The conditions and results of the $\mathrm{O}_{3}+$ DMSO rate constant determination experiments are summarized on Table 5, and the relative changes in $\mathrm{O}_{3}$ concentrations at the various $\mathrm{O}_{3}$ levels are shown on Figure 1. Figure 1 also shows the relative changes of DMSO calculated for the average $\mathrm{O}_{3}$ concentration of Run $1(\sim 54 \mathrm{ppm})$ if the $\mathrm{O}_{3}+$ DMSO rate constant were $3 \times 10^{-20} \mathrm{~cm}^{3} \mathrm{molec}^{-1} \mathrm{~s}^{-1}$.

Although the first DMSO measurements in all the experiments appeared to be anomalous (perhaps due to incomplete mixing), essentially no change in measured DMSO levels occurred in the subsequent measurements, especially in Run 1, which had the highest $\mathrm{O}_{3}$ concentration. The slow DMSO decay in Run 2 could not be due to an $\mathrm{O}_{3}$ reaction because that run had the lowest $\mathrm{O}_{3}$ levels of all three experiments. The relative DMSO decay rate calculated using the $\mathrm{O}_{3}+$ DMSO rate constant of $3 \times 10^{-20}$ $\mathrm{cm}^{3}$ molec $^{-1} \mathrm{~s}^{-1}$ and the $\mathrm{O}_{3}$ level of Run 1 is clearly much higher than that observed in the experiment, indicating that the $\mathrm{O}_{3}+$ DMSO rate constant must be less than that. Since decay rates calculated with lower rate constants may be within the scatter of the data, the rate constant of $3 \times 10^{-20} \mathrm{~cm}^{3} \mathrm{molec}^{-1} \mathrm{~s}^{-1}$ is taken as the upper limit as indicated by our data.

The upper limit rate constant determined in this work is about a factor of 3 lower than the upper limit of Falbe-Hansen et al (2000). This corresponds an average atmospheric half life of over a year, based on the average $\mathrm{O}_{3}$ concentration used in the tropospheric lifetime estimates of Falbe-Hansen et al (2000). This confirms that reaction with $\mathrm{O}_{3}$ is a negligible atmospheric loss process for DMSO.

## Environmental Chamber Experiments

## Summary of Experiments

Table 6 gives a chronological listing of all the environmental chamber experiments carried out for this program. These consisted primarily of incremental reactivity and DMSO - NO ${ }_{x}$ experiments, whose conditions and selected results are summarized on Table 7, and which are discussed in more detail in the following sections. In addition, several control and characterization runs were carried out to determine the chamber-dependent inputs needed for the model simulations of the experiments and to assure consistency with previous results. The results of these experiments, summarized in Table 6, indicated that there were no significant problems with chamber characterization or conditions during the course of this study. See Carter (1995c) and references therein for more detailed discussions of the chamber characterization experiments and methods.

Table 5. Summary of conditions and results of the $\mathrm{O}_{3}$ rate constant determination experiments.

| Run | 1 |  | 2 |  | 3 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Avg. O3 (ppm) | 54.3 |  | 5.0 |  | 14.5 |  |
| Avg. Cyc-C6 (ppm) | $178 \pm 2$ |  | $137 \pm 2$ |  | $128 \pm 2$ |  |
| Avg Temp (K) | $294.2 \pm 0.4$ |  | $294.2 \pm 0.5$ |  | $293.9 \pm 0.2$ |  |
| Relative Injection times (min) [a] |  |  |  |  |  |  |
| O3 | -22 |  | <-140 |  | -21 |  |
| Cyclohexane | -42 |  | -140 |  | [b] |  |
| DMSO | -7 |  | -10 |  | [b] |  |
| DMSO Data | Time | ppm | Time | ppm | Time | ppm |
|  | 0 | 0.246 | 0 | 0.339 | 0 | 0.292 |
|  | 28 | 0.197 | 22 | 0.447 | 24 | 0.321 |
|  | 52 | 0.196 | 50 | 0.444 | 49 | 0.320 |
|  | 76 | 0.198 | 75 | 0.436 |  |  |
|  | 98 | 0.195 | 122 | 0.423 |  |  |

[a] Times are relative to the time of the first DMSO measurement.
[b] Continuation of Run 2, with only $\mathrm{O}_{3}$ injected. First DMSO measurement made 130 minutes after first DMSO measurement in Run 2.



Figure 1. Plots of experimental and calculated relative DMSO concentrations atainst time in the $\mathrm{O}_{3}$ + DMSO rate constant determination experiments. The calculated values are based on the $\mathrm{O}_{3}$ concentration in Run 1 and an upper limit $\mathrm{k}_{03+\text { DMSO }}$ of $3 \times 10^{-20} \mathrm{~cm}^{3}$ molec $^{-1} \mathrm{~s}^{-1}$.

Table 6. Chronological listing of the environmental chamber experiments carried out for this program.

| Run No. | Date | Title | Comments |
| :---: | :---: | :---: | :---: |
| DTC751 | 12/22/98 | n-Butane + Chlorine Actinometry | Run to measure the light intensity by determining the $\mathrm{Cl}_{2}$ photolysis rate, as discussed by Carter et al (1995c). The results yielded a calculated $\mathrm{NO}_{2}$ photolysis rate of 0.153 $\mathrm{min}^{-1}$, which is reasonably consistent with the results of the quartz tube actinometry experiments carried out previously, which indicated an $\mathrm{NO}_{2}$ photolysis rate of $0.16 \mathrm{~min}^{-1}$. |
| DTC752 | 1/5/99 | n-Butane + NOx | Run to measure the rate of the chamber radical source, as discussed by Carter et al (1995c). Results are reasonably well simulated using the standard chamber model assigned to this series of experiments (see Table A-4), though Side B has a somewhat higher radical source than Side A. |
| DTC766 | 2/5/99 | Mini Surrogate + DMSO | Standard mini-surrogate - $\mathrm{NO}_{\mathrm{x}}$ reactivity experiment with $\sim 0.5 \mathrm{ppm}$ of DMSO injected into Side A. Conditions and selected results are summarized on Table 7. The results of this experiment are not used for mechanism evaluation because the DMSO injection procedure and analysis is considered to be more uncertain in subsequent experiments. However, Table 7 shows the incremental reactivities in this experiment are close to those observed in the other minisurrogate + DMSO run, DTC783. |
| DTC767 | 2/8/99 | n -Butane +NOx | Run to measure the rate of the chamber radical source. Results are simulated very well using the standard chamber model assigned to this series of experiments (see Table A4), and good side equivalency is observed. This indicates that that the magnitude of the chamber radical source is in the normal range, and that the side differences observed in DTC752 are no longer occurring. |
| DTC768 | 2/9/99 | Mixed surrogate + DMSO (not used for model evaluation) | This was intended to be a standard mini-surrogate experiment, but the n-octane, toluene, m-xylene liquid mixture used in the full surrogate experiments was used instead of the $n$-hexane, $m$-xylene mixture for the minisurrogate runs. The DMSO was injected into Side B. Because of uncertainties in the DMSO injection method and analysis and the other problems with the run, the results are not used for mechanism evaluation. |

Table 6 (continued)

| Run No. | Date | Title | Comments |
| :---: | :---: | :---: | :---: |
| DTC780 | 3/2/99 | Full Surrogate + DMSO | Standard high $\mathrm{NO}_{\mathrm{x}}$ full surrogate reactivity experiment with 0.06 ppm of DMSO added to Side B. Conditions and selected results are summarized on Table 7. Because of uncertainties in DMSO injection and analysis method, the results of this run are not used for mechanism evaluation. However, Table 7 shows that the measured incremental reactivities are reasonably consistent with the results of the other full surrogate + DMSO experiment (Run DTC786). |
|  | $\begin{gathered} \text { Around } \\ 4 / 99 \end{gathered}$ |  | DMSO injection and calibration method changed. DMSO analyses for subsequent runs verified by using a total sulfur analyzer. |
| DTC781 | 4/21/99 | Full Surrogate Side Equivalency Test | Same full surrogate - $\mathrm{NO}_{\mathrm{x}}$ mixture was irradiated on both sides to determine equivalency of the results in the two reactors. Due to an injection error, the concentrations of the gaseous organic reactants were approximately half the normal values, so only 0.2 ppm of $\mathrm{O}_{3}$ was formed. Good side equivalency was observed. |
| DTC782 | 4/22/99 | Low NOx Full <br> Surrogate + DMSO | Standard low $\mathrm{NO}_{\mathrm{x}}$ full surrogate reactivity experiment with 0.18 ppm of DMSO added to Side B. Conditions and selected results are summarized on Table 7, and plots of selected results are shown on Figure 4. |
| DTC783 | 4/23/99 | Mini-Surrogate + DMSO | Standard mini-surrogate - $\mathrm{NO}_{\mathrm{x}}$ reactivity experiment with 0.35 ppm of DMSO injected into Side A. Conditions and selected results are summarized on Table 7, and plots of selected results are shown on Figure 4. |
| DTC784 | 4/27/99 | n-Butane +NOx | Run to measure the rate of the chamber radical source. Results are well simulated using the standard chamber model assigned to this series of experiments (see Table A- <br> 4). The NO oxidation rate on Side B was slightly greater than in Side A but the difference was not significant. |
| DTC785 | 4/28/99 | $\mathrm{DMSO}+\mathrm{NOx}$ | Approximately 0.3 ppm of DMSO injected into both sides of the chamber, 0.13 ppm of $\mathrm{NO}_{\mathrm{x}}$ injected into Side A and 0.13 ppm injected into Side B. Results are summarized on Table 7 and shown on Figure 2. |
| DTC786 | 4/29/99 | Full Surrogate + DMSO | Standard high $\mathrm{NO}_{\mathrm{x}}$ full surrogate reactivity experiment with 0.27 ppm of DMSO added to Side B. Conditions and selected results are summarized on Table 7, and plots of selected results are shown on Figure 4. |
| DTC787 | 4/30/99 | Low NOx Full <br> Surrogate + DMSO | Standard low $\mathrm{NO}_{\mathrm{x}}$ full surrogate reactivity experiment with 0.09 ppm of DMSO added to Side B. Conditions and selected results are summarized on Table 7, and plots of selected results are shown on Figure 4. |

Table 6 (continued)

| Run No. | Date | Title | Comments |
| :--- | :---: | :--- | :--- |
| DTC788 | $5 / 7 / 99$ | DMSO + NOx | Approximately 0.15 ppm of DMSO injected into both sides <br> of the chamber, 0.15 ppm of $\mathrm{NO}_{\mathrm{x}}$ injected into Side A and <br>  |
|  |  | 0.3 ppm injected into Side B. Results are summarized on <br> Table 7 and shown on Figure 2. |  |

## Results of DMSO - NO $\mathbf{x}_{\mathrm{x}}$ Experiments

Two dual-chamber DMSO - $\mathrm{NO}_{\mathrm{x}}$ experiments were carried out during the course of this program to provide data to evaluate the mechanism for DMSO in the absence of other reactants. Such experiments do not provide useful data for compounds that do not have significant internal radical sources (see Carter et al, 1982; Carter and Lurmann, 1991) because they tend to be dominated by the chamber radical source, so they were not included in the original work plan for this project. However, the results of the DMSO reactivity experiments, discussed in the following section, indicated that DMSO does have significant internal radical sources, so the $\mathrm{DMSO}-\mathrm{NO}_{x}$ experiments were included in this project. Although these runs were carried out around the end of the project, the results will be discussed first because they represent simpler chemical systems.

The four DMSO - $\mathrm{NO}_{\mathrm{x}}$ mixtures irradiated provided mechanism evaluation data at different $\mathrm{NO}_{\mathrm{x}}$, and DMSO concentrations and at different $\mathrm{DMSO} / \mathrm{NO}_{\mathrm{x}}$ ratios. Concentration-time plots of selected species measured during these experiments are shown on Figure 2. Results of model calculations using the three mechanisms ("C", "F", and "B") that give the best simulations of the data are also shown on that figure. Figure 3 gives plots of the same data, except showing the model calculations using the other three of the six alternative mechanisms that were examined.

The figures shows that these $\mathrm{DMSO}-\mathrm{NO}_{\mathrm{x}}$ systems are highly reactive, with relatively rapid NO oxidation and $\mathrm{O}_{3}$ formation in all experiments except the run with the highest $\mathrm{NO}_{\mathrm{x}}$ and lowest DMSO levels. This is despite the fact that the $\mathrm{DMSO} / \mathrm{NO}_{\mathrm{x}}$ ratios are relatively low in these experiments, ranging from $\sim 1$ to 4 , on a carbon basis. Most of the DMSO was oxidized within 3 or 4 hours, with essentially all of the DMSO reacted at the end of the 6 hour experiments. Relatively large amounts of formaldehyde were formed in these experiments, with final yields comparable to the initial DMSO. Since formaldehyde also reacts relatively rapidly in these experiments, this indicates that more than one mole of formaldehyde must be formed for each mole of DMSO that reacts.

The total gas-phase sulfur analyzer had a $100 \%$ response to DMSO, as indicated by the relatively good agreement between the measurements using this instrument at the beginning of the experiments and the initial DMSO concentrations as measured by GC. However, during the course of the experiment, the gas-phase sulfur declined to a much lesser extent than the DMSO, with the final values being about half

Table 7, Summary of conditions and selected results of environmental chamber experiments with DMSO.

| Run | $\begin{gathered} \text { DMSO } \\ (\mathrm{ppm}) \end{gathered}$ | $\begin{gathered} \mathrm{NO}_{\mathrm{x}} \\ (\mathrm{ppm}) \end{gathered}$ | Surg. (ppm C) | $\Delta\left(\left[\mathrm{O}_{3}\right]\right.$ [ NO$\left.]\right)(\mathrm{ppm})$ |  |  |  |  |  | $\begin{gathered} 5^{\text {th }} \text { Hour IntOH } \\ \left(10^{-6} \mathrm{~min}\right) \end{gathered}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $2^{\text {nd }}$ Hour |  |  | $5^{\text {th }}$ Hour |  |  |  |  |  |
|  |  |  |  | Base | Test | IR [a] | Base | Test | IR [a] | Base | Test | IR [a] |
| Mixed Surrogate [b] |  |  |  |  |  |  |  |  |  |  |  |  |
| DTC768B [c] | $\sim 0.12$ ? | 0.37 | 3.91 | 0.08 | 0.13 | $\sim 0.5$ ? | 0.35 | 0.44 | $\sim 0.4$ ? | 13 | 11 | $\sim-11$ ? |
| Mini-Surrogate |  |  |  |  |  |  |  |  |  |  |  |  |
| DTC766A [c] | $\sim 0.52$ ? | 0.37 | 5.83 | 0.08 | 0.57 | $\sim 0.9$ ? | 0.40 | 0.90 | $\sim 1.0$ ? | 10 | 16 | $\sim 12$ ? |
| DTC783B | 0.35 | 0.39 | 6.06 | 0.13 | 0.56 | 1.2 | 0.45 | 0.74 | 0.81 | 13 | 17 | 10 |
| High $\mathrm{NO}_{x}$ Full Surrogate |  |  |  |  |  |  |  |  |  |  |  |  |
| DTC780B [c] | $\sim 0.11$ ? | 0.30 | 4.16 | 0.25 | 0.37 | $\sim 1.0$ ? | 0.50 | 0.60 | $\sim 0.8$ ? | 21 | 19 | $\sim-16$ ? |
| DTC786A | 0.27 | 0.29 | 4.54 | 0.23 | 0.63 | 1.5 | 0.44 | 0.67 | 0.85 | 22 | 21 | -2 |
| Low $\mathrm{NO}_{\underline{\chi}}$ Full Surrogate |  |  |  |  |  |  |  |  |  |  |  |  |
| DTC782A | 0.18 | 0.09 | 4.69 | 0.26 | 0.35 | 0.5 | 0.26 | 0.33 | 0.38 | 18 | 9 | -53 |
| DTC787B | 0.09 | 0.10 | 4.26 | 0.25 | 0.31 | 0.7 | 0.26 | 0.30 | 0.45 | 19 | 14 | -58 |
| DMSO - $\mathrm{NO}_{\underline{x}}$ |  |  |  |  |  |  |  |  |  |  |  |  |
| DTC785B | 0.31 | 0.24 | - | - | 0.33 | - | - | 0.58 | - | - | - | - |
| DTC785A | 0.27 | 0.13 | - | - | 0.32 | - | - | 0.43 | - | - | - | - |
| DTC788B | 0.15 | 0.27 | - | - | 0.15 | - | - | 0.28 | - | - | - | - |
| DTC788A | 0.13 | 0.16 | - | - | 0.16 | - | - | 0.33 | - | - | - | - |

[a] IR $=$ Incremental Reactivity $=([$ Test $]-[$ Base $]) /[$ DMSO $]$
[b] Non-standard surrogate mixture employed because of an injection error
[c] Amounts of DMSO added and DMSO analysis is uncertain. Run not used for mechanism evaluation.
the initial concentration. (The gas-phase sulfur data in the reactivity experiments, where the initial DMSO was also completely consumed, were similar.) This can be attributed to the formation of $\mathrm{SO}_{2}$ in the oxidation of DMSO, to which the instrument would also respond. However, it is clear that DMSO is forming sulfur-containing oxidation products that are not measured on this analyzer, apparently in approximately $50 \%$ yields. This could be due to the formation of MSA or sulfate, which presumably would be lost on the walls of the filter before being detected by this instrument. It is unknown whether the instrument would respond to gas-phase $\mathrm{DMSO}_{2}$ that is expected to be formed in at least $\sim 25 \%$ yields in this system.

Figure 2 and Figure 3 show that all six of the alternative mechanisms tend to underpredict the NO oxidation and $\mathrm{O}_{3}$ formation rates in these experiments, and also underpredict the final $\mathrm{O}_{3}$ yields except for the simulation of DTC785A by Mechanism "C". The DMSO consumption rates are also underpredicted in all cases, with the mechanisms shown on Figure 2 being the best performing in that regard. Overall Mechanism " C " performs the least poorly in this regard, underpredicting the NO oxidation and $\mathrm{O}_{3}$


Figure 2. Experimental and calculated concentration-time plots for selected species in the DMSO$\mathrm{NO}_{\mathrm{x}}$ environmental chamber experiments. Calculations are for the three best performing mechanisms.


Figure 3. Experimental and calculated concentration-time plots for selected species in the DMSO$\mathrm{NO}_{\mathrm{x}}$ environmental chamber experiments. Calculations are for the three worst performing mechanisms
formation rates by about a factor of 1.5 in these experiments. It also predicts that not all the DMSO is consumed in runs DTC788A and B, contrary to the experimental observations.

All the mechanisms except " B " also tend to underpredict the formaldehyde yields in these experiments, though (except for " B ") Mechanism " C " is least unsatisfactory in this regard. Mechanism " C " gives reasonably good fits to the final total sulfur readings at the end of the experiments for the runs where it correctly predicts that most of the DMSO is consumed, if it is assumed that $\mathrm{DMSO}_{2}$ is not measured as gas-phase sulfur. These points are discussed further below.

## Results of the Incremental Reactivity Experiments

The results of the seven incremental reactivity experiments carried out for this program are summarized on Table 7. As indicated in Table 6, the first runs done with DMSO, DTC766, DTC768 and DTC780 were judged not to be useful for mechanism evaluation because of uncertainties in the DMSO injection. However, as noted in Table 6, the incremental reactivities observed in runs DTC766 and DTC780 were qualitatively similar to those observed in the comparable experiment that was subsequently carried out. Figure 4 shows plots of the major results of these experiments for the four runs that are used for mechanism evaluation. The figure also shows the results of the model simulations using the Mechanisms "C" and " F ", the two mechanisms that gave the least unsatisfactory simulations of the DMSO - $\mathrm{NO}_{\mathrm{x}}$ experiments discussed above. Figure 5 shows representative results of model calculations of selected data from selected experiments using the other four alternative mechanisms, with the top series of plots giving the fits for Mechanisms "A" and "B", and the bottom showing the fits for "D" and "E".

Figure 4 and Figure 5 show that the three "best fit" mechanisms fit the results of the reactivity experiments somewhat better than they fit the results of the DMSO - $\mathrm{NO}_{\mathrm{x}}$ experiments, particularly the reactivity experiments with the more realistic "full surrogate" base ROG mixture. The tendency of the mechanisms to underpredict the IntOH reactivities in the low $\mathrm{NO}_{\mathrm{x}}$ full surrogate experiments is observed with almost all VOCs (see Carter, 2000), and is attributed to possible problems with the representation of low $\mathrm{NO}_{\mathrm{x}}$ conditions in the base mechanism. Therefore, this does not necessarily indicate problems with the mechanism for the test compound. The three best fit mechanisms give quite good predictions of the $\Delta\left(\left[\mathrm{O}_{3}\right]-[\mathrm{NO}]\right)$ reactivities in the full surrogate runs and of the IntOH reactivities in the reactivity experiments with the higher $\mathrm{NO}_{\mathrm{x}}$ conditions. However, consistent with its simulations of the DMSO $\mathrm{NO}_{\mathrm{x}}$ experiments, the mechanisms tend to underpredict the effect of DMSO on NO oxidation and $\mathrm{O}_{3}$ formation rates in the mini-surrogate reactivity experiment.

Consistent with the results of the simulations of the DMSO - NO $x$ runs, Mechanisms "D", and "E" tend to significantly underpredict the DMSO's $\Delta\left(\left[\mathrm{O}_{3}\right]-[\mathrm{NO}]\right)$ and IntOH reactivities in these experiments. However, unlike the result with the DMSO - $\mathrm{NO}_{\mathrm{x}}$ experiments, the reactivity results with Mechanism "A" are only slightly different from those with Mechanism "B".


Figure 4 Experimental and calculated results of the incremental reactivity experiments with DMSO. Calculations are for the two "best fit" mechanisms.


Figure 5. Experimental and calculated results of selected incremental reactivity experiments with DMSO. Calculations are for Mechanisms A, B, D, and E.

The performance of the mechanisms in predicting the effects of DMSO on formaldehyde yields in the reactivity experiments was consistent with the predictions for the DMSO - $\mathrm{NO}_{\mathrm{x}}$ runs. In particular, all the mechanisms other than " B " consistently underpredicted the effects of DMSO on formaldehyde, though the underprediction was not large for the best fit mechanism "C".

## Mechanistic Implications

Of the six alternative mechanisms examined, the mechanisms that gave the best fit to most of the reactivity data were the ones that predicted the highest reactivity characteristics for DMSO. The main contributors to ozone reactivity are number of NO to $\mathrm{NO}_{2}$ conversions involved in the oxidation of the reactant, the radical initiation or termination characteristics of the reactions, the tendency of the VOC to enhance or remove $\mathrm{NO}_{\mathrm{x}}$ levels, and the reactivities of the products. Mechanism "C" is the most reactive of the mechanisms examined because it involves the largest number of NO to $\mathrm{NO}_{2}$ conversions and the highest formaldehyde yield in the OH reaction, consistent with the assumption of $20 \% \mathrm{DMSO}_{2}$ formation. High yields of formaldehyde cause relatively high reactivity because the formaldehyde photolysis to form radicals is a significant radical initiation process.

Mechanism " $F$ " predicts lower reactivity than the high reactivity mechanism " $C$ " because it assumes significant MSA formation in the OH reaction, which means lower formaldehyde yields. Mechanism " B " predicts the same formaldehyde yields in the OH reaction as mechanism "C", but predicts fewer NO to $\mathrm{NO}_{2}$ conversions because the oxidation of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})$. to $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O}$. is assumed to involve $\mathrm{NO}_{2}$ to NO conversions, rather than vise-versa, as is the case with " C ". Mechanism " A ", which involves the initial formation and subsequent reaction of MSIA in the $\mathrm{OH}+$ DMSO reaction, involves the same overall number of NO to $\mathrm{NO}_{2}$ conversions and formaldehyde yield once the MSIA reacts as does Mechanism "C". However, although the MSIA is assumed to react with OH with an almost gas kinetic rate constant, the delay caused by MSIA formation and reaction causes a substantial reduction in predicted rates of NO oxidation and $\mathrm{O}_{3}$ formation.

Mechanisms "D" and "E" predict substantially lower DMSO reactivity than the other mechanisms they both have non-negligible radical termination processes. In the case of Mechanism "E" this is the formation of MSPN, which (unlike Mechanisms "D" and "F", where MSPN formation is also assumed to be important), is assumed to decompose relatively slowly. The most inhibiting mechanism is "D", which assumes significant radical loss by formation of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{O}$, which (unlike Mechanism " F ") is assumed not to regenerate radicals. The very poor performance of these mechanisms in fitting the DMSO reactivity data indicate that these radical terminating processes cannot be important under the conditions of our experiments.

The model predicted that reaction of DMSO with $\mathrm{NO}_{3}$ was a non-negligible process under the conditions of our environmental chamber experiments. For most experiments, the fraction of DMSO reacting with $\mathrm{NO}_{3}$ was predicted to be in the $25 \%-35 \%$ range, except for the mini-surrogate experiments where $\mathrm{NO}_{3}$ reaction was predicted to occur $\sim 50 \%$ of the time, and for runs DTC788A and B, where it was
predicted to be relatively less important. Note that all the alternative mechanisms assume a relatively unreactive $\mathrm{DMSO}+\mathrm{NO}_{3}$ mechanism, involving formation of an unreactive product $\left(\mathrm{DMSO}_{2}\right)$ and the net destruction of $\mathrm{O}_{3}$ (by conversion of $\mathrm{NO}_{3}$, formed from $\mathrm{O}_{3}+\mathrm{NO}_{2}$, to $\mathrm{NO}_{2}$.) However, assuming a more reactive mechanism, e.g.,

$$
\begin{equation*}
\mathrm{DMSO}+\mathrm{NO}_{3} \rightarrow \mathrm{HNO}_{3}+\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2} . \tag{21}
\end{equation*}
$$

occurs at a significant rate results in predictions of incorrect shapes in $\mathrm{O}_{3}$ concentration-time profiles and higher peak $\mathrm{O}_{3}$ concentrations than observed in the DMSO - $\mathrm{NO}_{\mathrm{x}}$ experiments than are sensitive to the $\mathrm{NO}_{3}$ reaction. Using a lower DMSO $+\mathrm{NO}_{3}$ rate constant slightly improves the mechanism predictions of formaldehyde yields by increasing the relative importance of the OH reaction that forms formaldehyde, but does not significantly increase the $\mathrm{O}_{3}$ formation rate. The effects of these changes in the DMSO + $\mathrm{NO}_{3}$ mechanism and rate constant are shown on Figure 6 for the two $\mathrm{DMSO}+\mathrm{NO}_{\mathrm{x}}$ experiments that are the most sensitive to them. Note that the alternative assumptions on the $\mathrm{DMSO}+\mathrm{NO}_{3}$ mechanism have very little effects on the simulations of runs DTC788A and DTC788B, which are not shown. In particular, the alternative mechanisms still underpredict the $\mathrm{O}_{3}$ formation rates and yields in these experiments.


Figure 6. Effects of alternative concerning the mechanism and rate constant for the reactions of DMSO with $\mathrm{NO}_{3}$ on model simulations of the two $\mathrm{DMSO}+\mathrm{NO}_{x}$ experiments that are the most sensitive to this reaction.

The role of the $\mathrm{NO}_{3}+$ DMSO reaction is the reason that Mechanism "C" predicts higher overall formaldehyde yields than does Mechanisms "C", despite the fact that the overall formaldehyde yield in the OH reaction is the same. The reduced number of NO to $\mathrm{NO}_{2}$ conversions predicted by Mechanism "C" means that it predicts lower $\mathrm{O}_{3}$ and thus lower $\mathrm{NO}_{3}$ radical levels, and thus lower rates of reaction of DMSO with $\mathrm{NO}_{3}$. On the other hand it predicts similar formaldehyde and thus overall radical levels, and thus the predicted OH radical levels are approximately the same. This means that relatively more of the DMSO consumption is reaction with OH radicals, forming formaldehyde, than is the case for the other mechanisms.

As discussed above, the reaction of $\mathrm{NO}_{3}$ with DMSO is expected to be less important under atmospheric conditions than in these experiments because of the relatively rapid photolysis rates for $\mathrm{NO}_{3}$ in sunlight compared to blacklight irradiation (Carter et al, 1995d). Therefore, the uncertainties concerning this reaction probably do not have a large impact on the atmospheric reactivity simulations discussed in the following section.

The fact that even the most reactive mechanism (Mechanism "C") tends to underpredict $\mathrm{O}_{3}$ formation rates in the DMSO $-\mathrm{NO}_{\mathrm{x}}$ experiments suggests that there may be other radical sources in the DMSO oxidation system that are not being represented in the mechanisms being considered. As discussed above, the possibility of radical formation in the $\mathrm{DMSO}+\mathrm{NO}_{3}$ reaction was considered, but this gives predictions that are not consistent with the data. Radical formation from reaction of DMSO with $\mathrm{O}_{3}$ cannot be significant given the low $\mathrm{DMSO}+\mathrm{O}_{3}$ rate constant measured in this and previous studies. Assuming higher yields of formaldehyde by reducing the $\mathrm{DMSO}_{2}$ yield gives predictions that are more consistent with the formaldehyde data but still results in a tendency to underpredict $\mathrm{O}_{3}$ formation rates in most of the experiments.

Although the best fit Mechanism "C" is not satisfactory in all respects in that it has a bias towards underpredicting $\mathrm{O}_{3}$ formation rates in the $\mathrm{DMSO}+\mathrm{NO}_{\mathrm{x}}$ and the mini-surrogate incremental reactivity experiments, it gives reasonably good simulations of incremental ozone reactivities in the full surrogate experiments. This is important in terms of the suitability of this mechanism for atmospheric reactivity simulations, since the chemical conditions of the full surrogate experiments are more representative of those in the atmosphere than is the case for the other types of experiments. The more realistic chemical conditions of the full surrogate runs appear to be less sensitive to whatever errors or omissions in the DMSO mechanisms are causing the biases in the simulations of the $\mathrm{DMSO}+\mathrm{NO}_{\mathrm{x}}$ or the mini-surrogate runs. For that reason, it may not be inappropriate to use Mechanism "C" as the basis for estimating the impacts of DMSO on ozone formation in the atmosphere.

## ATMOSPHERIC REACTIVITY CALCULATIONS

Incremental reactivities of VOCs have been shown to be highly dependent on environmental conditions, so reactivities measured in environmental chamber experiments cannot necessarily be assumed to be the same as those under atmospheric conditions (Carter and Atkinson, 1989; Carter et al, 1995b). Because of this, the only method available to obtain quantitative estimates of incremental reactivities of VOCs in ambient air pollution episodes is to conduct airshed model simulations of the episodes. Since these simulations cannot be any more reliable than the chemical mechanisms used, the major objective of this program was to assess the reliability of the DMSO mechanisms for use in such calculations. As discussed above, the results of this study suggest that DMSO Mechanism "C" may serve as an appropriate basis for estimating the effects of DMSO on ozone under atmospheric conditions. The estimates based on this mechanism are discussed in this section.

## Scenarios Used for Reactivity Assessment

The set of airshed scenarios employed to assess the DMSO reactivities for this study is the same as those used for calculating the MIR and other reactivity scales in our previous work (Carter, 1994a), and also in the update using the SAPRC-99 mechanism (Carter, 2000). These scenarios, and the reasons for using them, are briefly described below.

The objective is to use a set of scenarios which represents, as much as possible, a comprehensive distribution of the environmental conditions where unacceptable levels of ozone are formed. Although a set of scenarios has not been developed for the specific purpose of VOC reactivity assessment, the EPA developed an extensive set of scenarios for conducting analyses of effects of ROG and $\mathrm{NO}_{\mathrm{x}}$ controls on ozone formation using the EKMA modeling approach (Gipson et al. 1981; Gipson and Freas, 1983; EPA, 1984; Gery et al. 1987; Baugues, 1990). The EKMA approach involves the use of single-cell box models to simulate how the ozone formation in one day episodes is affected by changes in ROG and $\mathrm{NO}_{\mathrm{x}}$ inputs. Although single-cell models cannot represent realistic pollution episodes in great detail, they can represent dynamic injection of pollutants, time-varying changes of inversion heights, entrainment of pollutants from aloft as the inversion height raises, and time-varying photolysis rates, temperatures, and humidities (Gipson and Freas, 1981; EPA, 1984; Gipson, 1984; Hogo and Gery, 1988). Thus, they can be used to simulate a wide range of the chemical conditions which affect ozone formation from ROG and $\mathrm{NO}_{x}$, and which affect VOC reactivity. Therefore, at least to the extent they are suitable for their intended purpose, an appropriate set of EKMA scenarios should also be suitable for assessing reactivities over a wide range of conditions.

## Base Case Scenarios

The set of EKMA scenarios used in this study were developed by the United States EPA for assessing how various ROG and $\mathrm{NO}_{\mathrm{x}}$ control strategies would affect ozone nonattainment in various areas
of the country (Baugues, 1990). The characteristics of these scenarios and the methods used to derive their input data are described in more detail elsewhere (Baugues, 1990; Carter, 1994b). Briefly, 39 urban areas in the United States were selected based on geographical representativeness of ozone nonattainment areas and data availability, and a representative high ozone episode was selected for each. The initial nonmethane organic carbon (NMOC) and $\mathrm{NO}_{\mathrm{x}}$ concentrations, the aloft $\mathrm{O}_{3}$ concentrations, and the mixing height inputs were based on measurement data for the various areas, the hourly emissions in the scenarios were obtained from the National Acid Precipitation Assessment Program emissions inventory (Baugues, 1990), and biogenic emissions were also included. Table 8 gives a summary of the urban areas represented and other selected characteristics of the scenarios.

Several changes to the scenario inputs were made based on discussions with the California ARB staff and others (Carter, 1994a,b). Two percent of the initial $\mathrm{NO}_{\mathrm{x}}$ and $0.1 \%$ of the emitted $\mathrm{NO}_{\mathrm{x}}$ in all the scenarios was assumed to be in the form of HONO. The photolysis rates were calculated using solar light intensities and spectra calculated by Jeffries (1991) for 640 meters, the approximate mid-point of the mixed layer during daylight hours. The composition of the NMOCs entrained from aloft was based on the analysis of Jeffries et al. (1989). The composition of the initial and emitted reactive organics was derived as discussed below. Complete listings of the input data for the scenarios are given elsewhere (Carter, 1994b).

This set of 39 EKMA scenarios are referred to as "base case" to distinguish them from the scenarios derived from them by adjusting $\mathrm{NO}_{\mathrm{x}}$ inputs to yield standard conditions of $\mathrm{NO}_{\mathrm{x}}$ availability as discussed below. No claim is made as to the accuracy of these scenarios in representing any real episode, but they are a result of an effort to represent, as accurately as possible given the available data and the limitations of the formulation of the EKMA model, the range of conditions occurring in urban areas throughout the United States. When developing general reactivity scales it is more important that the scenarios employed represent a realistic distribution of chemical conditions than accurately representing the details of any one particular episode.

The Base ROG mixture is the mixture of reactive organic gases used to represent the chemical composition of the initial and emitted anthropogenic reactive organic gases from all sources in the scenarios. Consistent with the approach used in the original EPA scenarios, the same mixture was used for all scenarios. The speciation for this mixture was derived by Croes (1991) based on an analysis of the EPA database (Jeffries et al. 1989) for the hydrocarbons and the 1987 Southern California Air Quality Study (SCAQS) database for the oxygenates (Croes et al. 1994; Lurmann and Main 1992). This mixture consists of $52 \%$ (by carbon) alkanes, $15 \%$ alkenes, $27 \%$ aromatics, $1 \%$ formaldehyde, $2 \%$ higher aldehydes, $1 \%$ ketones, and $2 \%$ acetylene. The detailed composition of this mixture is given elsewhere (Carter, 1994b; Carter, 2000).

Table 8. Summary of the conditions of the scenarios used for atmospheric reactivity assessment.

|  | Scenario | Max $\mathrm{O}_{3}$ (ppb) | Max 8Hr Avg $\mathrm{O}_{3}(\mathrm{ppb})$ | $\begin{aligned} & \mathrm{ROG} \\ & / \mathrm{NO}_{\mathrm{x}} \end{aligned}$ | $\begin{gathered} \mathrm{NO}_{\mathrm{x}} \\ / \mathrm{MOIR}^{2} \\ \mathrm{NO}_{\mathrm{x}} \end{gathered}$ | Height <br> (kM) | Init., Emit ROG (m. $\mathrm{mol} \mathrm{m}{ }^{-2}$ ) | $\begin{gathered} \mathrm{O}_{3} \text { aloft } \\ (\mathrm{ppb}) \end{gathered}$ | Integrated OH (ppt-min) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Avg. | MIR | 187 | 119 | 3.1 | 1.5 | 1.8 | 15 | 70 | 128 |
| Cond. | MOIR | 239 | 165 | 4.5 | 1.0 | 1.8 | 15 | 70 | 209 |
|  | EBIR | 227 | 172 | 6.4 | 0.7 | 1.8 | 15 | 70 | 210 |
| Base | Atlanta, GA | 179 | 132 | 7.3 | 0.7 | 2.1 | 12 | 63 | 200 |
| Case | Austin, TX | 175 | 144 | 9.3 | 0.5 | 2.1 | 11 | 85 | 179 |
|  | Baltimore, MD | 334 | 215 | 5.2 | 1.1 | 1.2 | 17 | 84 | 186 |
|  | Baton Rouge, LA | 241 | 173 | 6.8 | 0.9 | 1.0 | 11 | 62 | 186 |
|  | Birmingham, AL | 244 | 202 | 6.9 | 0.5 | 1.8 | 13 | 81 | 208 |
|  | Boston, MA | 197 | 167 | 6.5 | 0.6 | 2.6 | 14 | 105 | 262 |
|  | Charlotte, NC | 143 | 126 | 7.8 | 0.3 | 3.0 | 7 | 92 | 212 |
|  | Chicago, IL | 278 | 226 | 11.6 | 0.5 | 1.4 | 25 | 40 | 164 |
|  | Cincinnati, OH | 205 | 153 | 6.4 | 0.7 | 2.8 | 17 | 70 | 220 |
|  | Cleveland, OH | 252 | 179 | 6.6 | 0.9 | 1.7 | 16 | 89 | 187 |
|  | Dallas, TX | 208 | 141 | 4.7 | 1.2 | 2.3 | 18 | 75 | 176 |
|  | Denver, CO | 204 | 139 | 6.3 | 1.1 | 3.4 | 29 | 57 | 143 |
|  | Detroit, MI | 246 | 177 | 6.8 | 0.7 | 1.8 | 17 | 68 | 235 |
|  | El Paso, TX | 182 | 135 | 6.6 | 1.0 | 2.0 | 12 | 65 | 138 |
|  | Hartford, CT | 172 | 144 | 8.4 | 0.5 | 2.3 | 11 | 78 | 220 |
|  | Houston, TX | 312 | 217 | 6.1 | 0.9 | 1.7 | 25 | 65 | 225 |
|  | Indianapolis, IN | 212 | 148 | 6.6 | 0.9 | 1.7 | 12 | 52 | 211 |
|  | Jacksonville, FL | 155 | 115 | 7.6 | 0.6 | 1.5 | 8 | 40 | 206 |
|  | Kansas City, MO | 159 | 126 | 7.1 | 0.6 | 2.2 | 9 | 65 | 233 |
|  | Lake Charles, LA | 286 | 209 | 7.4 | 0.6 | 0.5 | 7 | 40 | 233 |
|  | Los Angeles, CA | 568 | 406 | 7.6 | 1.0 | 0.5 | 23 | 100 | 134 |
|  | Louisville, KY | 212 | 155 | 5.5 | 0.8 | 2.5 | 14 | 75 | 260 |
|  | Memphis, TN | 229 | 180 | 6.8 | 0.6 | 1.8 | 15 | 58 | 249 |
|  | Miami, FL | 132 | 111 | 9.6 | 0.4 | 2.7 | 9 | 57 | 181 |
|  | Nashville, TN | 167 | 138 | 8.0 | 0.4 | 1.6 | 7 | 50 | 225 |
|  | New York, NY | 365 | 294 | 8.1 | 0.7 | 1.5 | 39 | 103 | 159 |
|  | Philadelphia, PA | 247 | 169 | 6.2 | 0.9 | 1.8 | 19 | 53 | 227 |
|  | Phoenix, AZ | 277 | 193 | 7.6 | 1.0 | 3.3 | 40 | 60 | 153 |
|  | Portland, OR | 166 | 126 | 6.5 | 0.7 | 1.6 | 6 | 66 | 233 |
|  | Richmond, VA | 242 | 172 | 6.2 | 0.8 | 1.9 | 16 | 64 | 217 |
|  | Sacramento, CA | 204 | 142 | 6.6 | 0.8 | 1.1 | 7 | 60 | 209 |
|  | St Louis, MO | 324 | 209 | 6.1 | 1.1 | 1.6 | 26 | 82 | 176 |
|  | Salt Lake City, UT | 186 | 150 | 8.5 | 0.6 | 2.2 | 11 | 85 | 182 |
|  | San Antonio, TX | 133 | 98 | 3.9 | 1.0 | 2.3 | 6 | 60 | 192 |
|  | San Diego, CA | 193 | 150 | 7.1 | 0.9 | 0.9 | 8 | 90 | 146 |
|  | San Francisco, CA | 229 | 126 | 4.8 | 1.8 | 0.7 | 25 | 70 | 61 |
|  | Tampa, FL | 230 | 153 | 4.4 | 1.0 | 1.0 | 8 | 68 | 211 |
|  | Tulsa, OK | 231 | 160 | 5.3 | 0.9 | 1.8 | 15 | 70 | 264 |
|  | Washington, DC | 283 | 209 | 5.3 | 0.8 | 1.4 | 13 | 99 | 239 |

## Adjusted $\mathbf{N O}_{\mathbf{x}}$ scenarios

Incremental reactivities in the base case scenarios would be expected to vary widely, since incremental reactivities depend on the $\mathrm{ROG} / \mathrm{NO}_{\mathrm{x}}$ ratio, and that ratio varies widely among the base case scenarios. To obtain reactivity scales for specified $\mathrm{NO}_{\mathrm{x}}$ conditions, separate scenarios, designated MIR (for maximum incremental reactivity), MOIR (for maximum ozone incremental reactivity), and Equal Benefit Incremental Reactivity (EBIR) were developed (Carter, 1994a). In the MIR scenarios, the $\mathrm{NO}_{\mathrm{x}}$ inputs were adjusted so the base ROG mixture (and most other VOCs) have their highest incremental reactivity. This is representative of the highest $\mathrm{NO}_{\mathrm{x}}$ conditions of relevance to VOC reactivity assessment because at higher $\mathrm{NO}_{\mathrm{x}}$ levels $\mathrm{O}_{3}$ yields become significantly suppressed, but is also the condition where $\mathrm{O}_{3}$ is most sensitive to VOC emissions. In the MOIR scenarios, the $\mathrm{NO}_{\mathrm{x}}$ inputs were adjusted to yield the highest ozone concentration. In the EBIR scenarios, the $\mathrm{NO}_{\mathrm{x}}$ inputs were adjusted so that the relative effects of $\mathrm{NO}_{\mathrm{x}}$ reductions and total ROG reductions on peak ozone levels were equal. This represents the lowest $\mathrm{NO}_{\mathrm{x}}$ condition of relevance for VOC reactivity assessment, because $\mathrm{O}_{3}$ formation becomes more sensitive to $\mathrm{NO}_{\mathrm{x}}$ emissions than VOC emissions at lower $\mathrm{NO}_{\mathrm{x}}$ levels. As discussed by Carter (1994a) the MIR and EBIR ROG/ $\mathrm{NO}_{\mathrm{x}}$ ratios are respectively $\sim 1.5$ and $\sim 0.7$ times those for the MOIR scenarios in all cases.

For this study, the MIR, MOIR, and EBIR reactivities were calculated using the "averaged conditions" scenarios with the corresponding adjusted $\mathrm{NO}_{\mathrm{x}}$ conditions. As discussed by Carter (1994a), averaged conditions scenarios have all inputs derived by averaging the corresponding inputs of the base case scenarios, except that the $\mathrm{NO}_{\mathrm{x}}$ inputs were adjusted to yield the specified $\mathrm{NO}_{\mathrm{x}}$ conditions as discussed above. This is slightly different than the approach used by Carter (1994a, 2000) to derive the MIR, MOIR, and EBIR scales, which involved adjusting $\mathrm{NO}_{x}$ conditions separately for each of the 39 base case scenarios, and then averaging the reactivities derived from them. Since Carter (1994a) showed that both approaches yield essentially the same result. For this work use of the averaged conditions approach is preferred because it is computationally much more straightforward, and gives an equally a good indication of how the relative reactivities of compounds vary with varying $\mathrm{NO}_{x}$ conditions.

## $\mathrm{NO}_{\mathbf{x}}$ Conditions in the Base Case Scenarios

The variability of $\mathrm{ROG} / \mathrm{NO}_{\mathrm{x}}$ ratios in the base case scenarios suggests a variability of reactivity characteristics in those scenarios. However, as discussed previously (Carter, 1994a), the ROG/ $\mathrm{NO}_{\mathrm{x}}$ ratio is also variable in the MIR or MOIR scenarios, despite the fact that the $\mathrm{NO}_{\mathrm{x}}$ inputs in these scenarios are adjusted to yield a specified reactivity characteristic. Thus, the $\mathrm{ROG} / \mathrm{NO}_{\mathrm{x}}$ ratio, by itself, is not necessarily a good predictor of reactivity characteristics of a particular scenario. The $\mathrm{NO}_{\mathrm{x}} / \mathrm{NO}_{\mathrm{x}}{ }^{\text {MOIR }}$ ratio is a much better predictor of this, with values greater than 1 indicating relatively high $\mathrm{NO}_{\mathrm{x}}$ conditions where ozone formation is more sensitive to VOCs, and values less than 1 indicating $\mathrm{NO}_{x}$-limited conditions. $\mathrm{NO}_{\mathrm{x}} / \mathrm{NO}_{\mathrm{x}}{ }^{\text {MOIR }}$ ratios less than 0.7 represent conditions where $\mathrm{NO}_{\mathrm{x}}$ control is a more effective ozone control strategy than ROG control (Carter, 1994a). Note that more than half of the base case scenarios represent $\mathrm{NO}_{\mathrm{x}}$-limited conditions, and $\sim 25 \%$ of them represent conditions where $\mathrm{NO}_{\mathrm{x}}$ control is more
beneficial than VOC control. A relatively small number of scenarios represent MIR or near MIR conditions. However, as discussed elsewhere (Carter, 1994a), this set of scenarios is based on near-worstcase conditions for ozone formation in each of the airsheds. Had scenarios representing less-than-worstcase conditions been included, one might expect a larger number of MIR or near MIR scenarios. This is because $\mathrm{NO}_{\mathrm{x}}$ is consumed more slowly on days with lower light intensity or temperature, and thus the scenario is less likely to become $\mathrm{NO}_{\mathrm{x}}$-limited.

## Quantification of Atmospheric Reactivity

The reactivity of a VOC in an airshed scenario is measured by its incremental reactivity. For ambient scenarios, this is defined as the change in ozone caused by adding the VOC to the emissions, divided by the amount of VOC added, calculated for sufficiently small amounts of added VOC that the incremental reactivity is independent of the amount added ${ }^{1}$.

$$
\begin{equation*}
\mathrm{IR}(\text { VOC,Scenario })=\lim _{V O C \rightarrow 0}\left[\frac{\mathrm{O}_{3}(\text { Scenario with VOC })-\mathrm{O}_{3}(\text { Base Scenario })}{\text { Amount of VOC Added }}\right] \tag{IV}
\end{equation*}
$$

The specific calculation procedure is discussed in detail elsewhere (Carter, 1994a,b).
Incremental reactivities derived as given above tend to vary from scenario to scenario because they differ in their overall sensitivity of $\mathrm{O}_{3}$ formation to VOCs. These differences can be factored out to some extent by using "relative reactivities", which are defined as ratios of incremental reactivities to the incremental reactivity of the base ROG mixture, which is used to represent emissions of reactive VOCs from all sources.

$$
\begin{equation*}
\mathrm{RR}(\text { VOC }, \text { Scenario })=\frac{\mathrm{IR}(\text { VOC, Scenario })}{\operatorname{IR}(\text { Base ROG, Scenario })} \tag{V}
\end{equation*}
$$

These relative reactivities can also be thought of as the relative effect on $\mathrm{O}_{3}$ of controlling emissions of the particular VOC by itself, compared to controlling emissions from all VOC sources equally. Thus, they are more meaningful in terms of control strategy assessment than absolute reactivities, which can vary greatly depending on the episode and local meteorology.

In addition to depending on the VOC and the scenario, the incremental and relative reactivities depend on how the amounts of VOC added and amounts of ozone formed are quantified. In this work, the amount of added VOC is quantified on a mass basis, since this is how VOCs are regulated, and generally approximates how VOC substitutions are made in practice. Note that relative reactivities will be different if they are quantified on a molar basis, with VOCs with higher molecular weight having higher reactivities on a mole basis than a gram basis.

[^0]Relative reactivities can also depend significantly on how ozone impacts are quantified (Carter, 1994a). Two different ozone quantification methods are used in this work, as follows:
"Ozone Yield" incremental reactivities measure the effect of the VOC on the total amount of ozone formed in the scenario at the time of its maximum concentration. Incremental reactivities are quantified as grams $\mathrm{O}_{3}$ formed per gram VOC added. Most previous recent studies of incremental reactivity (Dodge, 1984; Carter and Atkinson, 1987, 1989, Chang and Rudy, 1990; Jeffries and Crouse, 1991) have been based on this quantification method. The MIR, MOIR, and EBIR scales of Carter (1994a) also use this quantification.
"Max 8 Hour Average" incremental measure the effect of the VOC on the average ozone concentration during the 8 -hour period when the average ozone concentration was the greatest, which in these one-day scenarios was the last 8 hours of the simulation. This provides a measure of ozone impact that is more closely related to the new Federal ozone standard that is given in terms of an 8 hour average. This quantification is used for relative reactivities in this work.

In previous reports, we have reported reactivities in terms of integrated $\mathrm{O}_{3}$ over a standard concentration of 0.09 or 0.12 ppm . This provides a measure of the effect of the VOC on exposure to unacceptable levels of ozone. This is replaced by the Max 8 Hour Average reactivities because it is more representative of the new Federal ozone standard and because reactivities relative to integrated $\mathrm{O}_{3}$ over a standard tend to be between those relative to ozone yield and those relative to 8 -hour averages. Therefore, presenting both ozone yield and maximum 8 -hour average relative reactivities should be sufficient to provide information on how relative reactivities vary with ozone quantification method. Incremental reactivities are quantified as $\mathrm{ppm} \mathrm{O}_{3}$ per milligram VOC emitted per square meter.

## Results

Table 9 lists the ozone yield incremental reactivities for DMSO and the mixture of emitted reactive organic compounds (the base ROG), and gives the ozone yield and maximum 8 -hour average reactivities relative to the base ROG for DMSO, ethane, and m-xylene. Ethane and m-xylene are chosen for comparison because as discussed in the Introduction ethane has been used by the EPA as the informal standard to determine "negligible" reactivity for VOC exemption purposes (Dimitriades, 1999), and m -xylene is an example of a compound considered to be highly reactive. It can be seen that DMSO is calculated to be highly reactive towards ozone formation, being about twice as reactive as the mixture of emitted VOCs in most scenarios, and of comparable reactivity to m -xylene. The relative reactivity of DMSO appears to be somewhat higher with respect to the maximum 8-hour average than with respect to peak ozone yields, though not as much so as is the case for m-xylene. In general, the relative reactivities of DMSO do not appear to be highly dependent on $\mathrm{NO}_{\mathrm{x}}$ and other scenarios conditions, with the standard deviation with respect to the average for the base case scenarios being only $20 \%$ and $13 \%$ for ozone yield and maximum 8-hour average relative reactivities, respectively.

Table 9. Summary of calculated incremental and relative reactivities (gram basis) for DMSO, the mixture of emitted reactive organic compounds (base ROG), ethane, and m-xylene.

|  |  | Increm | ntal |  | eactivities | lative to | base R | (mass b |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Scenario | $\begin{array}{r} \text { Reacti } \\ \left(\mathrm{gm} \mathrm{O}_{3} / \mathrm{g}\right. \end{array}$ | $\begin{aligned} & \text { vities } \\ & \text { m VOC) } \end{aligned}$ |  | Ozone Yield |  |  | ax 8 Hour A |  |
|  |  | Base ROG | DMSO | Ethane | m-Xylene | DMSO | Ethane | m-Xylene | DMSO |
| Adj'd | MIR | 3.79 | 7.15 | 0.08 | 2.87 | 1.89 | 0.07 | 3.07 | 2.26 |
| NOx | MOIR | 1.46 | 2.43 | 0.14 | 2.18 | 1.67 | 0.09 | 2.92 | 2.32 |
|  | EBIR | 0.83 | 1.49 | 0.17 | 1.80 | 1.80 | 0.10 | 2.85 | 2.60 |
| Base | Average | 1.03 | 1.88 | 0.16 | 1.96 | 1.88 | 0.10 | 2.88 | 2.59 |
| Case | St.Dev | 0.42 | 0.77 | 0.04 | 0.51 | 0.37 | 0.02 | 0.27 | 0.34 |
|  | ATL GA | 0.82 | 1.59 | 0.16 | 1.98 | 1.94 | 0.09 | 2.90 | 2.74 |
|  | AUS TX | 0.63 | 1.36 | 0.19 | 1.55 | 2.17 | 0.11 | 2.58 | 3.21 |
|  | BAL MD | 1.59 | 2.62 | 0.12 | 2.25 | 1.65 | 0.08 | 2.97 | 2.29 |
|  | BAT LA | 0.85 | 1.79 | 0.13 | 2.36 | 2.12 | 0.08 | 3.17 | 2.73 |
|  | BIR AL | 0.72 | 1.16 | 0.22 | 1.16 | 1.62 | 0.12 | 2.52 | 2.52 |
|  | BOS MA | 0.72 | 1.28 | 0.20 | 1.49 | 1.76 | 0.13 | 2.35 | 2.42 |
|  | CHA NC | 0.53 | 1.20 | 0.21 | 1.39 | 2.26 | 0.14 | 2.26 | 3.20 |
|  | CHI IL | 0.26 | 0.80 | 0.28 | 0.53 | 3.06 | 0.13 | 2.71 | 3.64 |
|  | CIN OH | 1.12 | 1.65 | 0.18 | 1.65 | 1.47 | 0.10 | 2.66 | 2.31 |
|  | CLE OH | 1.17 | 2.21 | 0.13 | 2.16 | 1.89 | 0.08 | 2.91 | 2.47 |
|  | DAL TX | 2.14 | 3.81 | 0.11 | 2.59 | 1.78 | 0.08 | 2.99 | 2.27 |
|  | DEN CO | 1.66 | 3.34 | 0.09 | 2.66 | 2.01 | 0.06 | 3.17 | 2.54 |
|  | DET MI | 0.98 | 1.50 | 0.18 | 1.65 | 1.53 | 0.10 | 2.75 | 2.32 |
|  | ELP TX | 1.45 | 2.94 | 0.10 | 2.64 | 2.02 | 0.07 | 3.18 | 2.62 |
|  | HAR CT | 0.77 | 1.27 | 0.20 | 1.50 | 1.63 | 0.12 | 2.61 | 2.53 |
|  | HOU TX | 1.10 | 1.79 | 0.16 | 2.01 | 1.64 | 0.09 | 2.92 | 2.36 |
|  | IND IN | 1.24 | 2.13 | 0.14 | 2.21 | 1.71 | 0.09 | 3.11 | 2.41 |
|  | JAC FL | 0.67 | 1.43 | 0.16 | 2.16 | 2.12 | 0.09 | 3.16 | 2.84 |
|  | KAN MO | 1.07 | 1.47 | 0.19 | 1.59 | 1.37 | 0.11 | 2.64 | 2.28 |
|  | LAK LA | 0.42 | 0.95 | 0.22 | 1.86 | 2.26 | 0.11 | 3.30 | 3.04 |
|  | LOS CA | 0.76 | 1.67 | 0.11 | 2.60 | 2.21 | 0.07 | 3.33 | 2.59 |
|  | LOU KY | 1.24 | 1.92 | 0.17 | 1.89 | 1.55 | 0.11 | 2.71 | 2.37 |
|  | MEM TN | 0.76 | 1.23 | 0.20 | 1.63 | 1.61 | 0.11 | 2.79 | 2.43 |
|  | MIA FL | 0.49 | 1.24 | 0.20 | 1.79 | 2.54 | 0.11 | 2.85 | 3.45 |
|  | NAS TN | 0.67 | 1.15 | 0.23 | 1.52 | 1.71 | 0.15 | 2.52 | 2.64 |
|  | NEW NY | 0.39 | 1.04 | 0.17 | 1.33 | 2.66 | 0.08 | 2.88 | 3.08 |
|  | PHI PA | 1.08 | 1.79 | 0.16 | 1.97 | 1.67 | 0.09 | 2.85 | 2.40 |
|  | PHO AZ | 1.46 | 2.47 | 0.12 | 2.34 | 1.69 | 0.08 | 3.15 | 2.32 |
|  | POR OR | 0.96 | 1.65 | 0.17 | 1.90 | 1.72 | 0.11 | 2.80 | 2.60 |
|  | RIC VA | 1.06 | 1.55 | 0.17 | 1.69 | 1.46 | 0.09 | 2.74 | 2.39 |
|  | SAC CA | 1.22 | 1.94 | 0.15 | 2.12 | 1.60 | 0.09 | 3.11 | 2.28 |
|  | SAI MO | 1.38 | 2.61 | 0.11 | 2.34 | 1.89 | 0.07 | 2.98 | 2.47 |
|  | SAL UT | 0.90 | 1.66 | 0.17 | 1.69 | 1.85 | 0.10 | 2.81 | 2.60 |
|  | SAN TX | 1.62 | 2.67 | 0.13 | 2.26 | 1.65 | 0.09 | 2.80 | 2.26 |
|  | SDO CA | 0.85 | 1.95 | 0.11 | 2.56 | 2.31 | 0.08 | 3.22 | 2.75 |
|  | SFO CA | 1.87 | 4.19 | 0.05 | 3.30 | 2.25 | 0.04 | 3.38 | 2.58 |
|  | TAM FL | 1.52 | 2.78 | 0.12 | 2.40 | 1.82 | 0.08 | 3.02 | 2.43 |
|  | TUL OK | 1.17 | 1.77 | 0.17 | 1.86 | 1.51 | 0.10 | 2.76 | 2.28 |
|  | WAS DC | 0.99 | 1.57 | 0.18 | 1.70 | 1.59 | 0.10 | 2.71 | 2.37 |

## CONCLUSIONS

This study has achieved its objective in providing information concerning the relative tendency of DMSO to promote ozone formation in the atmosphere. Prior to this study, although it was known that DMSO reacted relatively rapidly in the atmosphere, it was uncertain whether these reactions were such that they promoted ozone formation in the atmosphere. In particular, the possibility existed that DMSO's reactions may involve sufficient termination processes that DMSO might be an ozone inhibitor and thus inappropriate to regulate as an ozone precursor. However, the environmental chamber experiments carried out for this program showed conclusively that DMSO in fact has significant positive effects on ozone formation under all experimental conditions studied, and its reactions tend to involve more radical initiation than termination. The DMSO photooxidation mechanism that best fit the chamber data predicted that on a mass basis DMSO emissions cause about twice as much ozone formation as the mixture of VOCs emitted from all sources. Therefore, it is clearly inappropriate to exempt DMSO from regulation as a VOC ozone precursor.

In contrast to what is observed with a number of other VOCs (e.g., Carter and Atkinson, 1989; Carter, 1994a), the predicted ozone impacts of DMSO relative to the mixture of VOCs emitted from other sources was found not to be highly scenario dependent. This suggests that uncertainties in scenario conditions may not be a large factor affecting predictions of DMSO's relative atmospheric impacts.

However, there continues to be a number of uncertainties concerning the details of the atmospheric reactions of DMSO and related sulfur-containing compounds. The kinetic and mechanistic data available in the literature are not sufficient to derive an unambiguous mechanism for DMSO. Uncertainties concern the initial reaction of OH radicals with DMSO, and fates of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \cdot \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \cdot$, and $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{O}$. radicals if they are formed, and the decomposition rate of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OONO}_{2}$ if it is formed. The mechanism giving the best fit to the chamber data in this study was found to be the one that assumed that $\sim 75 \%$ of the reaction of OH with DMSO involves abstraction to form $\mathrm{CH}_{2} \mathrm{~S}(\mathrm{O}) \mathrm{CH}_{2}$, which then reacts through a number of steps to ultimately form two formaldehydes and $\mathrm{SO}_{2}$ after two NO to $\mathrm{NO}_{2}$ conversions, with the intermediate reactions involving rapid reaction of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})$. with $\mathrm{O}_{2}$, and rapid decomposition of $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{O} \cdot$ to $\mathrm{CH}_{3} \cdot+\mathrm{SO}_{2}$. (The other $25 \%$ was assumed to involve formation of $\mathrm{DMSO}_{2}$, based on $\mathrm{DMSO}_{2}$ yields obtained in a number of laboratories.) However, this mechanism is not consistent with all the laboratory data, where a more complex mixture of products, including $\mathrm{CH}_{2} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OH}(\mathrm{MSA}), \mathrm{CH}_{3} \mathrm{~S}(\mathrm{O})_{2} \mathrm{OONO}_{2}(\mathrm{MSPN})$, and $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OH}$ (MSIA) are observed or inferred to be formed in significant or non-negligible yields (Barnes et al, 1989; Becker and Patroescu, 1996, Hynes and Wine, 1996, Sørensen et al, 1996, Urbanski et al, 1998) . Indeed, recent data from Wine's laboratory (Hynes and Wine, 1996, Urbanski et al, 1998) suggest that the major reaction involves initial formation of MSIA, whose subsequent reactions can account for the other products observed in the previous laboratory studies. However this mechanism (Model "A") is not consistent with the environmental chamber data obtained in this study.

Although comprehensive product analyses were not carried out in this project, some information was obtained concerning the compounds formed when DMSO reacts under atmospheric conditions. Very high formaldehyde yields were observed, suggesting that much of the carbon in DMSO is converted into formaldehyde when it reacts in the atmosphere. Although the none of the sulfur-containing products were directly monitored, data from a total gas-phase sulfur analyzer indicated that approximately half of the sulfur in the reacting DMSO is converted into a form that this analyzer does not detect; i.e., something other than $\mathrm{SO}_{2}$. The mechanism that best fit the chamber data suggests that this product is primarily $\mathrm{DMSO}_{2}$, which is predicted to be formed in about $25 \%$ yield in the OH reaction, and in high yield in the reaction of DMSO with $\mathrm{NO}_{3}$, which is predicted to occur up to $50 \%$ of the time in the conditions of our chamber experiments. However, it is unknown whether $\mathrm{DMSO}_{2}$ is in fact not detected by this total gas phase sulfur analyzer, or whether formation of other products, such as MSA, could be contributing to this loss of gas phase sulfur.

Finally, it was found that even the most reactive of the alternative mechanisms that were considered tended to underpredict the rates of NO oxidation and $\mathrm{O}_{3}$ formation during the initial periods of some of the experiments. This suggests that there may be some radical initiation process involved in the oxidation of DMSO or (more likely) its reactive products that is not being considered. This also suggests that our atmospheric reactivity calculations using this mechanism may in fact be underestimating the actual ozone impact of DMSO. However, that mechanism was found to give reasonably good simulations of DMSO's ozone impacts in the experiments most representative of atmospheric conditions, suggesting that the biases in that mechanisms predictions of DMSO's atmospheric ozone impacts are not likely to be large.

## REFERENCES

Atkinson, R. (1989): "Kinetics and Mechanisms of the Gas-Phase Reactions of the Hydroxyl Radical with Organic Compounds," J. Phys. Chem. Ref. Data, Monograph no 1.

Atkinson, R., D. L. Baulch, R. A. Cox, R. F. Hampson, Jr., J. A. Kerr, M. J. Rossi, and J. Troe (1997): "Evaluated Kinetic, Photochemical and Heterogeneous Data for Atmospheric Chemistry: Supplement V and VI, IUPAC Subcommittee on Gas Kinetic Data Evaluation for Atmospheric Chemistry," Phys. Chem. Ref. Data, 26, 521-1011 (Supplement V) and 1329-1499 (Supplement VI).

Barnes, I., V. Bastian, K. H. Becker, and D. Martin (1989), "Fourier Transform IR Studies of the Reaction of Dimethyl Sulfoxide with $\mathrm{OH}, \mathrm{NO}_{3}$, and Cl Radicals," in "Biogenic Sulfur in the Environment," E. S. Saltzman and W. J. Cooper, Editors, ACS Symposium Series 393, American Chemical Society, Washington, DC.

Baugues, K. (1990): "Preliminary Planning Information for Updating the Ozone Regulatory Impact Analysis Version of EKMA," Draft Document, Source Receptor Analysis Branch, Technical Support Division, U. S. Environmental Protection Agency, Research Triangle Park, NC, January.

Becker, K. H. and J. V. Patroescu (1996): "Reaktionen von Organischen Schwefelverbindun-gen in der Atmophäre," Bergische Universität Gesamthochschule Wuppertal, Germany, March.

CARB (1999) California Air Resources Board, Proposed Regulation for Title 17, California Code of Regulations, Division 3, Chapter 1, Subchapter 8.5, Article 3.1, sections 94560-94539.

Carter, W. P. L. (1990): "A Detailed Mechanism for the Gas-Phase Atmospheric Reactions of Organic Compounds," Atmos. Environ., 24A, 481-518.

Carter, W. P. L. (1994a): "Development of Ozone Reactivity Scales for Volatile Organic Compounds," J. Air \& Waste Manage. Assoc., 44, 881-899.

Carter, W. P. L. (1994b): "Calculation of Reactivity Scales Using an Updated Carbon Bond IV Mechanism," Report Prepared for Systems Applications International Under Funding from the Auto/Oil Air Quality Improvement Research Program, April 12. Available at http://helium.ucr.edu/~carter/absts.htm\#cb4rct.

Carter, W. P. L. (1997): "Estimated Atmospheric Reactivity Ozone Formation Potentials of Dimetlyl Sulfide and Dimethyl Sulfoxide," Letter to Mr. Louis Zeillmann or Gaylord Chemical Corporation, December 31.

Carter, W. P. L. (2000): "Documentation of the SAPRC-99 Chemical Mechanism for VOC Reactivity Assessment," Report to the California Air Resources Board, Contracts 92-329 and 95-308, May 8. Available at http://helium.ucr.edu/~carter/absts.htm\#saprc99.

Carter, W. P. L., R. Atkinson, A. M. Winer, and J. N. Pitts, Jr. (1982): "Experimental Investigation of Chamber-Dependent Radical Sources," Int. J. Chem. Kinet., 14, 1071.

Carter, W. P. L. and R. Atkinson (1987): "An Experimental Study of Incremental Hydrocarbon Reactivity," Environ. Sci. Technol., 21, 670-679

Carter, W. P. L. and R. Atkinson (1989): "A Computer Modeling Study of Incremental Hydrocarbon Reactivity", Environ. Sci. Technol., 23, 864.

Carter, W. P. L., and F. W. Lurmann (1990): "Evaluation of the RADM Gas-Phase Chemical Mechanism," Final Report, EPA-600/3-90-001.

Carter, W. P. L. and F. W. Lurmann (1991): "Evaluation of a Detailed Gas-Phase Atmospheric Reaction Mechanism using Environmental Chamber Data," Atm. Environ. 25A, 2771-2806.

Carter, W. P. L., J. A. Pierce, I. L. Malkina, D. Luo and W. D. Long (1993): "Environmental Chamber Studies of Maximum Incremental Reactivities of Volatile Organic Compounds," Report to Coordinating Research Council, Project No. ME-9, California Air Resources Board Contract No. A032-0692; South Coast Air Quality Management District Contract No. C91323, United States Environmental Protection Agency Cooperative Agreement No. CR-814396-01-0, University Corporation for Atmospheric Research Contract No. 59166, and Dow Corning Corporation. April 1. Available at http://helium.ucr.edu/~carter/absts.htm\#rct1rept.

Carter, W. P. L., J. A. Pierce, D. Luo, and I. L. Malkina (1995a): "Environmental Chamber Studies of Maximum Incremental Reactivities of Volatile Organic Compounds," Atmos. Environ. 29, 24992511.

Carter, W. P. L., D. Luo, I. L. Malkina, and J. A. Pierce (1995b): "Environmental Chamber Studies of Atmospheric Reactivities of Volatile Organic Compounds. Effects of Varying ROG Surrogate and $\mathrm{NO}_{\mathrm{x}}$," Final report to Coordinating Research Council, Inc., Project ME-9, California Air Resources Board, Contract A032-0692, and South Coast Air Quality Management District, Contract C91323. March 24. Available at http://helium.ucr.edu/~carter/absts.htm\#rct2rept.

Carter, W. P. L., D. Luo, I. L. Malkina, and D. Fitz (1995c): "The University of California, Riverside Environmental Chamber Data Base for Evaluating Oxidant Mechanism. Indoor Chamber Experiments through 1993," Report submitted to the U. S. Environmental Protection Agency, EPA/AREAL, Research Triangle Park, NC., March 20. Available at http://helium.ucr.edu/ ~carter/absts.htm\#databas.

Carter, W. P. L., D. Luo, I. L. Malkina, and J. A. Pierce (1995d): "Environmental Chamber Studies of Atmospheric Reactivities of Volatile Organic Compounds. Effects of Varying Chamber and Light Source," Final report to National Renewable Energy Laboratory, Contract XZ-2-12075, Coordinating Research Council, Inc., Project M-9, California Air Resources Board, Contract A032-0692, and South Coast Air Quality Management District, Contract C91323, March 26. Available at http://helium.ucr.edu/~carter/absts.htm\#explrept.

Carter, W. P. L., D. Luo, and I. L. Malkina (1997): "Environmental Chamber Studies for Development of an Updated Photochemical Mechanism for VOC Reactivity Assessment," Final report to the California Air Resources Board, the Coordinating Research Council, and the National Renewable Energy Laboratory, November 26. Available at http://helium.ucr.edu/~carter/absts.htm\#rct3rept.

Carter. W. P. L., D. Luo and I. L. Malkina (2000): "Investigation of Atmospheric Reactivities of Selected Consumer Product VOCs," Report to California Air Resources Board, May 30. Available at http://helium.ucr.edu/~carter/absts.htm\#cpreport.

Chang, T. Y. and S. J. Rudy (1990): "Ozone-Forming Potential of Organic Emissions from AlternativeFueled Vehicles," Atmos. Environ., 24A, 2421-2430.

Croes, B. E., Technical Support Division, California Air Resources Board, personal communication (1991).

Croes, B. E., et al. (1994): "Southern California Air Quality Study Data Archive," Research Division, California Air Resources Board.

Dasgupta, P. K, Dong, S. and Hwang, H. (1988): "Continuous Liquid Phase Fluorometry Coupled to a Diffusion Scrubber for the Determination of Atmospheric Formaldehyde, Hydrogen Peroxide, and Sulfur Dioxide," Atmos. Environ. 22, 949-963.

Dasgupta, P. K, Dong, S. and Hwang, H. (1990): Aerosol Science and Technology 12, 98-104
Dimitriades, B. (1999): "Scientific Basis of an Improved EPA Policy on Control of Organic Emissions for Ambient Ozone Reduction," J. Air \& Waste Manage. Assoc. 49, 831-838

Dodge, M. C. (1984): "Combined effects of organic reactivity and NMHC/NOx ratio on photochemical oxidant formation -- a modeling study," Atmos. Environ., 18, 1657.

EPA (1984): "Guideline for Using the Carbon Bond Mechanism in City-Specific EKMA," EPA-450/4-84-005, February.

Falbe-Hansen, H., S. Sørensen, N. R. Jensen, T. Pedersen, and J. Hjorth (2000): "Atmospheric Gas-Phase Reactions of Dimethylsulfoxide and dimethylsulfone with OH and $\mathrm{NO}_{3}$ radicals, Cl atoms and ozone," Atmos. Environ. 35, 1543-1551.

Gery, M. W., R. D. Edmond and G. Z. Whitten (1987): "Tropospheric Ultraviolet Radiation. Assessment of Existing Data and Effects on Ozone Formation," Final Report, EPA-600/3-87-047, October.

Gipson, G. L., W. P. Freas, R. A. Kelly and E. L. Meyer (1981): "Guideline for Use of City-Specific EKMA in Preparing Ozone SIPs, EPA-450/4-80-027, March.

Gipson, G. L. and W. P. Freas (1983): "Use of City-Specific EKMA in the Ozone RIA," U. S. Environmental Protection Agency, July.

Gipson, G. L. (1984): "Users Manual for OZIPM-2: Ozone Isopleth Plotting Package With Optional Mechanism/Version 2," EPA-450/4-84-024, August.

Hogo, H. and M. W. Gery (1988): "Guidelines for Using OZIPM-4 with CBM-IV or Optional Mechanisms. Volume 1. Description of the Ozone Isopleth Plotting Package Version 4", Final Report for EPA Contract No. 68-02-4136, Atmospheric Sciences Research Laboratory, Research Triangle Park, NC. January.

Hynes, A. J. and P. H. Wine (1996): "The Atmospheric Chemistry of Dimethylsulfoxide (DMSO) Kinetics and Mechanism of the OH + DMSO Reaction," J. Atm. Chem. 24, 23-37.

Jeffries, H. E. (1991): "UNC Solar Radiation Models," unpublished draft report for EPA Cooperative Agreements CR813107, CR813964 and CR815779".

Jeffries, H. E., K. G. Sexton, J. R. Arnold, and T. L. Kale (1989): "Validation Testing of New Mechanisms with Outdoor Chamber Data. Volume 2: Analysis of VOC Data for the CB4 and CAL Photochemical Mechanisms," Final Report, EPA-600/3-89-010b.

Jeffries, H. E. and R. Crouse (1991): "Scientific and Technical Issues Related to the Application of Incremental Reactivity. Part II: Explaining Mechanism Differences," Report prepared for Western States Petroleum Association, Glendale, CA, October.

Johnson, G. M. (1983): "Factors Affecting Oxidant Formation in Sydney Air," in "The Urban Atmosphere -- Sydney, a Case Study." Eds. J. N. Carras and G. M. Johnson (CSIRO, Melbourne), pp. 393-408.

Lurmann, F. W. and H. H. Main (1992): "Analysis of the Ambient VOC Data Collected in the Southern California Air Quality Study," Final Report to California Air Resources Board Contract No. A832-130, February.

McBride, S., M. Oravetz, and A.G. Russell. 1997. "Cost-Benefit and Uncertainty Issues Using Organic Reactivity to Regulate Urban Ozone." Environ. Sci. Technol. 35, A238-44.

RRWG (1999): "VOC Reactivity Policy White Paper," Prepared by the Reactivity Research Work Group Policy Team, October 1. Available at http://www.cgenv.com/Narsto/reactinfo.html.

Sørensen, S., J. Falbe-Hansen, M. Mangoni, J. Hjorth, and N. R. Jensen (1996): "Observation of dMSO and $\mathrm{CH}_{3} \mathrm{~S}(\mathrm{O}) \mathrm{OH}$ from the Gas Phase Reaction Between DMS and OH," J. Atm. Chem, 24, 299315.

Urbanski, S. B, R. E. Stickel, and P. H. Wine (1998): "Mechanistic and Kinetic Study of the Gas-Phase Reaction of Hydroxyl Radical with Dimethyl Sulfoxide," J. Phys. Chem. A, 102, 10522-10529.

Yin, F, D. Grosjean, J. H. Seinfeld (1990): "Photooxidation of Dimethyl Sulfide and Dimethyl Disulfide. I: Mechanism Development," J. Atmos. Chem. 11. 309-364.

Zafonte, L., P. L. Rieger, and J. R. Holmes (1977): "Nitrogen Dioxide Photolysis in the Los Angeles Atmosphere," Environ. Sci. Technol. 11, 483-487.

## APPENDIX A.

## MECHANISM LISTING AND TABULATIONS

This Appendix gives a complete listing of the mechanisms used to represent the reactions of species other than DMSO and its unique products in the model simulations in this report. This includes the "base mechanism" giving the reactions of the inorganic compounds and the common organic products, the mechanisms for the ROG surrogate VOCs used in the chamber experiments, and the mechanisms for the individual VOCs and lumped VOC species used in the atmospheric reactivity simulations. (The mechanisms used for DMSO are discussed in the body of the report.) Table A-1 contains a list of all the model species used in the mechanism, Table A-2 lists all the reactions and rate parameters, and Table A-3 lists the absorption cross sections and quantum yields for the photolysis reactions. In addition, Table A-4 gives the chamber-dependent parameters used in the model simulations of the chamber experiments.

Table A-1. Listing of model species used in the SAPRC-99 mechanism for the base case environmental chamber and atmospheric reactivity simulations.
Type and Name Description

## Species used in Base Mechanism

Constant Species.

| O2 | Oxygen |
| :--- | :--- |
| M | Air |
| H2O | Water |
| H2 | Hydrogen Molecules |
| HV | Light |

Active Inorganic Species.

| O3 | Ozone |
| :--- | :--- |
| NO | Nitric Oxide |
| NO2 | Nitrogen Dioxide |
| NO3 | Nitrate Radical |
| N2O5 | Nitrogen Pentoxide |
| HONO | Nitrous Acid |
| HNO3 | Nitric Acid |
| HNO4 | Peroxynitric Acid |
| HO2H | Hydrogen Peroxide |
| CO | Carbon Monoxide |
| SO2 | Sulfur Dioxide |

Active Radical Species and Operators.

| HO. | Hydroxyl Radicals |
| :--- | :--- |
| HO2. | Hydroperoxide Radicals |
| C-O2. | Methyl Peroxy Radicals |
| RO2-R. | Peroxy Radical Operator representing NO to NO2 conversion with HO2 formation. |
| R2O2. | Peroxy Radical Operator representing NO to NO2 conversion without HO2 formation. |
| RO2-N. | Peroxy Radical Operator representing NO consumption with organic nitrate formation. |
| CCO-O2. | Acetyl Peroxy Radicals |
| RCO-O2. | Peroxy Propionyl and higher peroxy acyl Radicals |
| BZCO-O2. | Peroxyacyl radical formed from Aromatic Aldehydes |
| MA-RCO3. | Peroxyacyl radicals formed from methacrolein and other acroleins. |

Steady State Radical Species

| O3P | Ground State Oxygen Atoms |
| :--- | :--- |
| O*1D2 | Excited Oxygen Atoms |
| TBU-O. | t-Butoxy Radicals |
| BZ-O. | Phenoxy Radicals |
| BZ(NO2)-O. | Nitro-substituted Phenoxy Radical |
| HOCOO. | Radical formed when Formaldehyde reacts with HO2 |
| PAN and PAN Analogues |  |
| PAN | Peroxy Acetyl Nitrate |
| PAN2 | PPN and other higher alkyl PAN analogues |
| PBZN | PAN analogues formed from Aromatic Aldehydes |
| MA-PAN | PAN analogue formed from Methacrolein |

Table A-1 (continued)

| Type and Name | Description |
| :---: | :---: |
| Explicit and Lumped Molecule Reactive Organic Product Species |  |
| HCHO | Formaldehyde |
| CCHO | Acetaldehyde |
| RCHO | Lumped C3+ Aldehydes |
| ACET | Acetone |
| MEK | Ketones and other non-aldehyde oxygenated products which react with OH radicals slower than $5 \times 10^{-12} \mathrm{~cm}^{3} \mathrm{molec}^{-2} \mathrm{sec}^{-1}$. |
| MEOH | Methanol |
| COOH | Methyl Hydroperoxide |
| ROOH | Lumped higher organic hydroperoxides |
| GLY | Glyoxal |
| MGLY | Methyl Glyoxal |
| BACL | Biacetyl |
| PHEN | Phenol |
| CRES | Cresols |
| NPHE | Nitrophenols |
| BALD | Aromatic aldehydes (e.g., benzaldehyde) |
| METHACRO | Methacrolein |
| MVK | Methyl Vinyl Ketone |
| ISO-PROD | Lumped isoprene product species |
| Lumped Parameter Products |  |
| PROD2 | Ketones and other non-aldehyde oxygenated products which react with OH radicals faster than $5 \times 10^{-12} \mathrm{~cm}^{3} \mathrm{molec}^{-2} \mathrm{sec}^{-1}$. |
| RNO3 | Lumped Organic Nitrates |
| Uncharacterized Reactive Aromatic Ring Fragmentation Products |  |
| DCB1 | Reactive Aromatic Fragmentation Products that do not undergo signficant photodecomposition to radicals. |
| DCB2 | Reactive Aromatic Fragmentation Products which photolyze with alpha-dicarbonyl-like action spectrum. |
| DCB3 | Reactive Aromatic Fragmentation Products which photolyze with acrolein action spectrum. |
| Non-Reacting Species |  |
| CO2 | Carbon Dioxide |
| XC | Lost Carbon |
| XN | Lost Nitrogen |
| SULF | Sulfates ( $\mathrm{SO}_{3}$ or $\mathrm{H}_{2} \mathrm{SO}_{4}$ ) |
| Low Reactivity Compounds or Unknown Products Represented as Unreactive |  |
| H2 | Hydrogen |
| HCOOH | Formic Acid |
| CCO-OH | Acetic Acid |
| RCO-OH | Higher organic acids |
| CCO-OOH | Peroxy Acetic Acid |
| RCO-OOH | Higher organic peroxy acids |
| NROG | Unspecified Unreactive Carbon |

Table A-1 (continued)
Type and Name Description

## Base ROG VOC Species used in the Chamber Simulations

| N-C4 | n-Butane |
| :--- | :--- |
| N-C6 | n-Hexane |
| N-C8 | n-Octane |
| ETHENE | Ethene |
| PROPENE | Propene |
| T-2-BUTE | Trans-2-Butene |
| TOLUENE | Toluene |
| M-XYLENE | m-Xylene |

## Explicit and Lumped VOC Species used in the Ambient Simulations

Primary Organics Represented explicitly
CH4 Methane

ETHENE Ethene
ISOPRENE Isoprene
Lumped Parameter Species
ALK1 Alkanes and other non-aromatic compounds that react only with OH , and have $\mathrm{kOH}<5$ $\times 10^{2} \mathrm{ppm}-1 \mathrm{~min}-1$. (Primarily ethane)
ALK2 Alkanes and other non-aromatic compounds that react only with OH , and have kOH between $5 \times 10^{2}$ and $2.5 \times 10^{3} \mathrm{ppm}-1 \mathrm{~min}-1$. (Primarily propane and acetylene)
ALK3 Alkanes and other non-aromatic compounds that react only with OH , and have kOH between $2.5 \times 10^{3}$ and $5 \times 10^{3} \mathrm{ppm}-1 \mathrm{~min}-1$.
ALK4 Alkanes and other non-aromatic compounds that react only with OH , and have kOH between $5 \times 10^{3}$ and $1 \times 10^{4} \mathrm{ppm}-1 \mathrm{~min}-1$.
ALK5 Alkanes and other non-aromatic compounds that react only with OH , and have kOH greater than $1 \times 10^{4} \mathrm{ppm}-1 \mathrm{~min}-1$.
ARO1 Aromatics with $\mathrm{kOH}<2 \times 10^{4} \mathrm{ppm}-1 \mathrm{~min}-1$.
ARO2 Aromatics with $\mathrm{kOH}>2 \times 10^{4} \mathrm{ppm}-1 \mathrm{~min}-1$.
OLE1 Alkenes (other than ethene) with $\mathrm{kOH}<7 \mathrm{x} 10^{4} \mathrm{ppm}-1 \mathrm{~min}-1$.
OLE2 Alkenes with $\mathrm{kOH}>7 \times 10^{4} \mathrm{ppm}-1 \mathrm{~min}-1$.
TERP Terpenes

Table A-2. Listing of the reactions used in the SAPRC-99 mechanism for the base case environmental chamber and atmospheric reactivity simulations. See Carter (2000) for documentation.

| Label | Rate Parameters [a] |  |  |  | Reaction and Products [b] |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{k}(298)$ | A | Ea | B |  |
| Inorganic Reactions |  |  |  |  |  |
| , |  | Phot Set= NO2 |  |  | $\mathrm{NO} 2+\mathrm{HV}=\mathrm{NO}+\mathrm{O} 3 \mathrm{P}$ |
| 2 | 5.79e-34 | $5.68 \mathrm{e}-34$ | 0.00 | -2.8 | $\mathrm{O} 3 \mathrm{P}+\mathrm{O} 2+\mathrm{M}=\mathrm{O} 3+\mathrm{M}$ |
| 3 | 7.96e-15 | 8.00e-12 | 4.09 |  | $\mathrm{O} 3 \mathrm{P}+\mathrm{O} 3=\# 2 \mathrm{O} 2$ |
| 4 | 1.01e-31 | $1.00 \mathrm{e}-31$ | 0.00 | -1.6 | $\mathrm{O} 3 \mathrm{P}+\mathrm{NO}+\mathrm{M}=\mathrm{NO} 2+\mathrm{M}$ |
| 5 | $9.72 \mathrm{e}-12$ | 6.50e-12 | -0.24 |  | $\mathrm{O} 3 \mathrm{P}+\mathrm{NO} 2=\mathrm{NO}+\mathrm{O} 2$ |
| 6 | 1.82e-12 | Falloff, F=0.80 |  |  | $\mathrm{O} 3 \mathrm{P}+\mathrm{NO} 2=\mathrm{NO} 3+\mathrm{M}$ |
|  | 0 : | 9.00e-32 | 0.00 | $-2.0$ |  |
|  | inf: | $2.20 \mathrm{e}-11$ | 0.00 | 0.0 |  |
| 8 | 1.81e-14 | 1.80e-12 | 2.72 |  | $\mathrm{O} 3+\mathrm{NO}=\mathrm{NO} 2+\mathrm{O} 2$ |
| 9 | $3.52 \mathrm{e}-17$ | $1.40 \mathrm{e}-13$ | 4.91 |  | $\mathrm{O} 3+\mathrm{NO} 2=\mathrm{O} 2+\mathrm{NO} 3$ |
| 10 | $2.60 \mathrm{e}-11$ | $1.80 \mathrm{e}-11$ | -0.22 |  | $\mathrm{NO}+\mathrm{NO} 3=\# 2 \mathrm{NO} 2$ |
| 11 | 1.95e-38 | 3.30e-39 | -1.05 |  | $\mathrm{NO}+\mathrm{NO}+\mathrm{O} 2=\# 2 \mathrm{NO} 2$ |
| 12 | $1.54 \mathrm{e}-12$ | Falloff, $\mathrm{F}=0.45$ |  |  | $\mathrm{NO} 2+\mathrm{NO} 3=\mathrm{N} 2 \mathrm{O} 5$ |
|  |  | $2.80 \mathrm{e}-30$ | 0.00 | -3.5 |  |
|  | inf: | 2.00e-12 | 0.00 | 0.2 |  |
| 13 | 5.28e-2 | Falloff, F=0.45 |  |  | $\mathrm{N} 2 \mathrm{O} 5=\mathrm{NO} 2+\mathrm{NO} 3$ |
|  |  | $1.00 \mathrm{e}-3$ | 21.86 | -3.5 |  |
|  | inf: | $9.70 \mathrm{e}+14$ | 22.02 | 0.1 |  |
| 14 | 2.60e-22 | 2.60e-22 |  |  | $\mathrm{N} 2 \mathrm{O} 5+\mathrm{H} 2 \mathrm{O}=$ \#2 HNO3 |
| 15 |  | (Slow) |  |  | $\mathrm{N} 2 \mathrm{O} 5+\mathrm{HV}=\mathrm{NO} 3+\mathrm{NO}+\mathrm{O} 3 \mathrm{P}$ |
| 16 |  | (Slow) |  |  | $\mathrm{N} 2 \mathrm{O} 5+\mathrm{HV}=\mathrm{NO} 3+\mathrm{NO} 2$ |
| 17 | 6.56e-16 | $4.50 \mathrm{e}-14$ | 2.50 |  | $\mathrm{NO} 2+\mathrm{NO} 3=\mathrm{NO}+\mathrm{NO} 2+\mathrm{O} 2$ |
| 18 |  | Phot Set= N | O3NO |  | $\mathrm{NO} 3+\mathrm{HV}=\mathrm{NO}+\mathrm{O} 2$ |
| 19 |  | Phot Set= N03 | 3 NO 2 |  | $\mathrm{NO} 3+\mathrm{HV}=\mathrm{NO} 2+\mathrm{O} 3 \mathrm{P}$ |
| 20 |  | Phot Set= | 303P |  | $\mathrm{O} 3+\mathrm{HV}=\mathrm{O} 3 \mathrm{P}+\mathrm{O}_{2}$ |
| 21 |  | Phot $\mathrm{Set}=0$ | 301D |  | $\mathrm{O} 3+\mathrm{HV}=\mathrm{O} * 1 \mathrm{D} 2+\mathrm{O} 2$ |
| 22 | 2.20e-10 | 2.20e-10 |  |  | $\mathrm{O} * 1 \mathrm{D} 2+\mathrm{H} 2 \mathrm{O}=\# 2 \mathrm{HO}$. |
| 23 | 2.87e-11 | 2.09e-11 | -0.19 |  | $\mathrm{O} * 1 \mathrm{D} 2+\mathrm{M}=\mathrm{O} 3 \mathrm{P}+\mathrm{M}$ |
| 24 | 7.41e-12 | Falloff, F=0.60 |  |  | HO. $+\mathrm{NO}=\mathrm{HONO}$ |
|  |  | 7.00e-31 | 0.00 | -2.6 |  |
|  | inf: | $3.60 \mathrm{e}-11$ | 0.00 | -0.1 |  |
| 25 |  | hot Set= HO | NO-NO |  | $\mathrm{HONO}+\mathrm{HV}=\mathrm{HO} .+\mathrm{NO}$ |
| 26 |  | hot $\mathrm{Set}=\mathrm{HO}$ | NO-NO2 |  | $\mathrm{HONO}+\mathrm{HV}=\mathrm{HO} 2 .+\mathrm{NO} 2$ |
| 27 | 6.46e-12 | 2.70e-12 | -0.52 |  | $\mathrm{HO} .+\mathrm{HONO}=\mathrm{H} 2 \mathrm{O}+\mathrm{NO} 2$ |
| 28 | 8.98e-12 | Falloff, $\mathrm{F}=0.60$ |  |  | HO. $+\mathrm{NO} 2=\mathrm{HNO} 3$ |
|  | 0: | 2.43e-30 | 0.00 | -3.1 |  |
|  | inf: | 1.67e-11 | 0.00 | -2.1 |  |
| 29 | 2.00e-11 | $2.00 \mathrm{e}-11$ |  |  | HO. $+\mathrm{NO} 3=\mathrm{HO} 2 .+\mathrm{NO} 2$ |
| 30 | 1.47e-13 | $\mathrm{k}=\mathrm{k} 0+\mathrm{k} 3 \mathrm{M} /(1+\mathrm{k} 3 \mathrm{M} / \mathrm{k} 2)$ |  |  | $\mathrm{HO} .+\mathrm{HNO} 3=\mathrm{H} 2 \mathrm{O}+\mathrm{NO} 3$ |
|  | k0: | 7.20e-15 | -1.56 | 0.0 |  |
|  | k2. | 4.10e-16 | -2.86 | 0.0 |  |
|  |  | $1.90 \mathrm{e}-33$ | -1.44 | 0.0 |  |
| 31 |  | Phot $\mathrm{Set}=\mathrm{HNO} 3$ |  |  | $\mathrm{HNO} 3+\mathrm{HV}=\mathrm{HO} .+\mathrm{NO} 2$ |
| 32 | 2.09e-13 | k=k | + k2 [M |  | HO. $+\mathrm{CO}=\mathrm{HO} 2 .+\mathrm{CO} 2$ |
|  | k1: | 1.30e-13 | 0.00 | 0.0 |  |

Table A-2 (continued)

| Label | Rate Parameters [a] |  |  |  | Reaction and Products [b] |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | k(298) | A | Ea | B |  |
|  |  | : 3.19e-33 | 0.00 | 0.0 |  |
| 33 | $6.63 \mathrm{e}-14$ | 1.90e-12 | 1.99 |  | HO. $+\mathrm{O} 3=\mathrm{HO} 2 .+\mathrm{O} 2$ |
| 34 | $8.41 \mathrm{e}-12$ | 3.40e-12 | -0.54 |  | HO2. $+\mathrm{NO}=\mathrm{HO} .+\mathrm{NO} 2$ |
| 35 | 1.38e-12 | Falloff, F=0.60 |  |  | HO2. $+\mathrm{NO} 2=\mathrm{HNO} 4$ |
|  |  | : 1.80e-31 | 0.00 | -3.2 |  |
|  | inf: | : $4.70 \mathrm{e}-12$ | 0.00 | 0.0 |  |
| 36 | $7.55 \mathrm{e}-2$ | Falloff, F=0.50 |  |  | HNO4 $=\mathrm{HO} 2 .+\mathrm{NO} 2$ |
|  |  | : 4.10e-5 | 21.16 | 0.0 |  |
|  |  | : $5.70 \mathrm{e}+15$ | 22.20 | 0.0 |  |
| 37 |  | Phot Set= H | 2NO2 |  | HNO4 + HV = \#. 61 \{HO2. + NO2 $\}+$ \#. $39\{\mathrm{HO} .+\mathrm{NO} 3\}$ |
| 38 | 5.02e-12 | 1.50e-12 | -0.72 |  | $\mathrm{HNO} 4+\mathrm{HO}=\mathrm{H} 2 \mathrm{O}+\mathrm{NO}_{2}+\mathrm{O} 2$ |
| 39 | 1.87e-15 | 1.40e-14 | 1.19 |  | HO2. + O3 = HO. + \#2 O2 |
| 40A | 2.87e-12 | $\mathrm{k}=\mathrm{k} 1+\mathrm{k} 2[\mathrm{M}]$ |  |  | HO2. $+\mathrm{HO} 2 .=\mathrm{HO} 2 \mathrm{H}+\mathrm{O} 2$ |
|  |  | : $2.20 \mathrm{e}-13$ | -1.19 | 0.0 |  |
|  | k2: | : 1.85e-33 | -1.95 | 0.0 |  |
| 40B | 6.46e-30 | $\mathrm{k}=\mathrm{k} 1+\mathrm{k} 2[\mathrm{M}]$ |  |  | HO2. $+\mathrm{HO} 2 .+\mathrm{H} 2 \mathrm{O}=\mathrm{HO} 2 \mathrm{H}+\mathrm{O} 2+\mathrm{H} 2 \mathrm{O}$ |
|  |  | : 3.08e-34 | -5.56 | 0.0 |  |
|  |  | : $2.59 \mathrm{e}-54$ | -6.32 | 0.0 |  |
| 41 | $4.00 \mathrm{e}-12$ | 4.00e-12 |  |  | $\mathrm{NO} 3+\mathrm{HO} 2 .=\# .8\{\mathrm{HO} .+\mathrm{NO} 2+\mathrm{O} 2\}+\# .2\{\mathrm{HNO} 3+\mathrm{O} 2\}$ |
| 42 | 2.28e-16 | 8.50e-13 | 4.87 |  | $\mathrm{NO} 3+\mathrm{NO} 3=\# 2 \mathrm{NO} 2+\mathrm{O} 2$ |
| 43 |  | Phot Set= | H2O2 |  | $\mathrm{HO} 2 \mathrm{H}+\mathrm{HV}=\# 2 \mathrm{HO}$. |
| 44 | 1.70e-12 | 2.90e-12 | 0.32 |  | $\mathrm{HO} 2 \mathrm{H}+\mathrm{HO}=\mathrm{HO} 2 .+\mathrm{H} 2 \mathrm{O}$ |
| 45 | 1.11e-10 | 4.80e-11 | -0.50 |  | HO. $+\mathrm{HO} 2 .=\mathrm{H} 2 \mathrm{O}+\mathrm{O} 2$ |
| S2OH | 9.77e-13 | Falloff, F=0.45 |  |  | HO. + SO2 $=$ HO2. + SULF |
|  |  | : 4.00e-31 | 0.00 | -3.3 |  |
|  |  | : $2.00 \mathrm{e}-12$ | 0.00 | 0.0 |  |
| H 2 OH | 6.70e-15 | 7.70e-12 | 4.17 |  | HO. $+\mathrm{H} 2=\mathrm{HO} 2 .+\mathrm{H} 2 \mathrm{O}$ |
| Methyl peroxy and methoxy reactions |  |  |  |  |  |
| MER1 | $7.29 \mathrm{e}-12$ | $2.80 \mathrm{e}-12$ | -0.57 |  | $\mathrm{C}-\mathrm{O} 2 .+\mathrm{NO}=\mathrm{NO} 2+\mathrm{HCHO}+\mathrm{HO} 2$. |
| MER4 | 5.21e-12 | 3.80e-13 | -1.55 |  | $\mathrm{C}-\mathrm{O} 2 .+\mathrm{HO} 2 .=\mathrm{COOH}+\mathrm{O} 2$ |
| MEN3 | 1.30e-12 | 1.30e-12 |  |  | $\mathrm{C}-\mathrm{O} 2 .+\mathrm{NO} 3=\mathrm{HCHO}+\mathrm{HO} 2 .+\mathrm{NO} 2$ |
| MER5 | 2.65e-13 | 2.45e-14 | -1.41 |  | $\mathrm{C}-\mathrm{O} 2 .+\mathrm{C}-\mathrm{O} 2 .=\mathrm{MEOH}+\mathrm{HCHO}+\mathrm{O} 2$ |
| MER6 | 1.07e-13 | 5.90e-13 | 1.01 |  | $\mathrm{C}-\mathrm{O} 2 .+\mathrm{C}-\mathrm{O} 2 .=$ \#2 $\{\mathrm{HCHO}+\mathrm{HO} 2$. |
| Peroxy Racical Operators |  |  |  |  |  |
| RRNO | 9.04e-12 | 2.70e-12 | -0.72 |  | $\mathrm{RO} 2-\mathrm{R} .+\mathrm{NO}=\mathrm{NO} 2+\mathrm{HO} 2$. |
| RRH2 | $1.49 \mathrm{e}-11$ | $1.90 \mathrm{e}-13$ | -2.58 |  | $\mathrm{RO} 2-\mathrm{R} .+\mathrm{HO} 2 . \mathrm{ROOH}+\mathrm{O} 2+$ \#-3 XC |
| RRN3 | 2.30e-12 | 2.30e-12 |  |  | $\mathrm{RO} 2-\mathrm{R} .+\mathrm{NO} 3=\mathrm{NO} 2+\mathrm{O} 2+\mathrm{HO} 2$. |
| RRME | 2.00e-13 | 2.00e-13 |  |  | RO2-R. + C-O2. $=$ HO2. + \#. $75 \mathrm{HCHO}+$ \#. 25 MEOH |
| RRR2 | 3.50e-14 | 3.50e-14 |  |  | RO2-R. + RO2-R. $=$ HO2. |
| R2NO |  | ame k as rxn RRNO |  |  | $\mathrm{R} 2 \mathrm{O} 2 .+\mathrm{NO}=\mathrm{NO} 2$ |
| R2H2 |  | Same k as rxn RRH2 |  |  | $\mathrm{R} 2 \mathrm{O} 2 .+\mathrm{HO} 2 .=\mathrm{HO} 2$. |
| R2N3 |  | Same k as rxn RRN3 |  |  | $\mathrm{R} 2 \mathrm{O} 2 .+\mathrm{NO} 3=\mathrm{NO} 2$ |
| R2ME |  | Same $k$ as rxn RRME |  |  | $\mathrm{R} 2 \mathrm{O} 2 .+\mathrm{C}-\mathrm{O} 2 .=\mathrm{C}-\mathrm{O} 2$. |
| R2RR |  | Same k as rxn RRR2 |  |  | R2O2. $+\mathrm{RO} 2-\mathrm{R} .=\mathrm{RO} 2-\mathrm{R}$. |
| R2R3 |  | Same k as rxn RRR2 |  |  | R2O2. + R2O2. $=$ |
| RNNO |  | ame k as rxn RRNO |  |  | $\mathrm{RO} 2-\mathrm{N} .+\mathrm{NO}=\mathrm{RNO} 3$ |
| RNH2 |  | Same k as rx | RRH2 |  | RO2-N. + HO2. $=$ ROOH + \#3 XC |

Table A-2 (continued)

| Label | Rate Parameters [a] |  |  | Reaction and Products [b] |
| :---: | :---: | :---: | :---: | :---: |
|  | k(298) | A Ea | B |  |
| RNME |  | Same k as rxn RRME |  | $\begin{aligned} & \text { RO2-N. + C-O2. }=\text { HO2. + \#. } 25 \mathrm{MEOH}+\text { \#. } 5\{\mathrm{MEK}+\mathrm{PROD} 2\}+ \\ & \text { \#. } 75 \mathrm{HCHO}+\mathrm{XC} \end{aligned}$ |
| RNN3 |  | Same k as rxn RRN3 |  | RO2-N. $+\mathrm{NO} 3=\mathrm{NO} 2+\mathrm{O} 2+\mathrm{HO} 2 .+\mathrm{MEK}+$ \#2 XC |
| RNRR |  | Same k as rxn RRR2 |  | RO2-N. + RO2-R. $=$ HO2. + \#. $5\{\mathrm{MEK}+\mathrm{PROD} 2\}+\mathrm{O} 2+\mathrm{XC}$ |
| RNR2 |  | Same k as rxn RRR2 |  | RO2-N. + R2O2. $=$ RO2-N. |
| RNRN |  | Same k as rxn RRR2 |  | RO2-N. + RO2-N. $=$ MEK + HO2 + PROD2 + O2 + \#2 XC |
| APN2 | 1.05e-11 | 1 Falloff, F=0.30 |  | CCO-O2. + NO2 $=$ PAN |
|  |  | 0: $2.70 \mathrm{e}-280.00$ | -7.1 |  |
|  |  | f: $1.20 \mathrm{e}-11 \quad 0.00$ | -0.9 |  |
| DPAN | 5.21e-4 | Falloff, F=0.30 |  | $\mathrm{PAN}=\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{NO} 2$ |
|  |  | 0: 4.90e-3 24.05 | 0.0 |  |
|  |  | \% $4.00 \mathrm{e}+16 \quad 27.03$ | 0.0 |  |
| APNO | 2.13e-11 | $17.80 \mathrm{e}-12-0.60$ |  | $\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{NO}=\mathrm{C}-\mathrm{O} 2 .+\mathrm{CO} 2+\mathrm{NO} 2$ |
| APH2 | 1.41e-11 | $1 \begin{array}{lll}4.30 \mathrm{e}-13 & -2.07\end{array}$ |  | $\begin{aligned} & \mathrm{CCO}-\mathrm{O} 2 .+\mathrm{HO} 2 .=\# .75\{\mathrm{CCO}-\mathrm{OOH}+\mathrm{O} 2\}+\# .25\{\mathrm{CCO}-\mathrm{OH}+ \\ & \mathrm{O} 3\} \end{aligned}$ |
| APN3 | 4.00e-12 | 4.00e-12 |  | $\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{NO} 3=\mathrm{C}-\mathrm{O} 2 .+\mathrm{CO} 2+\mathrm{NO} 2+\mathrm{O} 2$ |
| APME | $9.64 \mathrm{e}-12$ | 2 1.80e-12 -0.99 |  | $\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{C}-\mathrm{O} 2 .=\mathrm{CCO}-\mathrm{OH}+\mathrm{HCHO}+\mathrm{O} 2$ |
| APRR | $7.50 \mathrm{e}-12$ | 7.50e-12 |  | CCO-O2. + RO2-R. $=\mathrm{CCO}-\mathrm{OH}$ |
| APR2 |  | Same k as rxn APRR |  | $\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{R} 2 \mathrm{O} 2 .=\mathrm{CCO}-\mathrm{O} 2$. |
| APRN |  | Same k as rxn APRR |  | CCO-O2. + RO2-N. = CCO-OH + PROD2 |
| APAP | 1.55e-11 | $12.90 \mathrm{e}-12-0.99$ |  | $\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{CCO}-\mathrm{O} 2 .=\# 2\{\mathrm{C}-\mathrm{O} 2 .+\mathrm{CO} 2\}+\mathrm{O} 2$ |
| PPN2 | 1.21e-11 | $\begin{array}{lll}1.20 \mathrm{e}-11 & 0.00\end{array}$ | -0.9 | RCO-O2. + NO2 $=$ PAN2 |
| PAN2 | $4.43 \mathrm{e}-4$ | $2.00 \mathrm{e}+15 \quad 25.44$ |  | $\mathrm{PAN} 2=\mathrm{RCO}-\mathrm{O} 2 .+\mathrm{NO} 2$ |
| PPNO | $2.80 \mathrm{e}-11$ | 1.25e-11 -0.48 |  | $\mathrm{RCO}-\mathrm{O} 2+\mathrm{NO}=\mathrm{NO} 2+\mathrm{CCHO}+\mathrm{RO} 2-\mathrm{R} .+\mathrm{CO} 2$ |
| PPH2 |  | Same k as rxn APH2 |  | $\begin{aligned} & \mathrm{RCO}-\mathrm{O} 2 .+\mathrm{HO} 2 .=\# .75\{\mathrm{RCO}-\mathrm{OOH}+\mathrm{O} 2\}+\# .25\{\mathrm{RCO}-\mathrm{OH}+ \\ & \mathrm{O} 3\} \end{aligned}$ |
| PPN3 |  | Same k as rxn APN3 |  | $\mathrm{RCO}-\mathrm{O} 2 .+\mathrm{NO} 3=\mathrm{NO} 2+\mathrm{CCHO}+\mathrm{RO} 2-\mathrm{R} .+\mathrm{CO} 2+\mathrm{O} 2$ |
| PPME |  | Same k as rxn APME |  | $\mathrm{RCO}-\mathrm{O} 2 .+\mathrm{C}-\mathrm{O} 2 .=\mathrm{RCO}-\mathrm{OH}+\mathrm{HCHO}+\mathrm{O} 2$ |
| PPRR |  | Same k as rxn APRR |  | RCO-O2. + RO2-R. $=\mathrm{RCO}-\mathrm{OH}+\mathrm{O} 2$ |
| PPR2 |  | Same k as rxn APRR |  | RCO-O2. + R2O2. $=$ RCO-O2. |
| PPRN |  | Same k as rxn APRR |  | $\mathrm{RCO}-\mathrm{O} 2+\mathrm{RO} 2-\mathrm{N} .=\mathrm{RCO}-\mathrm{OH}+\mathrm{PROD} 2+\mathrm{O} 2$ |
| PPAP |  | Same k as rxn APAP |  | $\mathrm{RCO}-\mathrm{O} 2 .+\mathrm{CCO}-\mathrm{O} 2 .=$ \#2 CO2 $+\mathrm{C}-\mathrm{O} 2 .+\mathrm{CCHO}+\mathrm{RO} 2-\mathrm{R} .+\mathrm{O} 2$ |
| PPPP |  | Same k as rxn APAP |  | RCO-O2. $+\mathrm{RCO}-\mathrm{O} 2 .=$ \#2 \{ $\mathrm{CCHO}+\mathrm{RO} 2-\mathrm{R} .+\mathrm{CO} 2\}$ |
| BPN2 | 1.37e-11 | $1.37 \mathrm{e}-11$ |  | BZCO-O2. + NO2 $=$ PBZN |
| BPAN | 3.12e-4 | $7.90 \mathrm{e}+16 \quad 27.82$ |  | PBZN $=$ BZCO-O2. + NO2 |
| BPNO |  | Same k as rxn PPNO |  | BZCO-O2. $+\mathrm{NO}=\mathrm{NO} 2+\mathrm{CO} 2+\mathrm{BZ}-\mathrm{O} .+\mathrm{R} 2 \mathrm{O} 2$. |
| BPH2 |  | Same $k$ as rxn APH2 |  | $\begin{aligned} & \mathrm{BZCO}-\mathrm{O} 2 .+\mathrm{HO} 2 .=\# .75\{\mathrm{RCO}-\mathrm{OOH}+\mathrm{O} 2\}+\# .25\{\mathrm{RCO}-\mathrm{OH}+ \\ & \mathrm{O} 3\}+\# 4 \mathrm{XC} \end{aligned}$ |
| BPN3 |  | Same k as rxn APN3 |  | BZCO-O2. $+\mathrm{NO} 3=\mathrm{NO} 2+\mathrm{CO} 2+\mathrm{BZ}-\mathrm{O} .+\mathrm{R} 2 \mathrm{O} 2 .+\mathrm{O} 2$ |
| BPME |  | Same k as rxn APME |  | BZCO-O2. $+\mathrm{C}-\mathrm{O} 2 .=\mathrm{RCO}-\mathrm{OH}+\mathrm{HCHO}+\mathrm{O} 2+$ \#4 XC |
| BPRR |  | Same k as rxn APRR |  | BZCO-O2. $+\mathrm{RO} 2-\mathrm{R} .=\mathrm{RCO}-\mathrm{OH}+\mathrm{O} 2+$ \#4 XC |
| BPR2 |  | Same k as rxn APRR |  | BZCO-O2. + R2O2. = BZCO-O2. |
| BPRN |  | Same k as rxn APRR |  | BZCO-O2. + RO2-N. $=$ RCO-OH + PROD2 + O2 + \#4 XC |
| BPAP |  | Same k as rxn APAP |  | BZCO-O2. $+\mathrm{CCO}-\mathrm{O} 2 .=$ \#2 CO2 $+\mathrm{C}-\mathrm{O} 2 .+\mathrm{BZ}-\mathrm{O} .+\mathrm{R} 2 \mathrm{O} 2$. |
| BPPP |  | Same k as rxn APAP |  | $\begin{aligned} & \mathrm{BZCO}-\mathrm{O} 2 .+\mathrm{RCO}-\mathrm{O} 2 .=\# 2 \mathrm{CO} 2+\mathrm{CCHO}+\mathrm{RO} 2-\mathrm{R} .+\mathrm{BZ}-\mathrm{O} .+ \\ & \mathrm{R} 2 \mathrm{O} 2 . \end{aligned}$ |
| BPBP |  | Same k as rxn APAP |  | BZCO-O2. + BZCO-O2. $=$ \#2 \{BZ-O. + R2O2. $+\mathrm{CO} 2\}$ |
| MPN2 |  | Same $k$ as rxn PPN2 |  | MA-RCO3 + NO2 $=$ MA-PAN |

Table A-2 (continued)

| Label | Rate Parameters [a] |  |  |  | Reaction and Products [b] |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | k(298) | A | Ea | B |  |
| MPPN | $3.55 \mathrm{e}-4$ | $1.60 \mathrm{e}+16$ | 26.80 |  | MA-PAN $=$ MA-RCO3 + NO2 |
| MPNO |  | Same k as rx | PPNO |  | $\mathrm{MA}-\mathrm{RCO} 3 .+\mathrm{NO}=\mathrm{NO} 2+\mathrm{CO} 2+\mathrm{HCHO}+\mathrm{CCO}-\mathrm{O} 2$. |
| MPH2 |  | Same k as rx | APH2 |  | $\begin{aligned} & \mathrm{MA}-\mathrm{RCO} 3 .+\mathrm{HO} 2 .=\# .75\{\mathrm{RCO}-\mathrm{OOH}+\mathrm{O} 2\}+\# .25\{\mathrm{RCO}-\mathrm{OH}+ \\ & \mathrm{O} 3\}+\mathrm{XC} \end{aligned}$ |
| MPN3 |  | Same k as rx | APN3 |  | $\mathrm{MA}-\mathrm{RCO} 3 .+\mathrm{NO} 3=\mathrm{NO} 2+\mathrm{CO} 2+\mathrm{HCHO}+\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{O} 2$ |
| MPME |  | Same k as rxn | APME |  | MA-RCO3. $+\mathrm{C}-\mathrm{O} 2 .=\mathrm{RCO}-\mathrm{OH}+\mathrm{HCHO}+\mathrm{XC}+\mathrm{O} 2$ |
| MPRR |  | Same k as rx | APRR |  | MA-RCO3. $+\mathrm{RO} 2-\mathrm{R} .=\mathrm{RCO}-\mathrm{OH}+\mathrm{XC}$ |
| MPR2 |  | Same k as rx | APRR |  | $\mathrm{MA}-\mathrm{RCO} 3 .+\mathrm{R} 2 \mathrm{O} 2 .=\mathrm{MA}-\mathrm{RCO} 3$. |
| MPRN |  | Same k as rx | APRR |  | MA-RCO3. $+\mathrm{RO} 2-\mathrm{N} .=$ \#2 RCO-OH $+\mathrm{O} 2+$ \#4 XC |
| MPAP |  | Same k as rx | APAP |  | $\begin{aligned} & \mathrm{MA}-\mathrm{RCO} 3 .+\mathrm{CCO}-\mathrm{O} 2 .=\# 2 \mathrm{CO} 2+\mathrm{C}-\mathrm{O} 2 .+\mathrm{HCHO}+\mathrm{CCO}-\mathrm{O} 2 .+ \\ & \mathrm{O} 2 \end{aligned}$ |
| MPPP |  | Same k as rx | APAP |  | $\begin{aligned} & \text { MA-RCO3. }+\mathrm{RCO}-\mathrm{O} 2 .=\mathrm{HCHO}+\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{CCHO}+\mathrm{RO} 2-\mathrm{R} . \\ & +\# 2 \mathrm{CO} 2 \end{aligned}$ |
| MPBP |  | Same k as rx | APAP |  | $\begin{aligned} & \text { MA-RCO } 3 .+\mathrm{BZCO}-\mathrm{O} 2 .=\mathrm{HCHO}+\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{BZ}-\mathrm{O} .+\mathrm{R} 2 \mathrm{O} 2 . \\ & +\# 2 \mathrm{CO} 2 \end{aligned}$ |
| MPMP |  | Same k as rx | APAP |  | MA-RCO3. $+\mathrm{MA}-\mathrm{RCO} 3 .=\# 2\{\mathrm{HCHO}+\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{CO} 2\}$ |
| Other Organic Radical Species |  |  |  |  |  |
| TBON | $2.40 \mathrm{e}-11$ | 1 2.40e-11 |  |  | TBU-O. + NO2 = RNO3 + \#-2 XC |
| TBOD | $9.87 \mathrm{e}+2$ | 7.50e+14 | 16.20 |  | TBU-O. $=\mathrm{ACET}+\mathrm{C}-\mathrm{O} 2$. |
| BRN2 | $3.80 \mathrm{e}-11$ | $1 \quad 2.30 \mathrm{e}-11$ | -0.30 |  | BZ-O. $+\mathrm{NO} 2=$ NPHE |
| BRH2 |  | Same k as rx | RRH2 |  | BZ-O. + HO2. = PHEN |
| BRXX | $1.00 \mathrm{e}-3$ | $1.00 \mathrm{e}-3$ |  |  | BZ-O. = PHEN |
| BNN2 |  | Same k as rx | BRN2 |  | BZ(NO2)-O. + NO2 = \#2 XN + \#6 XC |
| BNH2 |  | Same k as rx | RRH2 |  | BZ(NO2)-O. + HO2. = NPHE |
| BNXX |  | Same k as rxi | BRXX |  | BZ(NO2)-O. $=$ NPHE |
| Explicit and Lumped Molecule Organic Products |  |  |  |  |  |
| FAHV |  | Phot Set= H | HO_R |  | $\mathrm{HCHO}+\mathrm{HV}=\# 2 \mathrm{HO} 2 .+\mathrm{CO}$ |
| FAVS |  | Phot Set= H | HO_M |  | $\mathrm{HCHO}+\mathrm{HV}=\mathrm{H} 2+\mathrm{CO}$ |
| FAOH | $9.20 \mathrm{e}-12$ | 2 8.60e-12 | -0.04 |  | $\mathrm{HCHO}+\mathrm{HO} .=\mathrm{HO} 2 .+\mathrm{CO}+\mathrm{H} 2 \mathrm{O}$ |
| FAH2 | $7.90 \mathrm{e}-14$ | $49.70 \mathrm{e}-15$ | -1.24 |  | $\mathrm{HCHO}+\mathrm{HO} 2 .=\mathrm{HOCOO}$. |
| FAHR | $1.51 \mathrm{e}+2$ | $22.40 \mathrm{e}+12$ | 13.91 |  | HOCOO. $=\mathrm{HO} 2 .+\mathrm{HCHO}$ |
| FAHN |  | Same k as rxa | MER1 |  | $\mathrm{HOCOO} .+\mathrm{NO}=\mathrm{HCOOH}+\mathrm{NO} 2+\mathrm{HO} 2$. |
| FAN3 | 5.74e-16 | $6 \quad 2.00 \mathrm{e}-12$ | 4.83 |  | $\mathrm{HCHO}+\mathrm{NO} 3=\mathrm{HNO} 3+\mathrm{HO} 2 .+\mathrm{CO}$ |
| AAOH | $1.58 \mathrm{e}-11$ | $15.60 \mathrm{e}-12$ | -0.62 |  | $\mathrm{CCHO}+\mathrm{HO} .=\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{H} 2 \mathrm{O}$ |
| AAHV |  | Phot Set= C | HO_R |  | $\mathrm{CCHO}+\mathrm{HV}=\mathrm{CO}+\mathrm{HO} 2 .+\mathrm{C}-\mathrm{O} 2$. |
| AAN3 | $2.73 \mathrm{e}-15$ | $5 \quad 1.40 \mathrm{e}-12$ | 3.70 |  | $\mathrm{CCHO}+\mathrm{NO} 3=\mathrm{HNO} 3+\mathrm{CCO}-\mathrm{O} 2$. |
| PAOH | $2.00 \mathrm{e}-11$ | $1 \quad 2.00 \mathrm{e}-11$ |  |  | $\begin{aligned} & \mathrm{RCHO}+\mathrm{HO} .=\text { \#. } 034 \mathrm{RO} 2-\mathrm{R} .+\# .001 \mathrm{RO} 2-\mathrm{N} .+ \text { \#. } 965 \mathrm{RCO}-\mathrm{O} 2 . \\ & +\# .034 \mathrm{CO}+\# .034 \mathrm{CCHO}+\#-0.003 \mathrm{XC} \end{aligned}$ |
| PAHV |  | Phot Set= | 2 CHO |  | $\mathrm{RCHO}+\mathrm{HV}=\mathrm{CCHO}+\mathrm{RO} 2-\mathrm{R} .+\mathrm{CO}+\mathrm{HO} 2$. |
| PAN3 | 3.67e-15 | $5 \quad 1.40 \mathrm{e}-12$ | 3.52 |  | $\mathrm{RCHO}+\mathrm{NO} 3=\mathrm{HNO} 3+\mathrm{RCO}-\mathrm{O} 2$. |
| K 3 OH | $1.92 \mathrm{e}-13$ | $3 \quad 1.10 \mathrm{e}-12$ | 1.03 |  | $\mathrm{ACET}+\mathrm{HO} .=\mathrm{HCHO}+\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{R} 2 \mathrm{O} 2$. |
| K3HV |  | Phot Set= A | ETONE |  | $\mathrm{ACET}+\mathrm{HV}=\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{C}-\mathrm{O} 2$. |
| K 4 OH | $1.18 \mathrm{e}-12$ | 2 1.30e-12 | 0.05 | 2.0 | $\begin{aligned} & \text { MEK + HO. = \#. } 37 \text { RO2-R. + \#. } 042 \text { RO2-N. + \#. } 616 \text { R2O2. + } \\ & \text { \#. } 492 \text { CCO-O2. + \#. } 096 \text { RCO-O2. + \#. } 115 \text { HCHO + \#. } 482 \text { CCHO } \\ & \text { + \#. } 37 \text { RCHO + \#. } 287 \text { XC } \end{aligned}$ |
| K4HV | Phot S | Set $=$ KETON | , $\mathrm{qy}=1$. | e-1 | $\mathrm{MEK}+\mathrm{HV}=\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{CCHO}+\mathrm{RO} 2-\mathrm{R}$. |
| MeOH | $9.14 \mathrm{e}-13$ | 3 3.10e-12 | 0.72 | 2.0 | $\mathrm{MEOH}+\mathrm{HO} .=\mathrm{HCHO}+\mathrm{HO} 2$. |

Table A-2 (continued)


Table A-2 (continued)

| Label | Rate Parameters [a] |  |  |  | Reaction and Products [b] |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | k(298) | A | Ea | B |  |
| IPO3 | $4.18 \mathrm{e}-18$ | $4.18 \mathrm{e}-18$ |  |  | $\begin{aligned} & \text { ISO-PROD + O3 = \#. } 4 \mathrm{HO} 2 .+ \text { \#. } 048 \mathrm{RO} 2-\mathrm{R} .+ \text { + } .048 \mathrm{RCO}-\mathrm{O} 2 .+ \\ & \text { \#. } 285 \mathrm{HO} .+ \text { \#. } 498 \mathrm{CO}+\# .14 \mathrm{CO} 2+\text { \#. } 125 \mathrm{HCHO}+\# .047 \mathrm{CCHO} \\ & \text { + \#. } 21 \mathrm{MEK}+\# .023 \mathrm{GLY}+\# .742 \mathrm{MGLY}+\text { \#. } 1 \mathrm{HCOOH}+\# .372 \\ & \text { RCO-OH + \#-. } 33 \mathrm{XC} \end{aligned}$ |
| IPN3 | $1.00 \mathrm{e}-13$ | $1.00 \mathrm{e}-13$ |  |  | ISO-PROD + NO3 = \#. 799 RO2-R. + \#. 051 RO2-N. + \#. 15 MARCO3. + \#. $572 \mathrm{CO}+$ \#. $15 \mathrm{HNO} 3+$ \#. $227 \mathrm{HCHO}+$ \#. $218 \mathrm{RCHO}+$ \#. 008 MGLY + \#. 572 RNO3 + \#. 28 XN + \#-. 815 XC |
| IPHV | Phot Set | = ACROLE | , qy | e-3 | $\begin{aligned} & \text { ISO-PROD + HV = \#1.233 HO2. + \#. } 467 \mathrm{CCO}-\mathrm{O} 2 .+ \text { \#. } 3 \mathrm{RCO}- \\ & \mathrm{O} 2 .+\# 1.233 \mathrm{CO}+\# .3 \mathrm{HCHO}+\# .467 \mathrm{CCHO}+\# .233 \mathrm{MEK}+\#- \\ & .233 \mathrm{XC} \end{aligned}$ |
| Lumped Parameter Organic Products |  |  |  |  |  |
| K6OH | $1.50 \mathrm{e}-11$ | $1.50 \mathrm{e}-11$ |  |  | $\begin{aligned} & \text { PROD } 2+\mathrm{HO} .=\text { \#. } 379 \mathrm{HO} 2 .+ \text { \#. } 473 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 07 \mathrm{RO} 2-\mathrm{N} .+ \\ & \text { \#. } 029 \mathrm{CCO}-\mathrm{O} 2 .+ \text { \#. } 049 \mathrm{RCO}-\mathrm{O} 2 .+ \text { \#. } 213 \mathrm{HCHO}+\text { \#. } 084 \mathrm{CCHO} \\ & \text { + \#. } 558 \mathrm{RCHO}+\text { \#. } 115 \mathrm{MEK}+\text { \#. } 329 \text { PROD } 2+\text { \#. } 886 \mathrm{XC} \end{aligned}$ |
| K6HV | Phot Se | t $=$ KETON | qy |  | $\begin{aligned} & \text { PROD } 2+\mathrm{HV}=\text { \#. } 96 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 04 \mathrm{RO} 2-\mathrm{N} .+ \text { \#. } 515 \mathrm{R} 2 \mathrm{O} 2 .+ \\ & \text { \#. } 667 \mathrm{CCO}-\mathrm{O} 2 .+ \text { \#. } 333 \mathrm{RCO}-\mathrm{O} 2 .+ \text { \#. } 506 \mathrm{HCHO}+\text { \#. } 246 \mathrm{CCHO} \\ & \text { + \#. } 71 \mathrm{RCHO}+\# .299 \mathrm{XC} \end{aligned}$ |
| RNOH | $7.80 \mathrm{e}-12$ | 7.80e-12 |  |  | $\begin{aligned} & \mathrm{RNO} 3+\mathrm{HO} .=\text { \#. } 338 \mathrm{NO} 2+\# .113 \mathrm{HO} 2 .+ \text { \#. } 376 \mathrm{RO} 2-\mathrm{R} .+\# .173 \\ & \mathrm{RO} 2-\mathrm{N} .+ \text { \#. } 596 \mathrm{R} 2 \mathrm{O} 2 .+ \text { \#. } 01 \mathrm{HCHO}+\# .439 \mathrm{CCHO}+\# .213 \\ & \mathrm{RCHO}+\# .006 \mathrm{ACET}+\# .177 \mathrm{MEK}+\# .048 \mathrm{PROD} 2+\# .31 \mathrm{RNO} \\ & +\# .351 \mathrm{XN}+\# .56 \mathrm{XC} \end{aligned}$ |
| RNHV |  | hot Set= IC | ONO2 |  | RNO3 + HV = NO2 + \#. $341 \mathrm{HO} 2 .+$ \#. 564 RO2-R. + \#. 095 RO2N. + \#. 152 R2O2. + \#. 134 HCHO + \#. $431 \mathrm{CCHO}+$ \#. $147 \mathrm{RCHO}+$ \#. 02 ACET + \#. 243 MEK + \#. 435 PROD2 + \#. 35 XC |
| Uncharacterized Reactive Aromatic Ring Fragmentation Products |  |  |  |  |  |
| D1OH | 5.00e-11 | $5.00 \mathrm{e}-11$ |  |  | $\mathrm{DCB} 1+\mathrm{HO} .=\mathrm{RCHO}+\mathrm{RO} 2-\mathrm{R} .+\mathrm{CO}$ |
| D1HV |  | (Slow) |  |  | $\mathrm{DCB} 1+\mathrm{HV}=\mathrm{HO} 2 .+$ \# $2 \mathrm{CO}+\mathrm{RO} 2-\mathrm{R} .+\mathrm{GLY}+\mathrm{R} 2 \mathrm{O} 2$. |
| D1O3 | 2.00e-18 | $2.00 \mathrm{e}-18$ |  |  | $\mathrm{DCB} 1+\mathrm{O} 3=\# 1.5 \mathrm{HO} 2 .+$ \#.5 HO. + \#1.5 CO + \#.5 CO2 + GLY |
| D2OH | $5.00 \mathrm{e}-11$ | $5.00 \mathrm{e}-11$ |  |  | $\mathrm{DCB} 2+\mathrm{HO} .=\mathrm{R} 2 \mathrm{O} 2 .+\mathrm{RCHO}+\mathrm{CCO}-\mathrm{O} 2$. |
| D2HV | Phot Set= | = MGLY_A | BS, qy | .7e-1 | $\begin{aligned} & \mathrm{DCB} 2+\mathrm{HV}=\mathrm{RO} 2-\mathrm{R} .+\# .5\{\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{HO} 2 .\}+\mathrm{CO}+\mathrm{R} 2 \mathrm{O} 2 .+ \\ & \# .5\{\mathrm{GLY}+\mathrm{MGLY}+\mathrm{XC}\} \end{aligned}$ |
| D3OH | 5.00e-11 | $5.00 \mathrm{e}-11$ |  |  | $\mathrm{DCB} 3+\mathrm{HO} .=\mathrm{R} 2 \mathrm{O} 2 .+\mathrm{RCHO}+\mathrm{CCO}-\mathrm{O} 2$. |
| D3HV | Phot Set= | = ACROLE | $\mathrm{N}, \mathrm{qy}=$ | $3 \mathrm{e}+0$ | $\begin{aligned} & \text { DCB3 + HV = RO2-R. + \#. } 5\{\mathrm{CCO}-\mathrm{O} 2 .+\mathrm{HO} 2 .\}+\mathrm{CO}+\mathrm{R} 2 \mathrm{O} 2 .+ \\ & \# .5\{\mathrm{GLY}+\mathrm{MGLY}+\mathrm{XC}\} \end{aligned}$ |
| Base ROG VOCs Used in the Chamber Simulations and Explicit VOCs in the Ambient Simulations |  |  |  |  |  |
| c1OH | 6.37e-15 | $2.15 \mathrm{e}-12$ | 3.45 |  | $\mathrm{CH} 4+\mathrm{HO} .=\mathrm{H} 2 \mathrm{O}+\mathrm{C}-\mathrm{O} 2$. |
| c4OH | $2.44 \mathrm{e}-12$ | $1.52 \mathrm{e}-12$ | -0.29 | 2.0 | $\begin{aligned} & \mathrm{N}-\mathrm{C} 4+\mathrm{HO} .=\# .921 \mathrm{RO} 2-\mathrm{R} .+\# .079 \mathrm{RO} 2-\mathrm{N} .+\# .413 \mathrm{R} 2 \mathrm{O} 2 .+\# .632 \\ & \mathrm{CCHO}+\# .12 \mathrm{RCHO}+\# .485 \text { MEK }+ \text { \#- } 0.038 \text { XC } \end{aligned}$ |
| c6OH | $5.47 \mathrm{e}-12$ | $1.38 \mathrm{e}-12$ | -0.82 | 2.0 | $\begin{aligned} & \mathrm{N}-\mathrm{C} 6+\mathrm{HO} .=\# .775 \mathrm{RO} 2-\mathrm{R} .+\# .225 \mathrm{RO} 2-\mathrm{N} .+\# .787 \mathrm{R} 2 \mathrm{O} 2 .+\# .011 \\ & \mathrm{CCHO}+\# .113 \mathrm{RCHO}+\# .688 \text { PROD} 2+\# .162 \text { XC } \end{aligned}$ |
| c8OH | $8.70 \mathrm{e}-12$ | $2.48 \mathrm{e}-12$ | -0.75 | 2.0 | $\begin{aligned} & \mathrm{N}-\mathrm{C} 8+\mathrm{HO} .=\# .646 \mathrm{RO} 2-\mathrm{R} .+\# .354 \mathrm{RO} 2-\mathrm{N} .+\# .786 \mathrm{R} 2 \mathrm{O} 2 .+\# .024 \\ & \mathrm{RCHO}+\# .622 \text { PROD} 2+\# 2.073 \mathrm{XC} \end{aligned}$ |
| etOH | 8.52e-12 | 1.96e-12 | -0.87 |  | ETHENE + HO. = RO2-R. + \#1.61 HCHO + \#. 195 CCHO |
| etO3 | $1.59 \mathrm{e}-18$ | 9.14e-15 | 5.13 |  | $\begin{aligned} & \text { ETHENE }+\mathrm{O} 3=\text { \#. } 12 \mathrm{HO} .+ \text { \#. } 12 \mathrm{HO} 2 .+ \text { \#. } 5 \mathrm{CO}+\# .13 \mathrm{CO} 2+ \\ & \mathrm{HCHO}+\# .37 \mathrm{HCOOH} \end{aligned}$ |
| etN3 | 2.05e-16 | $4.39 \mathrm{e}-13$ | 4.53 | 2.0 | ETHENE + NO3 = RO2-R. + RCHO + \#-1 XC + XN |
| etOA | $7.29 \mathrm{e}-13$ | 1.04e-11 | 1.57 |  | $\begin{aligned} & \text { ETHENE + O3P }=\# .5 \mathrm{HO} 2 .+ \text { \#. } 2 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 3 \mathrm{C}-\mathrm{O} 2 .+\# .491 \\ & \mathrm{CO}+\# .191 \mathrm{HCHO}+\# .25 \mathrm{CCHO}+\# .009 \mathrm{GLY}+\# .5 \mathrm{XC} \end{aligned}$ |

Table A-2 (continued)

| Label | Rate Parameters [a] |  |  |  | Reaction and Products [b] |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | k(298) | A | Ea | B |  |
| prOH | $2.63 \mathrm{e}-11$ | $4.85 \mathrm{e}-12$ | $-1.00$ |  | $\begin{aligned} & \text { PROPENE + HO. = \#. } 984 \text { RO2-R. + \#. } 016 \text { RO2-N. }+ \text { \#. } 984 \text { HCHO + } \\ & \text { \#. } 984 \text { CCHO + \#-0.048 XC } \end{aligned}$ |
| prO3 | $1.01 \mathrm{e}-17$ | 5.51e-15 | 3.73 |  | PROPENE + O3 = \#. 32 HO. + \#. 06 HO2. + \#. 26 C-O2. + \#. $51 \mathrm{CO}+$ \#. 135 $\mathrm{CO} 2+$ \#. $5 \mathrm{HCHO}+$ \#. $5 \mathrm{CCHO}+$ \#. $185 \mathrm{HCOOH}+$ \#. $17 \mathrm{CCO}-\mathrm{OH}+$ \#. 07 INERT + \#. 07 XC |
| prN3 | $9.49 \mathrm{e}-15$ | $4.59 \mathrm{e}-13$ | 2.30 |  | PROPENE + NO3 = \#.949 RO2-R. + \#.051 RO2-N. + \#2.693 XC + XN |
| prOP | $3.98 \mathrm{e}-12$ | $1.18 \mathrm{e}-11$ | 0.64 |  | PROPENE + O3P = \#. $45 \mathrm{RCHO}+$ \#. $55 \mathrm{MEK}+$ \#-0.55 XC |
| t2OH | $6.40 \mathrm{e}-11$ | $1.01 \mathrm{e}-11$ | -1.09 |  | $\begin{aligned} & \text { T-2-BUTE + HO. }=\text { \#. } 965 \text { RO2-R. }+ \text { \#. } 035 \text { RO2-N. }+ \text { \#1. } 93 \text { CCHO + \#- } \\ & 0.07 \mathrm{XC} \end{aligned}$ |
| t2O3 | $1.90 \mathrm{e}-16$ | $6.64 \mathrm{e}-15$ | 2.10 |  | $\begin{aligned} & \mathrm{T}-2-\mathrm{BUTE}+\mathrm{O} 3=\# .52 \mathrm{HO} .+ \text { \#. } 52 \mathrm{C}-\mathrm{O} 2 .+\# .52 \mathrm{CO}+\# .14 \mathrm{CO} 2+ \\ & \mathrm{CCHO}+\# .34 \mathrm{CCO}-\mathrm{OH}+\# .14 \mathrm{INERT}+\# .14 \mathrm{XC} \end{aligned}$ |
| t2N3 | $3.91 \mathrm{e}-13$ | $1.10 \mathrm{e}-13$ | -0.76 | 2.0 | $\begin{aligned} & \text { T-2-BUTE }+\mathrm{NO} 3=\# .705 \mathrm{NO} 2+\# .215 \mathrm{RO} 2-\mathrm{R} .+\# .08 \mathrm{RO} 2-\mathrm{N} .+\# .705 \\ & \mathrm{R} 2 \mathrm{O} 2 .+\# 1.41 \mathrm{CCHO}+\# .215 \mathrm{RNO}+\#-0.59 \mathrm{XC}+\# .08 \mathrm{XN} \end{aligned}$ |
| t2OP | $2.18 \mathrm{e}-11$ | $2.18 \mathrm{e}-11$ |  |  | T-2-BUTE + O3P = MEK |
| isOH | $9.82 \mathrm{e}-11$ | $2.50 \mathrm{e}-11$ | -0.81 |  | ISOPRENE + HO. $=$ \#. 907 RO2-R. + \#. 093 RO2-N. + \#. 079 R2O2. + \#. 624 HCHO + \#. 23 METHACRO + \#. 32 MVK + \#. 357 ISO-PROD + \#-0.167 XC |
| isO3 | 1.28e-17 | 7.86e-15 | 3.80 |  | $\begin{aligned} & \text { ISOPRENE + O3 = \#. } 266 \mathrm{HO} .+ \text { \#. } 066 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 008 \mathrm{RO} 2-\mathrm{N} .+ \\ & \text { \#. } 126 \mathrm{R} 2 \mathrm{O} 2 .+ \text { \#. } 192 \mathrm{MA}-\mathrm{RCO} .+ \text { \#. } 275 \mathrm{CO}+\text { \#. } 122 \mathrm{CO} 2+ \\ & \text { \#. } 592 \mathrm{HCHO}+\text { \#. } 1 \text { PROD2 + \#. } 39 \mathrm{METHACRO}+\text { \#. } 16 \mathrm{MVK}+ \\ & \text { \#. } 204 \mathrm{HCOOH}+\text { \#. } 15 \mathrm{RCO}-\mathrm{OH}+\text { \#- } 0.259 \mathrm{XC} \end{aligned}$ |
| isN3 | $6.74 \mathrm{e}-13$ | $3.03 \mathrm{e}-12$ | 0.89 |  | $\begin{aligned} & \text { ISOPRENE + NO3 = \#. } 187 \mathrm{NO} 2+\text { \#. } 749 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 064 \mathrm{RO} 2-\mathrm{N} . \\ & \text { + \#. } 187 \mathrm{R} 2 \mathrm{O} 2 .+ \text { \#. } 936 \mathrm{ISO}-\mathrm{PROD}+\text { \#-0.064 XC + \#.813 XN } \end{aligned}$ |
| isOP | $3.60 \mathrm{e}-11$ | $3.60 \mathrm{e}-11$ |  |  | $\begin{aligned} & \text { ISOPRENE + O3P = \#. } 01 \mathrm{RO} 2-\mathrm{N} .+ \text { \#. } 24 \mathrm{R} 2 \mathrm{O} 2 .+ \text { \#. } 25 \mathrm{C}-\mathrm{O} 2 .+ \\ & \# .24 \mathrm{MA}-\mathrm{RCO} 3 .+ \text { \#. } 24 \mathrm{HCHO}+\text { \#. } 75 \mathrm{PROD} 2+\text { \#-1.01 XC } \end{aligned}$ |
| tlOH | $5.95 \mathrm{e}-12$ | $1.81 \mathrm{e}-12$ | -0.71 | 0.0 | TOLUENE + HO. $=$ \#. 234 HO2. + \#. 758 RO2-R. + \#. 008 RO2-N. + \#. 116 GLY + \#. 135 MGLY + \#. 234 CRES + \#. 085 BALD + \#. 46 DCB1 + \#. 156 DCB2 + \#. 057 DCB3 + \#1.178 XC |
| mxOH | $2.36 \mathrm{e}-11$ | $2.36 \mathrm{e}-11$ | 0.00 | 0.0 | M-XYLENE + HO. = \#. 21 HO2. + \#. 782 RO2-R. + \#. 008 RO2-N. + \#. 107 GLY + \#. 335 MGLY + \#. 21 CRES + \#. 037 BALD + \#. 347 DCB1 + \#. 29 DCB2 + \#. 108 DCB3 + \#1.628 XC |
| Lumped Organic Species used in the Ambient Reactivity Simulations |  |  |  |  |  |
| t1OH | 8.27e-11 | $1.83 \mathrm{e}-11$ | -0.89 |  | $\begin{aligned} & \text { TERP + HO. }=\text { \#. } 75 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 25 \mathrm{RO} 2-\mathrm{N} .+ \text { \#.5 R2O2. + \#. } 276 \\ & \mathrm{HCHO}+\# .474 \mathrm{RCHO}+\# .276 \text { PROD2 + \#5.146 XC } \end{aligned}$ |
| t1O3 | 6.88e-17 | 1.08e-15 | 1.63 |  | $\begin{aligned} & \text { TERP + O3 = \#.567 HO. + \#. } 033 \mathrm{HO} 2 .+\# .031 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 18 \\ & \text { RO2-N. + \#. } 729 \mathrm{R} 2 \mathrm{O} 2 .+ \text { \#. } 123 \mathrm{CCO}-\mathrm{O} 2 .+\# .201 \mathrm{RCO}-\mathrm{O} 2 .+ \\ & \# .157 \mathrm{CO}+\# .037 \mathrm{CO} 2+\# .235 \mathrm{HCHO}+\# .205 \mathrm{RCHO}+\# .13 \\ & \text { ACET + \#. } 276 \text { PROD2 + \#.001 GLY + \#. } 031 \mathrm{BACL}+\# .103 \\ & \text { HCOOH + \#. } 189 \text { RCO-OH + \#4.183 XC } \end{aligned}$ |
| t1N3 | $6.57 \mathrm{e}-12$ | 3.66e-12 | -0.35 |  | $\begin{aligned} & \mathrm{TERP}+\mathrm{NO} 3=\text { \#. } 474 \mathrm{NO} 2+\# .276 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 25 \mathrm{RO} 2-\mathrm{N} .+ \text { \#. } 75 \\ & \mathrm{R} 2 \mathrm{O} 2 .+ \text { \#. } 474 \mathrm{RCHO}+\# .276 \mathrm{RNO} 3+\# 5.421 \mathrm{XC}+\# .25 \mathrm{XN} \end{aligned}$ |
| t1OP | 3.27e-11 | $3.27 \mathrm{e}-11$ |  |  | $\mathrm{TERP}+\mathrm{O} 3 \mathrm{P}=$ \#. $147 \mathrm{RCHO}+$ \#. $853 \mathrm{PROD} 2+$ \#4.441 XC |
| a1OH | $2.54 \mathrm{e}-13$ | 1.37e-12 | 0.99 | 2.0 | ALK1 + HO. $=$ RO2-R. +CCHO |
| a2OH | $1.04 \mathrm{e}-12$ | $9.87 \mathrm{e}-12$ | 1.33 |  | $\begin{aligned} & \text { ALK2 + HO. }=\text { \#. } 246 \mathrm{HO} .+ \text { \#. } 121 \mathrm{HO} 2 .+ \text { \#. } 612 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 021 \\ & \text { RO2-N. + \#. } 16 \mathrm{CO}+\# .039 \mathrm{HCHO}+\# .155 \mathrm{RCHO}+\# .417 \text { ACET } \\ & \text { + \#. } 248 \text { GLY + \#. } 121 \mathrm{HCOOH}+\text { \#0.338 XC } \end{aligned}$ |
| a3OH | $2.38 \mathrm{e}-12$ | $1.02 \mathrm{e}-11$ | 0.86 |  | $\begin{aligned} & \text { ALK3 + HO. }=\text { \#. } 695 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 07 \mathrm{RO} 2-\mathrm{N} .+ \text { \#. } 559 \mathrm{R} 2 \mathrm{O} 2 .+ \\ & \text { \#. } 236 \text { TBU-O. + \#. } 026 \mathrm{HCHO}+\text { \#. } 445 \mathrm{CCHO}+\text { \#. } 122 \mathrm{RCHO}+ \\ & \text { \#. } 024 \text { ACET + \#. } 332 \mathrm{MEK}+\text { \#- } 0.05 \mathrm{XC} \end{aligned}$ |

Table A-2 (continued)

| Label | Rate Parameters [a] |  |  |  | Reaction and Products [b] |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | k(298) | A | Ea | B |  |
| a4OH | $4.39 \mathrm{e}-12$ | $5.95 \mathrm{e}-12$ | 0.18 |  | $\begin{aligned} & \text { ALK4 + HO. = \#. } 835 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 143 \mathrm{RO} 2-\mathrm{N} .+ \text { \#. } 936 \mathrm{R} 2 \mathrm{O} 2 .+ \\ & \text { \#. } 011 \mathrm{C}-\mathrm{O} 2 .+ \text { \#. } 011 \mathrm{CCO} 2 .+ \text { \#. } 002 \mathrm{CO}+\text { \#. } 024 \mathrm{HCHO}+\text { \#. } 455 \\ & \text { CCHO + \#. } 244 \mathrm{RCHO}+\text { \#. } 452 \mathrm{ACET}+\text { \#. } 11 \mathrm{MEK}+\text { \#. } 125 \\ & \text { PROD } ~+~ \#-0.105 ~ X C ~ \end{aligned}$ |
| a5OH | $9.34 \mathrm{e}-12$ | $1.11 \mathrm{e}-11$ | 0.10 |  | $\begin{aligned} & \text { ALK5 + HO. }=\text { \#. } 653 \text { RO2-R. + \#. } 347 \mathrm{RO} 2-\mathrm{N} .+ \text { \#. } 948 \mathrm{R} 2 \mathrm{O} 2 .+ \\ & \text { \#. } 026 \mathrm{HCHO}+\# .099 \mathrm{CCHO}+\# .204 \mathrm{RCHO}+\text { \#. } 072 \mathrm{ACET}+ \\ & \text { \#. } 089 \mathrm{MEK}+\text { \#. } 417 \text { PROD2 + \#2.008 XC } \end{aligned}$ |
| b1OH | $5.95 \mathrm{e}-12$ | 1.81e-12 | -0.71 |  | $\begin{aligned} & \text { ARO1 + HO. }=\text { \#. } 224 \text { HO2. + \#. } 765 \text { RO2-R. + \#. } 011 \text { RO2-N. + } \\ & \text { \#. } 055 \text { PROD } 2+\text { \#.118 GLY + \#.119 MGLY + \#. } 017 \text { PHEN + } \\ & \text { \#. } 207 \text { CRES + \#. } 059 \text { BALD + \#. } 491 \text { DCB1 + \#. } 108 \text { DCB2 + \#. } 051 \\ & \text { DCB3 + \#1.288 XC } \end{aligned}$ |
| b2OH | $2.64 \mathrm{e}-11$ | $2.64 \mathrm{e}-11$ | 0.00 |  | $\begin{aligned} & \mathrm{ARO} 2+\mathrm{HO} .=\# .187 \mathrm{HO} 2 .+ \text { \#. } 804 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 009 \mathrm{RO} 2-\mathrm{N} .+ \\ & \text { \#. } 097 \mathrm{GLY}+\text { \#. } 287 \mathrm{MGLY}+\text { \#. } 087 \mathrm{BACL}+\# .187 \mathrm{CRES}+\# .05 \\ & \text { BALD + \#. } 561 \mathrm{DCB} 1+\text { \#. } 099 \text { DCB2 + \#. } 093 \text { DCB3 + \#1. } 68 \text { XC } \end{aligned}$ |
| olOH | $3.23 \mathrm{e}-11$ | $7.10 \mathrm{e}-12$ | -0.90 |  | $\begin{aligned} & \text { OLE1 + HO. }=\text { \#. } 91 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 09 \mathrm{RO} 2-\mathrm{N} .+ \text { \#. } 205 \mathrm{R} 2 \mathrm{O} 2 .+ \\ & \text { \#. } 732 \mathrm{HCHO}+\text { \#. } 294 \mathrm{CCHO}+\text { \#. } 497 \mathrm{RCHO}+\text { \#. } 005 \mathrm{ACET}+ \\ & \text { \#. } 119 \text { PROD2 + \#. } 92 \mathrm{XC} \end{aligned}$ |
| o103 | 1.06e-17 | $2.62 \mathrm{e}-15$ | 3.26 |  | $\begin{aligned} & \text { OLE } 1+\mathrm{O} 3=\# .155 \mathrm{HO} .+ \text { \#. } 056 \mathrm{HO} 2 .+ \text { \#. } 022 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 001 \\ & \text { RO2-N. + \#. } 076 \mathrm{C}-\mathrm{O} 2 .+ \text { \#. } 345 \mathrm{CO}+\text { \#. } 086 \mathrm{CO} 2+\# .5 \mathrm{HCHO}+ \\ & \# .154 \mathrm{CCHO}+\text { \#. } 363 \mathrm{RCHO}+\# .001 \text { ACET + \#. } 215 \mathrm{PROD} 2+ \\ & \text { \#. } 185 \mathrm{HCOOH}+\text { \#. } 05 \mathrm{CCO}-\mathrm{OH}+\# .119 \mathrm{RCO}-\mathrm{OH}+\text { \#. } 654 \mathrm{XC} \end{aligned}$ |
| o1N3 | 1.26e-14 | $4.45 \mathrm{e}-14$ | 0.75 |  | $\begin{aligned} & \text { OLE1 + NO3 }=\text { \#. } 824 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 176 \mathrm{RO} 2-\mathrm{N} .+ \text { \#. } 488 \mathrm{R} 2 \mathrm{O} 2 .+ \\ & \# .009 \mathrm{CCHO}+\# .037 \mathrm{RCHO}+\# .024 \mathrm{ACET}+\text { \#. } 511 \mathrm{RNO} 3+ \\ & \text { \#. } 677 \mathrm{XC}+\# .489 \mathrm{XN} \end{aligned}$ |
| o1OP | $4.90 \mathrm{e}-12$ | $1.07 \mathrm{e}-11$ | 0.47 |  | $\begin{aligned} & \mathrm{OLE} 1+\mathrm{O} 3 \mathrm{P}=\# .45 \mathrm{RCHO}+\text { \#. } 437 \mathrm{MEK}+\text { \#. } 113 \text { PROD2 + } \\ & \text { \#1.224 XC } \end{aligned}$ |
| o2OH | $6.33 \mathrm{e}-11$ | $1.74 \mathrm{e}-11$ | -0.76 |  | $\begin{aligned} & \text { OLE2 + HO. }=\text { \#. } 918 \text { RO2-R. + \#. } 082 \text { RO2-N. + \#. } 001 \mathrm{R} 2 \mathrm{O} 2 .+ \\ & \# .244 \mathrm{HCHO}+\text { \#. } 732 \mathrm{CCHO}+\text { \#. } 511 \mathrm{RCHO}+\text { \#. } 127 \mathrm{ACET}+ \\ & \text { \#. } 072 \mathrm{MEK}+\text { \#. } 061 \mathrm{BALD}+\text { \#. } 025 \mathrm{METHACRO}+\text { \#. } 025 \mathrm{ISO}- \\ & \text { PROD + \#-. } 054 \mathrm{XC} \end{aligned}$ |
| o2O3 | 1.07e-16 | 5.02e-16 | 0.92 |  | OLE2 + O3 = \#. 378 HO. + \#. 003 HO2. + \#. 033 RO2-R. + \#. 002 RO2-N. + \#. 137 R2O2. + \#. 197 C-O2. + \#. 137 CCO-O2. + \#. 006 RCO-O2. + \#. $265 \mathrm{CO}+$ \#. $07 \mathrm{CO} 2+$ \#. $269 \mathrm{HCHO}+$ \#. 456 CCHO + \#. 305 RCHO + \#. 045 ACET + \#. 026 MEK + \#. 006 PROD2 + \#. 042 BALD + \#. 026 METHACRO + \#. 073 HCOOH + \#. 129 CCO-OH + \#. 303 RCO-OH + \#. 155 XC |
| o2N3 | 7.27e-13 | 7.27e-13 | 0.00 |  | $\begin{aligned} & \text { OLE } 2+\mathrm{NO} 3=\# .391 \mathrm{NO} 2+\text { \#. } 442 \mathrm{RO} 2-\mathrm{R} .+ \text { \#. } 136 \mathrm{RO} 2-\mathrm{N} .+ \\ & \text { \#. } 711 \mathrm{R} 2 \mathrm{O} 2 .+ \text { \#. } 03 \mathrm{C}-\mathrm{O} 2 .+ \text { \#. } 079 \mathrm{HCHO}+\text { \#. } 507 \mathrm{CCHO}+\text { \#. } 151 \\ & \mathrm{RCHO}+\text { \#. } 102 \text { ACET + \#. } 001 \mathrm{MEK}+\text { \#. } 015 \mathrm{BALD}+\# .048 \mathrm{MVK} \\ & + \text { \#. } 321 \mathrm{RNO} 3+\# .075 \mathrm{XC}+\text { \#. } 288 \mathrm{XN} \end{aligned}$ |
| o2OP | $2.09 \mathrm{e}-11$ | $2.09 \mathrm{e}-11$ |  |  | $\mathrm{OLE} 2+\mathrm{O} 3 \mathrm{P}=$ \#. $013 \mathrm{HO} 2 .+$ \#. $012 \mathrm{RO} 2-\mathrm{R} .+$ \#.001 RO2-N. + \#. 012 CO + \#. 069 RCHO + \#. 659 MEK + \#. 259 PROD2 + \#. 012 METHACRO + \#. 537 XC |

[a] Except as indicated, the rate constants are given by $\mathrm{k}(\mathrm{T})=\mathrm{A} \cdot(\mathrm{T} / 300)^{\mathrm{B}} \cdot \mathrm{e}^{-\mathrm{E} / \mathrm{RT}}$, where the units of k and A are $\mathrm{cm}^{3}$ molec $^{-1} \mathrm{~s}^{-1}$, Ea are $\mathrm{kcal} \mathrm{mol}{ }^{-1}, \mathrm{~T}$ is ${ }^{\circ} \mathrm{K}$, and $\mathrm{R}=0.0019872 \mathrm{kcal} \mathrm{mol}^{-1} \mathrm{deg}^{-1}$. The following special rate constant expressions are used:

## Table A-2 (continued)

Phot Set = name: The absorption cross sections and quantum yields for the photolysis reaction are given in Table A-5, where "name" indicates the photolysis set used. If a "qy=number" notation is given, the number given is the overall quantum yield, which is assumed to be wavelength independent.
Falloff: The rate constant as a function of temperature and pressure is calculated using $k(T, M)=\{k 0(T) \cdot[\mathrm{M}] /[1$ $+\mathrm{k} 0(\mathrm{~T}) \cdot[\mathrm{M}] / \operatorname{kinf}(\mathrm{T})]\} \cdot \mathrm{F}^{\mathrm{Z}}$, where $\left.\mathrm{Z}=\left\{1+\left[\log _{10}\{\mathrm{k} 0(\mathrm{~T}) \cdot[\mathrm{M}]) / \operatorname{kinf}(\mathrm{T})\right\}\right]^{2}\right\}^{-1},[\mathrm{M}]$ is the total pressure in molecules $\mathrm{cm}^{-3}, \mathrm{~F}$ is as indicated on the table, and the temperature dependences of k 0 and kinf are as indicated on the table.
(Slow): The reaction is assumed to be negligible and is not included in the mechanism. It is shown on the listing for documentation purposes only.
$\mathrm{k}=\mathrm{k} 0+\mathrm{k} 3 \mathrm{M}(1+\mathrm{k} 3 \mathrm{M} / \mathrm{k} 2)$ : The rate constant as a function of temperature and pressure is calculated using $k(T, M)=k 0(T)+k 3(T) \cdot[M] \cdot(1+k 3(T) \cdot[M] / k 2(T))$, where $[M]$ is the total bath gas (air) concentration in molecules $\mathrm{cm}^{-3}$, and the temperature dependences for $\mathrm{k} 0, \mathrm{k} 2$ and k 3 are as indicated on the table.
$\mathrm{k}=\mathrm{k} 1+\mathrm{k} 2[\mathrm{M}]$ : The rate constant as a function of temperature and pressure is calculated using $\mathrm{k}(\mathrm{T}, \mathrm{M})=\mathrm{k} 1(\mathrm{~T})+\mathrm{k} 2(\mathrm{~T}) \cdot[\mathrm{M}]$, where $[\mathrm{M}]$ is the total bath gas (air) concentration in molecules $\mathrm{cm}^{-3}$, and the temperature dependences for k 1 , and k 2 are as indicated on the table.
Same k as Rxn label: The rate constant is the same as the reaction with the indicated label.
[b] Format of reaction listing: " $=$ " separates reactants from products; "\#number" indicates stoichiometric coefficient, "\#coefficient $\{$ product list \}" means that the stoichiometric coefficient is applied to all the products listed. See Table A-1 for a listing of the model species used.

Table A-3. Listing of the absorption cross sections and quantum yields for the photolysis reactions.

| $\begin{gathered} \mathrm{WL} \\ (\mathrm{~nm}) \end{gathered}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{aligned} & \mathrm{WL} \\ & (\mathrm{~nm}) \end{aligned}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{gathered} \text { WL } \\ (\mathrm{nm}) \end{gathered}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NO2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 205.0 | 4.31e-19 | 1.000 | 210.0 | 4.72e-19 | 1.000 | 215.0 | $4.95 \mathrm{e}-19$ | 1.000 | 220.0 | 4.56e-19 | 1.000 | 225.0 | $3.79 \mathrm{e}-19$ | 1.000 |
| 230.0 | 2.74e-19 | 1.000 | 235.0 | 1.67e-19 | 1.000 | 240.0 | $9.31 \mathrm{e}-20$ | 1.000 | 245.0 | $4.74 \mathrm{e}-20$ | 1.000 | 250.0 | $2.48 \mathrm{e}-20$ | 1.000 |
| 255.0 | 1.95e-20 | 1.000 | 260.0 | $2.24 \mathrm{e}-20$ | 1.000 | 265.0 | $2.73 \mathrm{e}-20$ | 1.000 | 270.0 | $4.11 \mathrm{e}-20$ | 1.000 | 275.0 | $4.90 \mathrm{e}-20$ | 1.000 |
| 280.0 | 5.92e-20 | 1.000 | 285.0 | $7.39 \mathrm{e}-20$ | 1.000 | 290.0 | $9.00 \mathrm{e}-20$ | 1.000 | 295.0 | $1.09 \mathrm{e}-19$ | 1.000 | 300.0 | $1.31 \mathrm{e}-19$ | 1.000 |
| 305.0 | 1.57e-19 | 1.000 | 310.0 | 1.86e-19 | 1.000 | 315.0 | 2.15e-19 | 0.990 | 320.0 | 2.48e-19 | 0.990 | 325.0 | $2.81 \mathrm{e}-19$ | 0.990 |
| 330.0 | $3.13 \mathrm{e}-19$ | 0.990 | 335.0 | $3.43 \mathrm{e}-19$ | 0.990 | 340.0 | $3.80 \mathrm{e}-19$ | 0.990 | 345.0 | $4.07 \mathrm{e}-19$ | 0.990 | 350.0 | $4.31 \mathrm{e}-19$ | 0.990 |
| 355.0 | 4.72e-19 | 0.990 | 360.0 | $4.83 \mathrm{e}-19$ | 0.980 | 365.0 | 5.17e-19 | 0.980 | 370.0 | $5.32 \mathrm{e}-19$ | 0.980 | 375.0 | $5.51 \mathrm{e}-19$ | 0.980 |
| 380.0 | 5.64e-19 | 0.970 | 385.0 | 5.76e-19 | 0.970 | 390.0 | 5.93e-19 | 0.960 | 395.0 | $5.85 \mathrm{e}-19$ | 0.935 | 400.0 | 6.02e-19 | 0.820 |
| 405.0 | 5.78e-19 | 0.355 | 410.0 | 6.00e-19 | 0.130 | 411.0 | 5.93e-19 | 0.110 | 412.0 | 5.86e-19 | 0.094 | 413.0 | $5.79 \mathrm{e}-19$ | 0.083 |
| 414.0 | 5.72e-19 | 0.070 | 415.0 | 5.65e-19 | 0.059 | 416.0 | $5.68 \mathrm{e}-19$ | 0.048 | 417.0 | $5.71 \mathrm{e}-19$ | 0.039 | 418.0 | $5.75 \mathrm{e}-19$ | 0.030 |
| 419.0 | 5.78e-19 | 0.023 | 420.0 | 5.81e-19 | 0.018 | 421.0 | $5.72 \mathrm{e}-19$ | 0.012 | 422.0 | 5.64e-19 | 0.008 | 423.0 | $5.55 \mathrm{e}-19$ | 0.004 |
| 424.0 | 5.47e-19 | 0.000 |  |  |  |  |  |  |  |  |  |  |  |  |
| NO3NO |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 585.0 | $2.89 \mathrm{e}-18$ | 0.000 | 586.0 | 3.32e-18 | 0.050 | 587.0 | 4.16e-18 | 0.100 | 588.0 | 5.04e-18 | 0.150 | 589.0 | $6.13 \mathrm{e}-18$ | 0.200 |
| 590.0 | 5.96e-18 | 0.250 | 591.0 | 5.44e-18 | 0.280 | 592.0 | 5.11e-18 | 0.310 | 593.0 | $4.58 \mathrm{e}-18$ | 0.340 | 594.0 | $4.19 \mathrm{e}-18$ | 0.370 |
| 595.0 | $4.29 \mathrm{e}-18$ | 0.400 | 596.0 | $4.62 \mathrm{e}-18$ | 0.370 | 597.0 | $4.36 \mathrm{e}-18$ | 0.340 | 598.0 | $3.67 \mathrm{e}-18$ | 0.310 | 599.0 | $3.10 \mathrm{e}-18$ | 0.280 |
| 600.0 | 2.76e-18 | 0.250 | 601.0 | 2.86e-18 | 0.240 | 602.0 | $3.32 \mathrm{e}-18$ | 0.230 | 603.0 | $3.80 \mathrm{e}-18$ | 0.220 | 604.0 | $4.37 \mathrm{e}-18$ | 0.210 |
| 605.0 | 4.36e-18 | 0.200 | 606.0 | 3.32e-18 | 0.200 | 607.0 | $2.40 \mathrm{e}-18$ | 0.200 | 608.0 | $1.85 \mathrm{e}-18$ | 0.200 | 609.0 | $1.71 \mathrm{e}-18$ | 0.200 |
| 610.0 | 1.77e-18 | 0.200 | 611.0 | 1.91e-18 | 0.180 | 612.0 | $2.23 \mathrm{e}-18$ | 0.160 | 613.0 | $2.63 \mathrm{e}-18$ | 0.140 | 614.0 | $2.55 \mathrm{e}-18$ | 0.120 |
| 615.0 | 2.26e-18 | 0.100 | 616.0 | $2.09 \mathrm{e}-18$ | 0.100 | 617.0 | $2.11 \mathrm{e}-18$ | 0.100 | 618.0 | $2.39 \mathrm{e}-18$ | 0.100 | 619.0 | $2.56 \mathrm{e}-18$ | 0.100 |
| 620.0 | 3.27e-18 | 0.100 | 621.0 | 5.24e-18 | 0.090 | 622.0 | $1.02 \mathrm{e}-17$ | 0.080 | 623.0 | 1.47e-17 | 0.070 | 624.0 | $1.21 \mathrm{e}-17$ | 0.060 |
| 625.0 | 8.38e-18 | 0.050 | 626.0 | 7.30e-18 | 0.050 | 627.0 | 7.53e-18 | 0.050 | 628.0 | $7.37 \mathrm{e}-18$ | 0.050 | 629.0 | $6.98 \mathrm{e}-18$ | 0.050 |
| 630.0 | 6.76e-18 | 0.050 | 631.0 | 4.84e-18 | 0.046 | 632.0 | $3.27 \mathrm{e}-18$ | 0.042 | 633.0 | $2.17 \mathrm{e}-18$ | 0.038 | 634.0 | $1.64 \mathrm{e}-18$ | 0.034 |
| 635.0 | 1.44e-18 | 0.030 | 636.0 | $1.69 \mathrm{e}-18$ | 0.024 | 637.0 | $2.07 \mathrm{e}-18$ | 0.018 | 638.0 | $2.03 \mathrm{e}-18$ | 0.012 | 639.0 | $1.58 \mathrm{e}-18$ | 0.006 |
| 640.0 | $1.23 \mathrm{e}-18$ | 0.000 |  |  |  |  |  |  |  |  |  |  |  |  |
| NO3NO2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 400.0 | $0.00 \mathrm{e}+00$ | 1.000 | 401.0 | $0.00 \mathrm{e}+00$ | 1.000 | 402.0 | 0.00e+00 | 1.000 | 403.0 | $2.00 \mathrm{e}-20$ | 1.000 | 404.0 | $0.00 \mathrm{e}+00$ | 1.000 |
| 405.0 | $3.00 \mathrm{e}-20$ | 1.000 | 406.0 | $2.00 \mathrm{e}-20$ | 1.000 | 407.0 | $1.00 \mathrm{e}-20$ | 1.000 | 408.0 | $3.00 \mathrm{e}-20$ | 1.000 | 409.0 | $0.00 \mathrm{e}+00$ | 1.000 |
| 410.0 | 1.00e-20 | 1.000 | 411.0 | $2.00 \mathrm{e}-20$ | 1.000 | 412.0 | $5.00 \mathrm{e}-20$ | 1.000 | 413.0 | $5.00 \mathrm{e}-20$ | 1.000 | 414.0 | $2.00 \mathrm{e}-20$ | 1.000 |
| 415.0 | $6.00 \mathrm{e}-20$ | 1.000 | 416.0 | $6.00 \mathrm{e}-20$ | 1.000 | 417.0 | $7.00 \mathrm{e}-20$ | 1.000 | 418.0 | $5.00 \mathrm{e}-20$ | 1.000 | 419.0 | $8.00 \mathrm{e}-20$ | 1.000 |
| 420.0 | $8.00 \mathrm{e}-20$ | 1.000 | 421.0 | 8.00e-20 | 1.000 | 422.0 | $9.00 \mathrm{e}-20$ | 1.000 | 423.0 | 1.10e-19 | 1.000 | 424.0 | $9.00 \mathrm{e}-20$ | 1.000 |
| 425.0 | $7.00 \mathrm{e}-20$ | 1.000 | 426.0 | 1.40e-19 | 1.000 | 427.0 | 1.40e-19 | 1.000 | 428.0 | 1.20e-19 | 1.000 | 429.0 | $1.10 \mathrm{e}-19$ | 1.000 |
| 430.0 | 1.70e-19 | 1.000 | 431.0 | 1.30e-19 | 1.000 | 432.0 | $1.50 \mathrm{e}-19$ | 1.000 | 433.0 | $1.80 \mathrm{e}-19$ | 1.000 | 434.0 | $1.80 \mathrm{e}-19$ | 1.000 |
| 435.0 | 1.60e-19 | 1.000 | 436.0 | 1.50e-19 | 1.000 | 437.0 | $1.80 \mathrm{e}-19$ | 1.000 | 438.0 | 2.10e-19 | 1.000 | 439.0 | $2.00 \mathrm{e}-19$ | 1.000 |
| 440.0 | 1.90e-19 | 1.000 | 441.0 | 1.80e-19 | 1.000 | 442.0 | $2.10 \mathrm{e}-19$ | 1.000 | 443.0 | 1.80e-19 | 1.000 | 444.0 | $1.90 \mathrm{e}-19$ | 1.000 |
| 445.0 | 2.00e-19 | 1.000 | 446.0 | 2.40e-19 | 1.000 | 447.0 | 2.90 e-19 | 1.000 | 448.0 | 2.40e-19 | 1.000 | 449.0 | $2.80 \mathrm{e}-19$ | 1.000 |
| 450.0 | 2.90e-19 | 1.000 | 451.0 | $3.00 \mathrm{e}-19$ | 1.000 | 452.0 | $3.30 \mathrm{e}-19$ | 1.000 | 453.0 | $3.10 \mathrm{e}-19$ | 1.000 | 454.0 | $3.60 \mathrm{e}-19$ | 1.000 |
| 455.0 | 3.60e-19 | 1.000 | 456.0 | 3.60e-19 | 1.000 | 457.0 | $4.00 \mathrm{e}-19$ | 1.000 | 458.0 | $3.70 \mathrm{e}-19$ | 1.000 | 459.0 | $4.20 \mathrm{e}-19$ | 1.000 |
| 460.0 | 4.00e-19 | 1.000 | 461.0 | 3.90e-19 | 1.000 | 462.0 | $4.00 \mathrm{e}-19$ | 1.000 | 463.0 | $4.10 \mathrm{e}-19$ | 1.000 | 464.0 | $4.80 \mathrm{e}-19$ | 1.000 |
| 465.0 | 5.10e-19 | 1.000 | 466.0 | $5.40 \mathrm{e}-19$ | 1.000 | 467.0 | $5.70 \mathrm{e}-19$ | 1.000 | 468.0 | $5.60 \mathrm{e}-19$ | 1.000 | 469.0 | $5.80 \mathrm{e}-19$ | 1.000 |
| 470.0 | 5.90e-19 | 1.000 | 471.0 | 6.20e-19 | 1.000 | 472.0 | $6.40 \mathrm{e}-19$ | 1.000 | 473.0 | $6.20 \mathrm{e}-19$ | 1.000 | 474.0 | $6.20 \mathrm{e}-19$ | 1.000 |
| 475.0 | 6.80e-19 | 1.000 | 476.0 | 7.80e-19 | 1.000 | 477.0 | $7.70 \mathrm{e}-19$ | 1.000 | 478.0 | 7.30e-19 | 1.000 | 479.0 | $7.30 \mathrm{e}-19$ | 1.000 |
| 480.0 | $7.00 \mathrm{e}-19$ | 1.000 | 481.0 | 7.10e-19 | 1.000 | 482.0 | $7.10 \mathrm{e}-19$ | 1.000 | 483.0 | $7.20 \mathrm{e}-19$ | 1.000 | 484.0 | $7.70 \mathrm{e}-19$ | 1.000 |
| 485.0 | 8.20e-19 | 1.000 | 486.0 | 9.10e-19 | 1.000 | 487.0 | $9.20 \mathrm{e}-19$ | 1.000 | 488.0 | $9.50 \mathrm{e}-19$ | 1.000 | 489.0 | $9.60 \mathrm{e}-19$ | 1.000 |
| 490.0 | $1.03 \mathrm{e}-18$ | 1.000 | 491.0 | 9.90e-19 | 1.000 | 492.0 | $9.90 \mathrm{e}-19$ | 1.000 | 493.0 | $1.01 \mathrm{e}-18$ | 1.000 | 494.0 | $1.01 \mathrm{e}-18$ | 1.000 |
| 495.0 | 1.06e-18 | 1.000 | 496.0 | 1.21e-18 | 1.000 | 497.0 | $1.22 \mathrm{e}-18$ | 1.000 | 498.0 | $1.20 \mathrm{e}-18$ | 1.000 | 499.0 | $1.17 \mathrm{e}-18$ | 1.000 |
| 500.0 | $1.13 \mathrm{e}-18$ | 1.000 | 501.0 | $1.11 \mathrm{e}-18$ | 1.000 | 502.0 | $1.11 \mathrm{e}-18$ | 1.000 | 503.0 | $1.11 \mathrm{e}-18$ | 1.000 | 504.0 | $1.26 \mathrm{e}-18$ | 1.000 |
| 505.0 | 1.28e-18 | 1.000 | 506.0 | $1.34 \mathrm{e}-18$ | 1.000 | 507.0 | $1.28 \mathrm{e}-18$ | 1.000 | 508.0 | $1.27 \mathrm{e}-18$ | 1.000 | 509.0 | $1.35 \mathrm{e}-18$ | 1.000 |
| 510.0 | 1.51e-18 | 1.000 | 511.0 | $1.73 \mathrm{e}-18$ | 1.000 | 512.0 | $1.77 \mathrm{e}-18$ | 1.000 | 513.0 | $1.60 \mathrm{e}-18$ | 1.000 | 514.0 | $1.58 \mathrm{e}-18$ | 1.000 |
| 515.0 | $1.58 \mathrm{e}-18$ | 1.000 | 516.0 | 1.56e-18 | 1.000 | 517.0 | $1.49 \mathrm{e}-18$ | 1.000 | 518.0 | $1.44 \mathrm{e}-18$ | 1.000 | 519.0 | $1.54 \mathrm{e}-18$ | 1.000 |
| 520.0 | $1.68 \mathrm{e}-18$ | 1.000 | 521.0 | $1.83 \mathrm{e}-18$ | 1.000 | 522.0 | $1.93 \mathrm{e}-18$ | 1.000 | 523.0 | $1.77 \mathrm{e}-18$ | 1.000 | 524.0 | $1.64 \mathrm{e}-18$ | 1.000 |
| 525.0 | $1.58 \mathrm{e}-18$ | 1.000 | 526.0 | $1.63 \mathrm{e}-18$ | 1.000 | 527.0 | $1.81 \mathrm{e}-18$ | 1.000 | 528.0 | 2.10e-18 | 1.000 | 529.0 | $2.39 \mathrm{e}-18$ | 1.000 |
| 530.0 | $2.23 \mathrm{e}-18$ | 1.000 | 531.0 | $2.09 \mathrm{e}-18$ | 1.000 | 532.0 | $2.02 \mathrm{e}-18$ | 1.000 | 533.0 | $1.95 \mathrm{e}-18$ | 1.000 | 534.0 | $2.04 \mathrm{e}-18$ | 1.000 |
| 535.0 | 2.30e-18 | 1.000 | 536.0 | 2.57e-18 | 1.000 | 537.0 | $2.58 \mathrm{e}-18$ | 1.000 | 538.0 | $2.34 \mathrm{e}-18$ | 1.000 | 539.0 | $2.04 \mathrm{e}-18$ | 1.000 |
| 540.0 | 2.10e-18 | 1.000 | 541.0 | 2.04e-18 | 1.000 | 542.0 | $1.88 \mathrm{e}-18$ | 1.000 | 543.0 | 1.68e-18 | 1.000 | 544.0 | $1.70 \mathrm{e}-18$ | 1.000 |
| 545.0 | 1.96e-18 | 1.000 | 546.0 | $2.42 \mathrm{e}-18$ | 1.000 | 547.0 | $2.91 \mathrm{e}-18$ | 1.000 | 548.0 | $2.98 \mathrm{e}-18$ | 1.000 | 549.0 | $2.71 \mathrm{e}-18$ | 1.000 |
| 550.0 | 2.48e-18 | 1.000 | 551.0 | $2.43 \mathrm{e}-18$ | 1.000 | 552.0 | $2.47 \mathrm{e}-18$ | 1.000 | 553.0 | $2.53 \mathrm{e}-18$ | 1.000 | 554.0 | $2.78 \mathrm{e}-18$ | 1.000 |
| 555.0 | $3.11 \mathrm{e}-18$ | 1.000 | 556.0 | 3.26e-18 | 1.000 | 557.0 | $3.29 \mathrm{e}-18$ | 1.000 | 558.0 | $3.51 \mathrm{e}-18$ | 1.000 | 559.0 | $3.72 \mathrm{e}-18$ | 1.000 |
| 560.0 | $3.32 \mathrm{e}-18$ | 1.000 | 561.0 | 2.98e-18 | 1.000 | 562.0 | 2.90 e-18 | 1.000 | 563.0 | $2.80 \mathrm{e}-18$ | 1.000 | 564.0 | $2.72 \mathrm{e}-18$ | 1.000 |
| 565.0 | $2.73 \mathrm{e}-18$ | 1.000 | 566.0 | $2.85 \mathrm{e}-18$ | 1.000 | 567.0 | $2.81 \mathrm{e}-18$ | 1.000 | 568.0 | $2.85 \mathrm{e}-18$ | 1.000 | 569.0 | $2.89 \mathrm{e}-18$ | 1.000 |
| 570.0 | 2.79e-18 | 1.000 | 571.0 | 2.76e-18 | 1.000 | 572.0 | $2.74 \mathrm{e}-18$ | 1.000 | 573.0 | $2.78 \mathrm{e}-18$ | 1.000 | 574.0 | $2.86 \mathrm{e}-18$ | 1.000 |
| 575.0 | 3.08e-18 | 1.000 | 576.0 | $3.27 \mathrm{e}-18$ | 1.000 | 577.0 | $3.38 \mathrm{e}-18$ | 1.000 | 578.0 | $3.31 \mathrm{e}-18$ | 1.000 | 579.0 | $3.24 \mathrm{e}-18$ | 1.000 |
| 580.0 | $3.34 \mathrm{e}-18$ | 1.000 | 581.0 | $3.55 \mathrm{e}-18$ | 1.000 | 582.0 | $3.28 \mathrm{e}-18$ | 1.000 | 583.0 | $2.93 \mathrm{e}-18$ | 1.000 | 584.0 | $2.82 \mathrm{e}-18$ | 1.000 |
| 585.0 | $2.89 \mathrm{e}-18$ | 1.000 | 586.0 | $3.32 \mathrm{e}-18$ | 0.950 | 587.0 | 4.16e-18 | 0.900 | 588.0 | $5.04 \mathrm{e}-18$ | 0.850 | 589.0 | $6.13 \mathrm{e}-18$ | 0.800 |
| 590.0 | 5.96e-18 | 0.750 | 591.0 | $5.44 \mathrm{e}-18$ | 0.720 | 592.0 | $5.11 \mathrm{e}-18$ | 0.690 | 593.0 | $4.58 \mathrm{e}-18$ | 0.660 | 594.0 | $4.19 \mathrm{e}-18$ | 0.630 |
| 595.0 | $4.29 \mathrm{e}-18$ | 0.600 | 596.0 | $4.62 \mathrm{e}-18$ | 0.590 | 597.0 | $4.36 \mathrm{e}-18$ | 0.580 | 598.0 | 3.67e-18 | 0.570 | 599.0 | $3.10 \mathrm{e}-18$ | 0.560 |
| 600.0 | 2.76e-18 | 0.550 | 601.0 | 2.86e-18 | 0.540 | 602.0 | $3.32 \mathrm{e}-18$ | 0.530 | 603.0 | $3.80 \mathrm{e}-18$ | 0.520 | 604.0 | $4.37 \mathrm{e}-18$ | 0.510 |
| 605.0 | 4.36e-18 | 0.400 | 606.0 | 3.32e-18 | 0.380 | 607.0 | $2.40 \mathrm{e}-18$ | 0.360 | 608.0 | $1.85 \mathrm{e}-18$ | 0.340 | 609.0 | $1.71 \mathrm{e}-18$ | 0.320 |

Table A-3 (continued)

| $\begin{aligned} & \mathrm{WL} \\ & (\mathrm{~nm}) \end{aligned}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{gathered} \mathrm{WL} \\ (\mathrm{~nm}) \end{gathered}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{gathered} \mathrm{WL} \\ (\mathrm{~nm}) \end{gathered}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{gathered} \mathrm{WL} \\ (\mathrm{~nm}) \end{gathered}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 610.0 | $1.77 \mathrm{e}-18$ | 0.300 | 611.0 | $1.91 \mathrm{e}-18$ | 0.290 | 612.0 | $2.23 \mathrm{e}-18$ | 0.280 | 613.0 | $2.63 \mathrm{e}-18$ | 0.270 | 614.0 | 2.55e-18 | 0.260 |
| 615.0 | 2.26e-18 | 0.250 | 616.0 | $2.09 \mathrm{e}-18$ | 0.240 | 617.0 | $2.11 \mathrm{e}-18$ | 0.230 | 618.0 | $2.39 \mathrm{e}-18$ | 0.220 | 619.0 | 2.56e-18 | 0.210 |
| 620.0 | 3.27e-18 | 0.200 | 621.0 | $5.24 \mathrm{e}-18$ | 0.190 | 622.0 | $1.02 \mathrm{e}-17$ | 0.180 | 623.0 | $1.47 \mathrm{e}-17$ | 0.170 | 624.0 | 1.21e-17 | 0.160 |
| 625.0 | $8.38 \mathrm{e}-18$ | 0.150 | 626.0 | $7.30 \mathrm{e}-18$ | 0.130 | 627.0 | $7.53 \mathrm{e}-18$ | 0.110 | 628.0 | 7.37e-18 | 0.090 | 629.0 | 6.98e-18 | 0.070 |
| 630.0 | $6.76 \mathrm{e}-18$ | 0.050 | 631.0 | $4.84 \mathrm{e}-18$ | 0.040 | 632.0 | $3.27 \mathrm{e}-18$ | 0.030 | 633.0 | 2.17e-18 | 0.020 | 634.0 | $1.64 \mathrm{e}-18$ | 0.010 |
| 635.0 | $1.44 \mathrm{e}-18$ | 0.000 |  |  |  |  |  |  |  |  |  |  |  |  |
| O3O3P |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 280.0 | $3.94 \mathrm{e}-18$ | 0.095 | 281.0 | $3.62 \mathrm{e}-18$ | 0.093 | 282.0 | $3.31 \mathrm{e}-18$ | 0.090 | 283.0 | 2.99e-18 | 0.088 | 284.0 | 2.70e-18 | 0.086 |
| 285.0 | $2.46 \mathrm{e}-18$ | 0.084 | 286.0 | $2.22 \mathrm{e}-18$ | 0.082 | 287.0 | $1.98 \mathrm{e}-18$ | 0.079 | 288.0 | 1.75e-18 | 0.077 | 289.0 | 1.59e-18 | 0.075 |
| 290.0 | $1.42 \mathrm{e}-18$ | 0.073 | 291.0 | $1.25 \mathrm{e}-18$ | 0.070 | 292.0 | $1.09 \mathrm{e}-18$ | 0.068 | 293.0 | 9.81e-19 | 0.066 | 294.0 | $8.73 \mathrm{e}-19$ | 0.064 |
| 295.0 | $7.65 \mathrm{e}-19$ | 0.061 | 296.0 | $6.58 \mathrm{e}-19$ | 0.059 | 297.0 | 5.81e-19 | 0.057 | 298.0 | 5.18e-19 | 0.055 | 299.0 | 4.55e-19 | 0.052 |
| 300.0 | 3.92e-19 | 0.050 | 301.0 | $3.35 \mathrm{e}-19$ | 0.035 | 302.0 | $3.01 \mathrm{e}-19$ | 0.025 | 303.0 | 2.66e-19 | 0.015 | 304.0 | 2.32e-19 | 0.010 |
| 305.0 | $1.97 \mathrm{e}-19$ | 0.020 | 306.0 | $1.73 \mathrm{e}-19$ | 0.050 | 307.0 | $1.55 \mathrm{e}-19$ | 0.123 | 308.0 | 1.37e-19 | 0.227 | 309.0 | 1.18e-19 | 0.333 |
| 310.0 | $9.98 \mathrm{e}-20$ | 0.400 | 311.0 | $8.92 \mathrm{e}-20$ | 0.612 | 312.0 | $7.94 \mathrm{e}-20$ | 0.697 | 313.0 | 6.96e-20 | 0.738 | 314.0 | 5.99e-20 | 0.762 |
| 315.0 | 5.01e-20 | 0.765 | 316.0 | 4.51e-20 | 0.779 | 317.0 | $4.00 \mathrm{e}-20$ | 0.791 | 318.0 | $3.50 \mathrm{e}-20$ | 0.806 | 319.0 | $2.99 \mathrm{e}-20$ | 0.822 |
| 320.0 | $2.49 \mathrm{e}-20$ | 0.852 | 321.0 | $2.23 \mathrm{e}-20$ | 0.879 | 322.0 | $1.97 \mathrm{e}-20$ | 0.903 | 323.0 | $1.72 \mathrm{e}-20$ | 0.908 | 324.0 | 1.46e-20 | 0.920 |
| 325.0 | $1.20 \mathrm{e}-20$ | 0.930 | 326.0 | $1.08 \mathrm{e}-20$ | 0.934 | 327.0 | $9.67 \mathrm{e}-21$ | 0.938 | 328.0 | $8.50 \mathrm{e}-21$ | 0.942 | 329.0 | 7.34e-21 | 0.946 |
| 330.0 | 6.17e-21 | 0.950 | 331.0 | 5.48e-21 | 0.950 | 332.0 | $4.80 \mathrm{e}-21$ | 0.950 | 333.0 | 4.11e-21 | 0.950 | 334.0 | $3.43 \mathrm{e}-21$ | 0.950 |
| 335.0 | 2.74e-21 | 0.950 | 336.0 | $2.43 \mathrm{e}-21$ | 0.960 | 337.0 | $2.11 \mathrm{e}-21$ | 0.970 | 338.0 | $1.80 \mathrm{e}-21$ | 0.980 | 339.0 | 1.48e-21 | 0.990 |
| 340.0 | $1.17 \mathrm{e}-21$ | 1.000 | 350.0 | $0.00 \mathrm{e}+00$ | 1.000 | 400.0 | $0.00 \mathrm{e}+00$ | 1.000 | 410.0 | $1.20 \mathrm{e}-23$ | 1.000 | 420.0 | 2.20e-23 | 1.000 |
| 440.0 | $1.12 \mathrm{e}-22$ | 1.000 | 460.0 | 3.28e-22 | 1.000 | 480.0 | $6.84 \mathrm{e}-22$ | 1.000 | 500.0 | $1.22 \mathrm{e}-21$ | 1.000 | 520.0 | 1.82e-21 | 1.000 |
| 540.0 | $2.91 \mathrm{e}-21$ | 1.000 | 560.0 | $3.94 \mathrm{e}-21$ | 1.000 | 580.0 | $4.59 \mathrm{e}-21$ | 1.000 | 600.0 | 5.11e-21 | 1.000 | 620.0 | $4.00 \mathrm{e}-21$ | 1.000 |
| 640.0 | $2.96 \mathrm{e}-21$ | 1.000 | 660.0 | $2.09 \mathrm{e}-21$ | 1.000 | 680.0 | $1.36 \mathrm{e}-21$ | 1.000 | 700.0 | $9.10 \mathrm{e}-22$ | 1.000 | 750.0 | $3.20 \mathrm{e}-22$ | 1.000 |
| 800.0 | $1.60 \mathrm{e}-22$ | 1.000 | 900.0 | $0.00 \mathrm{e}+00$ | 1.000 |  |  |  |  |  |  |  |  |  |
| O301D |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 280.0 | $3.94 \mathrm{e}-18$ | 0.905 | 281.0 | $3.62 \mathrm{e}-18$ | 0.907 | 282.0 | $3.31 \mathrm{e}-18$ | 0.910 | 283.0 | 2.99e-18 | 0.912 | 284.0 | 2.70e-18 | 0.914 |
| 285.0 | 2.46e-18 | 0.916 | 286.0 | $2.22 \mathrm{e}-18$ | 0.918 | 287.0 | $1.98 \mathrm{e}-18$ | 0.921 | 288.0 | 1.75e-18 | 0.923 | 289.0 | $1.59 \mathrm{e}-18$ | 0.925 |
| 290.0 | $1.42 \mathrm{e}-18$ | 0.927 | 291.0 | $1.25 \mathrm{e}-18$ | 0.930 | 292.0 | $1.09 \mathrm{e}-18$ | 0.932 | 293.0 | 9.81e-19 | 0.934 | 294.0 | $8.73 \mathrm{e}-19$ | 0.936 |
| 295.0 | $7.65 \mathrm{e}-19$ | 0.939 | 296.0 | $6.58 \mathrm{e}-19$ | 0.941 | 297.0 | 5.81e-19 | 0.943 | 298.0 | 5.18e-19 | 0.945 | 299.0 | $4.55 \mathrm{e}-19$ | 0.948 |
| 300.0 | 3.92e-19 | 0.950 | 301.0 | $3.35 \mathrm{e}-19$ | 0.965 | 302.0 | $3.01 \mathrm{e}-19$ | 0.975 | 303.0 | 2.66e-19 | 0.985 | 304.0 | 2.32e-19 | 0.990 |
| 305.0 | $1.97 \mathrm{e}-19$ | 0.980 | 306.0 | $1.73 \mathrm{e}-19$ | 0.950 | 307.0 | $1.55 \mathrm{e}-19$ | 0.877 | 308.0 | 1.37e-19 | 0.773 | 309.0 | 1.18e-19 | 0.667 |
| 310.0 | 9.98e-20 | 0.600 | 311.0 | $8.92 \mathrm{e}-20$ | 0.388 | 312.0 | $7.94 \mathrm{e}-20$ | 0.303 | 313.0 | 6.96e-20 | 0.262 | 314.0 | 5.99e-20 | 0.238 |
| 315.0 | 5.01e-20 | 0.235 | 316.0 | $4.51 \mathrm{e}-20$ | 0.221 | 317.0 | $4.00 \mathrm{e}-20$ | 0.209 | 318.0 | $3.50 \mathrm{e}-20$ | 0.194 | 319.0 | $2.99 \mathrm{e}-20$ | 0.178 |
| 320.0 | $2.49 \mathrm{e}-20$ | 0.148 | 321.0 | $2.23 \mathrm{e}-20$ | 0.121 | 322.0 | $1.97 \mathrm{e}-20$ | 0.097 | 323.0 | 1.72e-20 | 0.092 | 324.0 | 1.46e-20 | 0.080 |
| 325.0 | $1.20 \mathrm{e}-20$ | 0.070 | 326.0 | $1.08 \mathrm{e}-20$ | 0.066 | 327.0 | $9.67 \mathrm{e}-21$ | 0.062 | 328.0 | $8.50 \mathrm{e}-21$ | 0.058 | 329.0 | $7.34 \mathrm{e}-21$ | 0.054 |
| 330.0 | 6.17e-21 | 0.050 | 331.0 | 5.48e-21 | 0.050 | 332.0 | $4.80 \mathrm{e}-21$ | 0.050 | 333.0 | 4.11e-21 | 0.050 | 334.0 | $3.43 \mathrm{e}-21$ | 0.050 |
| 335.0 | 2.74e-21 | 0.050 | 336.0 | $2.43 \mathrm{e}-21$ | 0.040 | 337.0 | $2.11 \mathrm{e}-21$ | 0.030 | 338.0 | $1.80 \mathrm{e}-21$ | 0.020 | 339.0 | 1.48e-21 | 0.010 |
| 340.0 | $1.17 \mathrm{e}-21$ | 0.000 |  |  |  |  |  |  |  |  |  |  |  |  |
| HONO-NO |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 309.0 | 0.00e+00 | 0.410 | 310.0 | $1.30 \mathrm{e}-20$ | 0.410 | 311.0 | $1.90 \mathrm{e}-20$ | 0.411 | 312.0 | $2.80 \mathrm{e}-20$ | 0.421 | 313.0 | 2.20e-20 | 0.432 |
| 314.0 | $3.60 \mathrm{e}-20$ | 0.443 | 315.0 | $3.00 \mathrm{e}-20$ | 0.454 | 316.0 | $1.40 \mathrm{e}-20$ | 0.464 | 317.0 | $3.10 \mathrm{e}-20$ | 0.475 | 318.0 | $5.60 \mathrm{e}-20$ | 0.486 |
| 319.0 | $3.60 \mathrm{e}-20$ | 0.496 | 320.0 | 4.90e-20 | 0.507 | 321.0 | $7.80 \mathrm{e}-20$ | 0.518 | 322.0 | $4.90 \mathrm{e}-20$ | 0.529 | 323.0 | $5.10 \mathrm{e}-20$ | 0.539 |
| 324.0 | $7.10 \mathrm{e}-20$ | 0.550 | 325.0 | $5.00 \mathrm{e}-20$ | 0.561 | 326.0 | $2.90 \mathrm{e}-20$ | 0.571 | 327.0 | $6.60 \mathrm{e}-20$ | 0.582 | 328.0 | 1.17e-19 | 0.593 |
| 329.0 | $6.10 \mathrm{e}-20$ | 0.604 | 330.0 | $1.11 \mathrm{e}-19$ | 0.614 | 331.0 | $1.79 \mathrm{e}-19$ | 0.625 | 332.0 | $8.70 \mathrm{e}-20$ | 0.636 | 333.0 | 7.60e-20 | 0.646 |
| 334.0 | $9.60 \mathrm{e}-20$ | 0.657 | 335.0 | $9.60 \mathrm{e}-20$ | 0.668 | 336.0 | $7.20 \mathrm{e}-20$ | 0.679 | 337.0 | $5.30 \mathrm{e}-20$ | 0.689 | 338.0 | 1.00e-19 | 0.700 |
| 339.0 | $1.88 \mathrm{e}-19$ | 0.711 | 340.0 | $1.00 \mathrm{e}-19$ | 0.721 | 341.0 | $1.70 \mathrm{e}-19$ | 0.732 | 342.0 | 3.86e-19 | 0.743 | 343.0 | $1.49 \mathrm{e}-19$ | 0.754 |
| 344.0 | $9.70 \mathrm{e}-20$ | 0.764 | 345.0 | $1.09 \mathrm{e}-19$ | 0.775 | 346.0 | $1.23 \mathrm{e}-19$ | 0.786 | 347.0 | 1.04e-19 | 0.796 | 348.0 | $9.10 \mathrm{e}-20$ | 0.807 |
| 349.0 | 7.90e-20 | 0.818 | 350.0 | $1.12 \mathrm{e}-19$ | 0.829 | 351.0 | $2.12 \mathrm{e}-19$ | 0.839 | 352.0 | $1.55 \mathrm{e}-19$ | 0.850 | 353.0 | 1.91e-19 | 0.861 |
| 354.0 | 5.81e-19 | 0.871 | 355.0 | 3.64e-19 | 0.882 | 356.0 | $1.41 \mathrm{e}-19$ | 0.893 | 357.0 | 1.17e-19 | 0.904 | 358.0 | 1.20e-19 | 0.914 |
| 359.0 | $1.04 \mathrm{e}-19$ | 0.925 | 360.0 | $9.00 \mathrm{e}-20$ | 0.936 | 361.0 | $8.30 \mathrm{e}-20$ | 0.946 | 362.0 | $8.00 \mathrm{e}-20$ | 0.957 | 363.0 | $9.60 \mathrm{e}-20$ | 0.968 |
| 364.0 | $1.46 \mathrm{e}-19$ | 0.979 | 365.0 | $1.68 \mathrm{e}-19$ | 0.989 | 366.0 | $1.83 \mathrm{e}-19$ | 1.000 | 367.0 | 3.02e-19 | 1.000 | 368.0 | $5.20 \mathrm{e}-19$ | 1.000 |
| 369.0 | 3.88e-19 | 1.000 | 370.0 | $1.78 \mathrm{e}-19$ | 1.000 | 371.0 | $1.13 \mathrm{e}-19$ | 1.000 | 372.0 | 1.00e-19 | 1.000 | 373.0 | 7.70e-20 | 1.000 |
| 374.0 | $6.20 \mathrm{e}-20$ | 1.000 | 375.0 | $5.30 \mathrm{e}-20$ | 1.000 | 376.0 | $5.30 \mathrm{e}-20$ | 1.000 | 377.0 | 5.00e-20 | 1.000 | 378.0 | 5.80e-20 | 1.000 |
| 379.0 | $8.00 \mathrm{e}-20$ | 1.000 | 380.0 | 9.60e-20 | 1.000 | 381.0 | $1.13 \mathrm{e}-19$ | 1.000 | 382.0 | $1.59 \mathrm{e}-19$ | 1.000 | 383.0 | 2.10e-19 | 1.000 |
| 384.0 | 2.41e-19 | 1.000 | 385.0 | 2.03e-19 | 1.000 | 386.0 | $1.34 \mathrm{e}-19$ | 1.000 | 387.0 | $9.00 \mathrm{e}-20$ | 1.000 | 388.0 | $5.60 \mathrm{e}-20$ | 1.000 |
| 389.0 | $3.40 \mathrm{e}-20$ | 1.000 | 390.0 | 2.70e-20 | 1.000 | 391.0 | $2.00 \mathrm{e}-20$ | 1.000 | 392.0 | $1.50 \mathrm{e}-20$ | 1.000 | 393.0 | 1.10e-20 | 1.000 |
| 394.0 | 6.00e-21 | 1.000 | 395.0 | 1.00e-20 | 1.000 | 396.0 | $4.00 \mathrm{e}-21$ | 1.000 | 400.0 | $0.00 \mathrm{e}+00$ | 1.000 |  |  |  |
| HONO-NO2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 309.0 | 0.00e+00 | 0.590 | 310.0 | $1.30 \mathrm{e}-20$ | 0.590 | 311.0 | $1.90 \mathrm{e}-20$ | 0.589 | 312.0 | $2.80 \mathrm{e}-20$ | 0.579 | 313.0 | 2.20e-20 | 0.568 |
| 314.0 | $3.60 \mathrm{e}-20$ | 0.557 | 315.0 | $3.00 \mathrm{e}-20$ | 0.546 | 316.0 | $1.40 \mathrm{e}-20$ | 0.536 | 317.0 | $3.10 \mathrm{e}-20$ | 0.525 | 318.0 | $5.60 \mathrm{e}-20$ | 0.514 |
| 319.0 | $3.60 \mathrm{e}-20$ | 0.504 | 320.0 | 4.90e-20 | 0.493 | 321.0 | $7.80 \mathrm{e}-20$ | 0.482 | 322.0 | $4.90 \mathrm{e}-20$ | 0.471 | 323.0 | $5.10 \mathrm{e}-20$ | 0.461 |
| 324.0 | $7.10 \mathrm{e}-20$ | 0.450 | 325.0 | $5.00 \mathrm{e}-20$ | 0.439 | 326.0 | $2.90 \mathrm{e}-20$ | 0.429 | 327.0 | $6.60 \mathrm{e}-20$ | 0.418 | 328.0 | 1.17e-19 | 0.407 |
| 329.0 | $6.10 \mathrm{e}-20$ | 0.396 | 330.0 | $1.11 \mathrm{e}-19$ | 0.386 | 331.0 | $1.79 \mathrm{e}-19$ | 0.375 | 332.0 | $8.70 \mathrm{e}-20$ | 0.364 | 333.0 | $7.60 \mathrm{e}-20$ | 0.354 |
| 334.0 | $9.60 \mathrm{e}-20$ | 0.343 | 335.0 | $9.60 \mathrm{e}-20$ | 0.332 | 336.0 | $7.20 \mathrm{e}-20$ | 0.321 | 337.0 | 5.30e-20 | 0.311 | 338.0 | 1.00e-19 | 0.300 |
| 339.0 | $1.88 \mathrm{e}-19$ | 0.289 | 340.0 | $1.00 \mathrm{e}-19$ | 0.279 | 341.0 | $1.70 \mathrm{e}-19$ | 0.268 | 342.0 | 3.86e-19 | 0.257 | 343.0 | 1.49e-19 | 0.246 |
| 344.0 | $9.70 \mathrm{e}-20$ | 0.236 | 345.0 | $1.09 \mathrm{e}-19$ | 0.225 | 346.0 | $1.23 \mathrm{e}-19$ | 0.214 | 347.0 | 1.04e-19 | 0.204 | 348.0 | $9.10 \mathrm{e}-20$ | 0.193 |
| 349.0 | $7.90 \mathrm{e}-20$ | 0.182 | 350.0 | $1.12 \mathrm{e}-19$ | 0.171 | 351.0 | $2.12 \mathrm{e}-19$ | 0.161 | 352.0 | $1.55 \mathrm{e}-19$ | 0.150 | 353.0 | 1.91e-19 | 0.139 |
| 354.0 | 5.81e-19 | 0.129 | 355.0 | 3.64e-19 | 0.118 | 356.0 | 1.41e-19 | 0.107 | 357.0 | 1.17e-19 | 0.096 | 358.0 | $1.20 \mathrm{e}-19$ | 0.086 |
| 359.0 | $1.04 \mathrm{e}-19$ | 0.075 | 360.0 | $9.00 \mathrm{e}-20$ | 0.064 | 361.0 | $8.30 \mathrm{e}-20$ | 0.054 | 362.0 | $8.00 \mathrm{e}-20$ | 0.043 | 363.0 | $9.60 \mathrm{e}-20$ | 0.032 |

Table A-3 (continued)

| $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 364.0 | 1.46e-19 | 0.021 | 365.0 | 1.68e-19 | 0.011 | 366.0 | 1.83e-19 | 0.000 |  |  |  |  |  |  |
| HNO3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 190.0 | 1.36e-17 | 1.000 | 195.0 | $1.02 \mathrm{e}-17$ | 1.000 | 200.0 | 5.88e-18 | 1.000 | 205.0 | $2.80 \mathrm{e}-18$ | 1.000 | 210.0 | $1.04 \mathrm{e}-18$ | 1.000 |
| 215.0 | $3.65 \mathrm{e}-19$ | 1.000 | 220.0 | $1.49 \mathrm{e}-19$ | 1.000 | 225.0 | 8.81e-20 | 1.000 | 230.0 | $5.75 \mathrm{e}-20$ | 1.000 | 235.0 | $3.75 \mathrm{e}-20$ | 1.000 |
| 240.0 | $2.58 \mathrm{e}-20$ | 1.000 | 245.0 | $2.11 \mathrm{e}-20$ | 1.000 | 250.0 | 1.97e-20 | 1.000 | 255.0 | $1.95 \mathrm{e}-20$ | 1.000 | 260.0 | 1.91e-20 | 1.000 |
| 265.0 | 1.80e-20 | 1.000 | 270.0 | $1.62 \mathrm{e}-20$ | 1.000 | 275.0 | 1.38e-20 | 1.000 | 280.0 | $1.12 \mathrm{e}-20$ | 1.000 | 285.0 | $8.58 \mathrm{e}-21$ | 1.000 |
| 290.0 | $6.15 \mathrm{e}-21$ | 1.000 | 295.0 | $4.12 \mathrm{e}-21$ | 1.000 | 300.0 | $2.63 \mathrm{e}-21$ | 1.000 | 305.0 | $1.50 \mathrm{e}-21$ | 1.000 | 310.0 | $8.10 \mathrm{e}-22$ | 1.000 |
| 315.0 | $4.10 \mathrm{e}-22$ | 1.000 | 320.0 | $2.00 \mathrm{e}-22$ | 1.000 | 325.0 | $9.50 \mathrm{e}-23$ | 1.000 | 330.0 | $4.30 \mathrm{e}-23$ | 1.000 | 335.0 | $2.20 \mathrm{e}-23$ | 1.000 |
| 340.0 | $1.00 \mathrm{e}-23$ | 1.000 | 345.0 | $6.00 \mathrm{e}-24$ | 1.000 | 350.0 | $4.00 \mathrm{e}-24$ | 1.000 | 355.0 | $0.00 \mathrm{e}+00$ | 1.000 |  |  |  |
| $\mathrm{HO} 2 \mathrm{NO} 2$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 190.0 | 1.01e-17 | 1.000 | 195.0 | 8.16e-18 | 1.000 | 200.0 | $5.63 \mathrm{e}-18$ | 1.000 | 205.0 | 3.67e-18 | 1.000 | 210.0 | 2.39e-18 | 1.000 |
| 215.0 | 1.61e-18 | 1.000 | 220.0 | $1.18 \mathrm{e}-18$ | 1.000 | 225.0 | $9.32 \mathrm{e}-19$ | 1.000 | 230.0 | 7.88e-19 | 1.000 | 235.0 | 6.80e-19 | 1.000 |
| 240.0 | $5.79 \mathrm{e}-19$ | 1.000 | 245.0 | $4.97 \mathrm{e}-19$ | 1.000 | 250.0 | 4.11e-19 | 1.000 | 255.0 | $3.49 \mathrm{e}-19$ | 1.000 | 260.0 | $2.84 \mathrm{e}-19$ | 1.000 |
| 265.0 | $2.29 \mathrm{e}-19$ | 1.000 | 270.0 | $1.80 \mathrm{e}-19$ | 1.000 | 275.0 | $1.33 \mathrm{e}-19$ | 1.000 | 280.0 | $9.30 \mathrm{e}-20$ | 1.000 | 285.0 | $6.20 \mathrm{e}-20$ | 1.000 |
| 290.0 | $3.90 \mathrm{e}-20$ | 1.000 | 295.0 | $2.40 \mathrm{e}-20$ | 1.000 | 300.0 | $1.40 \mathrm{e}-20$ | 1.000 | 305.0 | $8.50 \mathrm{e}-21$ | 1.000 | 310.0 | $5.30 \mathrm{e}-21$ | 1.000 |
| 315.0 | $3.90 \mathrm{e}-21$ | 1.000 | 320.0 | $2.40 \mathrm{e}-21$ | 1.000 | 325.0 | $1.50 \mathrm{e}-21$ | 1.000 | 330.0 | $9.00 \mathrm{e}-22$ | 1.000 | 335.0 | $0.00 \mathrm{e}+00$ | 1.000 |
| H 2 O 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 190.0 | $6.72 \mathrm{e}-19$ | 1.000 | 195.0 | 5.63e-19 | 1.000 | 200.0 | $4.75 \mathrm{e}-19$ | 1.000 | 205.0 | 4.08e-19 | 1.000 | 210.0 | $3.57 \mathrm{e}-19$ | 1.000 |
| 215.0 | $3.07 \mathrm{e}-19$ | 1.000 | 220.0 | $2.58 \mathrm{e}-19$ | 1.000 | 225.0 | 2.17e-19 | 1.000 | 230.0 | 1.82e-19 | 1.000 | 235.0 | $1.50 \mathrm{e}-19$ | 1.000 |
| 240.0 | $1.24 \mathrm{e}-19$ | 1.000 | 245.0 | 1.02e-19 | 1.000 | 250.0 | $8.30 \mathrm{e}-20$ | 1.000 | 255.0 | $6.70 \mathrm{e}-20$ | 1.000 | 260.0 | $5.30 \mathrm{e}-20$ | 1.000 |
| 265.0 | $4.20 \mathrm{e}-20$ | 1.000 | 270.0 | $3.30 \mathrm{e}-20$ | 1.000 | 275.0 | $2.60 \mathrm{e}-20$ | 1.000 | 280.0 | 2.00e-20 | 1.000 | 285.0 | $1.50 \mathrm{e}-20$ | 1.000 |
| 290.0 | $1.20 \mathrm{e}-20$ | 1.000 | 295.0 | $9.00 \mathrm{e}-21$ | 1.000 | 300.0 | $6.80 \mathrm{e}-21$ | 1.000 | 305.0 | 5.10e-21 | 1.000 | 310.0 | $3.90 \mathrm{e}-21$ | 1.000 |
| 315.0 | $2.90 \mathrm{e}-21$ | 1.000 | 320.0 | $2.20 \mathrm{e}-21$ | 1.000 | 325.0 | $1.60 \mathrm{e}-21$ | 1.000 | 330.0 | $1.30 \mathrm{e}-21$ | 1.000 | 335.0 | $1.00 \mathrm{e}-21$ | 1.000 |
| 340.0 | $7.00 \mathrm{e}-22$ | 1.000 | 345.0 | $5.00 \mathrm{e}-22$ | 1.000 | 350.0 | $4.00 \mathrm{e}-22$ | 1.000 | 355.0 | $0.00 \mathrm{e}+00$ | 1.000 |  |  |  |
| HCHO_R |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 240.0 | 6.40e-22 | 0.270 | 241.0 | $5.60 \mathrm{e}-22$ | 0.272 | 242.0 | $1.05 \mathrm{e}-21$ | 0.274 | 243.0 | $1.15 \mathrm{e}-21$ | 0.276 | 244.0 | 8.20e-22 | 0.278 |
| 245.0 | $1.03 \mathrm{e}-21$ | 0.280 | 246.0 | $9.80 \mathrm{e}-22$ | 0.282 | 247.0 | $1.35 \mathrm{e}-21$ | 0.284 | 248.0 | $1.91 \mathrm{e}-21$ | 0.286 | 249.0 | $2.82 \mathrm{e}-21$ | 0.288 |
| 250.0 | $2.05 \mathrm{e}-21$ | 0.290 | 251.0 | $1.70 \mathrm{e}-21$ | 0.291 | 252.0 | 2.88e-21 | 0.292 | 253.0 | $2.55 \mathrm{e}-21$ | 0.293 | 254.0 | $2.55 \mathrm{e}-21$ | 0.294 |
| 255.0 | $3.60 \mathrm{e}-21$ | 0.295 | 256.0 | $5.09 \mathrm{e}-21$ | 0.296 | 257.0 | $3.39 \mathrm{e}-21$ | 0.297 | 258.0 | 2.26e-21 | 0.298 | 259.0 | 5.04e-21 | 0.299 |
| 260.0 | $5.05 \mathrm{e}-21$ | 0.300 | 261.0 | $5.49 \mathrm{e}-21$ | 0.308 | 262.0 | $5.20 \mathrm{e}-21$ | 0.316 | 263.0 | $9.33 \mathrm{e}-21$ | 0.324 | 264.0 | $8.23 \mathrm{e}-21$ | 0.332 |
| 265.0 | $4.30 \mathrm{e}-21$ | 0.340 | 266.0 | $4.95 \mathrm{e}-21$ | 0.348 | 267.0 | $1.24 \mathrm{e}-20$ | 0.356 | 268.0 | $1.11 \mathrm{e}-20$ | 0.364 | 269.0 | $8.78 \mathrm{e}-21$ | 0.372 |
| 270.0 | $9.36 \mathrm{e}-21$ | 0.380 | 271.0 | $1.79 \mathrm{e}-20$ | 0.399 | 272.0 | $1.23 \mathrm{e}-20$ | 0.418 | 273.0 | $6.45 \mathrm{e}-21$ | 0.437 | 274.0 | 6.56e-21 | 0.456 |
| 275.0 | $2.23 \mathrm{e}-20$ | 0.475 | 276.0 | $2.42 \mathrm{e}-20$ | 0.494 | 277.0 | $1.40 \mathrm{e}-20$ | 0.513 | 278.0 | $1.05 \mathrm{e}-20$ | 0.532 | 279.0 | 2.55e-20 | 0.551 |
| 280.0 | $2.08 \mathrm{e}-20$ | 0.570 | 281.0 | $1.48 \mathrm{e}-20$ | 0.586 | 282.0 | 8.81e-21 | 0.602 | 283.0 | $1.07 \mathrm{e}-20$ | 0.618 | 284.0 | $4.49 \mathrm{e}-20$ | 0.634 |
| 285.0 | $3.59 \mathrm{e}-20$ | 0.650 | 286.0 | $1.96 \mathrm{e}-20$ | 0.666 | 287.0 | $1.30 \mathrm{e}-20$ | 0.682 | 288.0 | 3.36e-20 | 0.698 | 289.0 | $2.84 \mathrm{e}-20$ | 0.714 |
| 290.0 | $1.30 \mathrm{e}-20$ | 0.730 | 291.0 | $1.75 \mathrm{e}-20$ | 0.735 | 292.0 | $8.32 \mathrm{e}-21$ | 0.740 | 293.0 | $3.73 \mathrm{e}-20$ | 0.745 | 294.0 | $6.54 \mathrm{e}-20$ | 0.750 |
| 295.0 | $3.95 \mathrm{e}-20$ | 0.755 | 296.0 | $2.33 \mathrm{e}-20$ | 0.760 | 297.0 | $1.51 \mathrm{e}-20$ | 0.765 | 298.0 | $4.04 \mathrm{e}-20$ | 0.770 | 299.0 | $2.87 \mathrm{e}-20$ | 0.775 |
| 300.0 | $8.71 \mathrm{e}-21$ | 0.780 | 301.0 | $1.72 \mathrm{e}-20$ | 0.780 | 302.0 | 1.06e-20 | 0.780 | 303.0 | $3.20 \mathrm{e}-20$ | 0.780 | 304.0 | $6.90 \mathrm{e}-20$ | 0.780 |
| 305.0 | $4.91 \mathrm{e}-20$ | 0.780 | 306.0 | $4.63 \mathrm{e}-20$ | 0.780 | 307.0 | $2.10 \mathrm{e}-20$ | 0.780 | 308.0 | $1.49 \mathrm{e}-20$ | 0.780 | 309.0 | 3.41e-20 | 0.780 |
| 310.0 | $1.95 \mathrm{e}-20$ | 0.780 | 311.0 | $5.21 \mathrm{e}-21$ | 0.764 | 312.0 | $1.12 \mathrm{e}-20$ | 0.748 | 313.0 | $1.12 \mathrm{e}-20$ | 0.732 | 314.0 | $4.75 \mathrm{e}-20$ | 0.716 |
| 315.0 | $5.25 \mathrm{e}-20$ | 0.700 | 316.0 | $2.90 \mathrm{e}-20$ | 0.684 | 317.0 | $5.37 \mathrm{e}-20$ | 0.668 | 318.0 | $2.98 \mathrm{e}-20$ | 0.652 | 319.0 | $9.18 \mathrm{e}-21$ | 0.636 |
| 320.0 | $1.26 \mathrm{e}-20$ | 0.620 | 321.0 | $1.53 \mathrm{e}-20$ | 0.585 | 322.0 | $6.69 \mathrm{e}-21$ | 0.550 | 323.0 | $3.45 \mathrm{e}-21$ | 0.515 | 324.0 | 8.16e-21 | 0.480 |
| 325.0 | $1.85 \mathrm{e}-20$ | 0.445 | 326.0 | $5.95 \mathrm{e}-20$ | 0.410 | 327.0 | $3.49 \mathrm{e}-20$ | 0.375 | 328.0 | $1.09 \mathrm{e}-20$ | 0.340 | 329.0 | $3.35 \mathrm{e}-20$ | 0.305 |
| 330.0 | $3.32 \mathrm{e}-20$ | 0.270 | 331.0 | $1.07 \mathrm{e}-20$ | 0.243 | 332.0 | $2.89 \mathrm{e}-21$ | 0.216 | 333.0 | $2.15 \mathrm{e}-21$ | 0.189 | 334.0 | $1.71 \mathrm{e}-21$ | 0.162 |
| 335.0 | $1.43 \mathrm{e}-21$ | 0.135 | 336.0 | $1.94 \mathrm{e}-21$ | 0.108 | 337.0 | 4.17e-21 | 0.081 | 338.0 | 2.36e-20 | 0.054 | 339.0 | $4.71 \mathrm{e}-20$ | 0.027 |
| 340.0 | $2.48 \mathrm{e}-20$ | 0.000 |  |  |  |  |  |  |  |  |  |  |  |  |
| HCHO M |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 240.0 | $6.40 \mathrm{e}-22$ | 0.490 | 241.0 | $5.60 \mathrm{e}-22$ | 0.490 | 242.0 | $1.05 \mathrm{e}-21$ | 0.490 | 243.0 | $1.15 \mathrm{e}-21$ | 0.490 | 244.0 | $8.20 \mathrm{e}-22$ | 0.490 |
| 245.0 | $1.03 \mathrm{e}-21$ | 0.490 | 246.0 | $9.80 \mathrm{e}-22$ | 0.490 | 247.0 | $1.35 \mathrm{e}-21$ | 0.490 | 248.0 | 1.91e-21 | 0.490 | 249.0 | $2.82 \mathrm{e}-21$ | 0.490 |
| 250.0 | $2.05 \mathrm{e}-21$ | 0.490 | 251.0 | $1.70 \mathrm{e}-21$ | 0.490 | 252.0 | $2.88 \mathrm{e}-21$ | 0.490 | 253.0 | $2.55 \mathrm{e}-21$ | 0.490 | 254.0 | $2.55 \mathrm{e}-21$ | 0.490 |
| 255.0 | $3.60 \mathrm{e}-21$ | 0.490 | 256.0 | $5.09 \mathrm{e}-21$ | 0.490 | 257.0 | $3.39 \mathrm{e}-21$ | 0.490 | 258.0 | 2.26e-21 | 0.490 | 259.0 | $5.04 \mathrm{e}-21$ | 0.490 |
| 260.0 | $5.05 \mathrm{e}-21$ | 0.490 | 261.0 | $5.49 \mathrm{e}-21$ | 0.484 | 262.0 | $5.20 \mathrm{e}-21$ | 0.478 | 263.0 | $9.33 \mathrm{e}-21$ | 0.472 | 264.0 | $8.23 \mathrm{e}-21$ | 0.466 |
| 265.0 | $4.30 \mathrm{e}-21$ | 0.460 | 266.0 | $4.95 \mathrm{e}-21$ | 0.454 | 267.0 | $1.24 \mathrm{e}-20$ | 0.448 | 268.0 | $1.11 \mathrm{e}-20$ | 0.442 | 269.0 | $8.78 \mathrm{e}-21$ | 0.436 |
| 270.0 | $9.36 \mathrm{e}-21$ | 0.430 | 271.0 | $1.79 \mathrm{e}-20$ | 0.419 | 272.0 | $1.23 \mathrm{e}-20$ | 0.408 | 273.0 | $6.45 \mathrm{e}-21$ | 0.397 | 274.0 | $6.56 \mathrm{e}-21$ | 0.386 |
| 275.0 | $2.23 \mathrm{e}-20$ | 0.375 | 276.0 | $2.42 \mathrm{e}-20$ | 0.364 | 277.0 | $1.40 \mathrm{e}-20$ | 0.353 | 278.0 | $1.05 \mathrm{e}-20$ | 0.342 | 279.0 | $2.55 \mathrm{e}-20$ | 0.331 |
| 280.0 | $2.08 \mathrm{e}-20$ | 0.320 | 281.0 | $1.48 \mathrm{e}-20$ | 0.312 | 282.0 | 8.81e-21 | 0.304 | 283.0 | $1.07 \mathrm{e}-20$ | 0.296 | 284.0 | $4.49 \mathrm{e}-20$ | 0.288 |
| 285.0 | $3.59 \mathrm{e}-20$ | 0.280 | 286.0 | $1.96 \mathrm{e}-20$ | 0.272 | 287.0 | $1.30 \mathrm{e}-20$ | 0.264 | 288.0 | 3.36e-20 | 0.256 | 289.0 | $2.84 \mathrm{e}-20$ | 0.248 |
| 290.0 | $1.30 \mathrm{e}-20$ | 0.240 | 291.0 | $1.75 \mathrm{e}-20$ | 0.237 | 292.0 | $8.32 \mathrm{e}-21$ | 0.234 | 293.0 | $3.73 \mathrm{e}-20$ | 0.231 | 294.0 | $6.54 \mathrm{e}-20$ | 0.228 |
| 295.0 | $3.95 \mathrm{e}-20$ | 0.225 | 296.0 | $2.33 \mathrm{e}-20$ | 0.222 | 297.0 | 1.51e-20 | 0.219 | 298.0 | $4.04 \mathrm{e}-20$ | 0.216 | 299.0 | 2.87e-20 | 0.213 |
| 300.0 | 8.71e-21 | 0.210 | 301.0 | $1.72 \mathrm{e}-20$ | 0.211 | 302.0 | 1.06e-20 | 0.212 | 303.0 | $3.20 \mathrm{e}-20$ | 0.213 | 304.0 | $6.90 \mathrm{e}-20$ | 0.214 |
| 305.0 | $4.91 \mathrm{e}-20$ | 0.215 | 306.0 | $4.63 \mathrm{e}-20$ | 0.216 | 307.0 | $2.10 \mathrm{e}-20$ | 0.217 | 308.0 | $1.49 \mathrm{e}-20$ | 0.218 | 309.0 | $3.41 \mathrm{e}-20$ | 0.219 |
| 310.0 | $1.95 \mathrm{e}-20$ | 0.220 | 311.0 | $5.21 \mathrm{e}-21$ | 0.236 | 312.0 | 1.12e-20 | 0.252 | 313.0 | 1.12e-20 | 0.268 | 314.0 | $4.75 \mathrm{e}-20$ | 0.284 |
| 315.0 | $5.25 \mathrm{e}-20$ | 0.300 | 316.0 | $2.90 \mathrm{e}-20$ | 0.316 | 317.0 | $5.37 \mathrm{e}-20$ | 0.332 | 318.0 | $2.98 \mathrm{e}-20$ | 0.348 | 319.0 | $9.18 \mathrm{e}-21$ | 0.364 |
| 320.0 | $1.26 \mathrm{e}-20$ | 0.380 | 321.0 | $1.53 \mathrm{e}-20$ | 0.408 | 322.0 | $6.69 \mathrm{e}-21$ | 0.436 | 323.0 | $3.45 \mathrm{e}-21$ | 0.464 | 324.0 | $8.16 \mathrm{e}-21$ | 0.492 |
| 325.0 | $1.85 \mathrm{e}-20$ | 0.520 | 326.0 | $5.95 \mathrm{e}-20$ | 0.548 | 327.0 | $3.49 \mathrm{e}-20$ | 0.576 | 328.0 | $1.09 \mathrm{e}-20$ | 0.604 | 329.0 | $3.35 \mathrm{e}-20$ | 0.632 |
| 330.0 | $3.32 \mathrm{e}-20$ | 0.660 | 331.0 | $1.07 \mathrm{e}-20$ | 0.650 | 332.0 | $2.89 \mathrm{e}-21$ | 0.640 | 333.0 | $2.15 \mathrm{e}-21$ | 0.630 | 334.0 | $1.71 \mathrm{e}-21$ | 0.620 |
| 335.0 | $1.43 \mathrm{e}-21$ | 0.610 | 336.0 | $1.94 \mathrm{e}-21$ | 0.600 | 337.0 | $4.17 \mathrm{e}-21$ | 0.590 | 338.0 | 2.36e-20 | 0.580 | 339.0 | $4.71 \mathrm{e}-20$ | 0.570 |
| 340.0 | $2.48 \mathrm{e}-20$ | 0.560 | 341.0 | $7.59 \mathrm{e}-21$ | 0.525 | 342.0 | 6.81e-21 | 0.490 | 343.0 | $1.95 \mathrm{e}-20$ | 0.455 | 344.0 | $1.14 \mathrm{e}-20$ | 0.420 |
| 345.0 | $3.23 \mathrm{e}-21$ | 0.385 | 346.0 | $1.13 \mathrm{e}-21$ | 0.350 | 347.0 | $6.60 \mathrm{e}-22$ | 0.315 | 348.0 | $1.22 \mathrm{e}-21$ | 0.280 | 349.0 | $3.20 \mathrm{e}-22$ | 0.245 |

Table A-3 (continued)

| $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 350.0 | $3.80 \mathrm{e}-22$ | 0.210 | 351.0 | $1.04 \mathrm{e}-21$ | 0.192 | 352.0 | $7.13 \mathrm{e}-21$ | 0.174 | 353.0 | $2.21 \mathrm{e}-20$ | 0.156 | 354.0 | $1.54 \mathrm{e}-20$ | 0.138 |
| 355.0 | 6.76e-21 | 0.120 | 356.0 | $1.35 \mathrm{e}-21$ | 0.102 | 357.0 | $3.60 \mathrm{e}-22$ | 0.084 | 358.0 | $5.70 \mathrm{e}-23$ | 0.066 | 359.0 | $5.80 \mathrm{e}-22$ | 0.048 |
| 360.0 | 8.20e-22 | 0.000 |  |  |  |  |  |  |  |  |  |  |  |  |
| CCHO_R |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 262.0 | $2.44 \mathrm{e}-20$ | 0.326 | 266.0 | $3.05 \mathrm{e}-20$ | 0.358 | 270.0 | $3.42 \mathrm{e}-20$ | 0.390 | 274.0 | $4.03 \mathrm{e}-20$ | 0.466 | 278.0 | $4.19 \mathrm{e}-20$ | 0.542 |
| 280.0 | 4.50e-20 | 0.580 | 281.0 | $4.69 \mathrm{e}-20$ | 0.575 | 282.0 | $4.72 \mathrm{e}-20$ | 0.570 | 283.0 | $4.75 \mathrm{e}-20$ | 0.565 | 284.0 | $4.61 \mathrm{e}-20$ | 0.560 |
| 285.0 | 4.49e-20 | 0.555 | 286.0 | 4.44e-20 | 0.550 | 287.0 | $4.59 \mathrm{e}-20$ | 0.545 | 288.0 | $4.72 \mathrm{e}-20$ | 0.540 | 289.0 | $4.77 \mathrm{e}-20$ | 0.535 |
| 290.0 | $4.89 \mathrm{e}-20$ | 0.530 | 291.0 | 4.78e-20 | 0.520 | 292.0 | $4.68 \mathrm{e}-20$ | 0.510 | 293.0 | $4.53 \mathrm{e}-20$ | 0.500 | 294.0 | $4.33 \mathrm{e}-20$ | 0.490 |
| 295.0 | $4.27 \mathrm{e}-20$ | 0.480 | 296.0 | $4.24 \mathrm{e}-20$ | 0.470 | 297.0 | $4.38 \mathrm{e}-20$ | 0.460 | 298.0 | $4.41 \mathrm{e}-20$ | 0.450 | 299.0 | 4.26e-20 | 0.440 |
| 300.0 | 4.16e-20 | 0.430 | 301.0 | 3.99e-20 | 0.418 | 302.0 | 3.86e-20 | 0.406 | 303.0 | $3.72 \mathrm{e}-20$ | 0.394 | 304.0 | 3.48e-20 | 0.382 |
| 305.0 | 3.42e-20 | 0.370 | 306.0 | 3.42e-20 | 0.354 | 307.0 | $3.36 \mathrm{e}-20$ | 0.338 | 308.0 | $3.33 \mathrm{e}-20$ | 0.322 | 309.0 | $3.14 \mathrm{e}-20$ | 0.306 |
| 310.0 | $2.93 \mathrm{e}-20$ | 0.290 | 311.0 | 2.76e-20 | 0.266 | 312.0 | $2.53 \mathrm{e}-20$ | 0.242 | 313.0 | $2.47 \mathrm{e}-20$ | 0.218 | 314.0 | $2.44 \mathrm{e}-20$ | 0.194 |
| 315.0 | 2.20e-20 | 0.170 | 316.0 | $2.04 \mathrm{e}-20$ | 0.156 | 317.0 | $2.07 \mathrm{e}-20$ | 0.142 | 318.0 | $1.98 \mathrm{e}-20$ | 0.128 | 319.0 | $1.87 \mathrm{e}-20$ | 0.114 |
| 320.0 | $1.72 \mathrm{e}-20$ | 0.100 | 321.0 | $1.48 \mathrm{e}-20$ | 0.088 | 322.0 | $1.40 \mathrm{e}-20$ | 0.076 | 323.0 | $1.24 \mathrm{e}-20$ | 0.064 | 324.0 | $1.09 \mathrm{e}-20$ | 0.052 |
| 325.0 | $1.14 \mathrm{e}-20$ | 0.040 | 326.0 | $1.07 \mathrm{e}-20$ | 0.032 | 327.0 | 8.58e-21 | 0.024 | 328.0 | 7.47e-21 | 0.016 | 329.0 | $7.07 \mathrm{e}-21$ | 0.008 |
| C 2 CHO |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 294.0 | 5.80e-20 | 0.890 | 295.0 | 5.57e-20 | 0.885 | 296.0 | $5.37 \mathrm{e}-20$ | 0.880 | 297.0 | 5.16e-20 | 0.875 | 298.0 | 5.02e-20 | 0.870 |
| 299.0 | 5.02e-20 | 0.865 | 300.0 | 5.04e-20 | 0.860 | 301.0 | $5.09 \mathrm{e}-20$ | 0.855 | 302.0 | 5.07e-20 | 0.850 | 303.0 | $4.94 \mathrm{e}-20$ | 0.818 |
| 304.0 | $4.69 \mathrm{e}-20$ | 0.786 | 305.0 | 4.32e-20 | 0.755 | 306.0 | $4.04 \mathrm{e}-20$ | 0.723 | 307.0 | 3.81e-20 | 0.691 | 308.0 | $3.65 \mathrm{e}-20$ | 0.659 |
| 309.0 | 3.62e-20 | 0.627 | 310.0 | 3.60e-20 | 0.596 | 311.0 | $3.53 \mathrm{e}-20$ | 0.564 | 312.0 | $3.50 \mathrm{e}-20$ | 0.532 | 313.0 | $3.32 \mathrm{e}-20$ | 0.500 |
| 314.0 | 3.06e-20 | 0.480 | 315.0 | 2.77e-20 | 0.460 | 316.0 | $2.43 \mathrm{e}-20$ | 0.440 | 317.0 | $2.18 \mathrm{e}-20$ | 0.420 | 318.0 | $2.00 \mathrm{e}-20$ | 0.400 |
| 319.0 | 1.86e-20 | 0.380 | 320.0 | $1.83 \mathrm{e}-20$ | 0.360 | 321.0 | $1.78 \mathrm{e}-20$ | 0.340 | 322.0 | $1.66 \mathrm{e}-20$ | 0.320 | 323.0 | $1.58 \mathrm{e}-20$ | 0.300 |
| 324.0 | $1.49 \mathrm{e}-20$ | 0.280 | 325.0 | $1.30 \mathrm{e}-20$ | 0.260 | 326.0 | $1.13 \mathrm{e}-20$ | 0.248 | 327.0 | 9.96e-21 | 0.236 | 328.0 | $8.28 \mathrm{e}-21$ | 0.223 |
| 329.0 | $6.85 \mathrm{e}-21$ | 0.211 | 330.0 | $5.75 \mathrm{e}-21$ | 0.199 | 331.0 | $4.94 \mathrm{e}-21$ | 0.187 | 332.0 | $4.66 \mathrm{e}-21$ | 0.174 | 333.0 | $4.30 \mathrm{e}-21$ | 0.162 |
| 334.0 | $3.73 \mathrm{e}-21$ | 0.150 | 335.0 | $3.25 \mathrm{e}-21$ | 0.133 | 336.0 | $2.80 \mathrm{e}-21$ | 0.117 | 337.0 | 2.30e-21 | 0.100 | 338.0 | $1.85 \mathrm{e}-21$ | 0.083 |
| 339.0 | 1.66e-21 | 0.067 | 340.0 | $1.55 \mathrm{e}-21$ | 0.050 | 341.0 | $1.19 \mathrm{e}-21$ | 0.033 | 342.0 | 7.60e-22 | 0.017 | 343.0 | $4.50 \mathrm{e}-22$ | 0.000 |
| ACETONE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 250.0 | 2.47e-20 | 0.760 | 254.0 | $3.04 \mathrm{e}-20$ | 0.776 | 258.0 | $3.61 \mathrm{e}-20$ | 0.792 | 262.0 | $4.15 \mathrm{e}-20$ | 0.768 | 266.0 | $4.58 \mathrm{e}-20$ | 0.704 |
| 270.0 | 4.91e-20 | 0.640 | 274.0 | 5.06e-20 | 0.604 | 278.0 | $5.07 \mathrm{e}-20$ | 0.568 | 280.0 | $5.05 \mathrm{e}-20$ | 0.550 | 281.0 | 5.01e-20 | 0.525 |
| 282.0 | $4.94 \mathrm{e}-20$ | 0.500 | 283.0 | 4.86e-20 | 0.475 | 284.0 | $4.76 \mathrm{e}-20$ | 0.450 | 285.0 | $4.68 \mathrm{e}-20$ | 0.425 | 286.0 | $4.58 \mathrm{e}-20$ | 0.400 |
| 287.0 | $4.50 \mathrm{e}-20$ | 0.375 | 288.0 | 4.41e-20 | 0.350 | 289.0 | $4.29 \mathrm{e}-20$ | 0.325 | 290.0 | $4.19 \mathrm{e}-20$ | 0.302 | 291.0 | $4.08 \mathrm{e}-20$ | 0.284 |
| 292.0 | $3.94 \mathrm{e}-20$ | 0.266 | 293.0 | $3.81 \mathrm{e}-20$ | 0.249 | 294.0 | $3.67 \mathrm{e}-20$ | 0.232 | 295.0 | $3.52 \mathrm{e}-20$ | 0.217 | 296.0 | $3.35 \mathrm{e}-20$ | 0.201 |
| 297.0 | $3.20 \mathrm{e}-20$ | 0.187 | 298.0 | $3.07 \mathrm{e}-20$ | 0.173 | 299.0 | $2.91 \mathrm{e}-20$ | 0.160 | 300.0 | 2.77e-20 | 0.147 | 301.0 | 2.66e-20 | 0.135 |
| 302.0 | $2.53 \mathrm{e}-20$ | 0.124 | 303.0 | $2.37 \mathrm{e}-20$ | 0.114 | 304.0 | $2.24 \mathrm{e}-20$ | 0.104 | 305.0 | $2.11 \mathrm{e}-20$ | 0.095 | 306.0 | $1.95 \mathrm{e}-20$ | 0.086 |
| 307.0 | $1.80 \mathrm{e}-20$ | 0.078 | 308.0 | $1.66 \mathrm{e}-20$ | 0.071 | 309.0 | $1.54 \mathrm{e}-20$ | 0.064 | 310.0 | $1.41 \mathrm{e}-20$ | 0.057 | 311.0 | $1.28 \mathrm{e}-20$ | 0.052 |
| 312.0 | 1.17e-20 | 0.046 | 313.0 | $1.08 \mathrm{e}-20$ | 0.042 | 314.0 | $9.67 \mathrm{e}-21$ | 0.037 | 315.0 | $8.58 \mathrm{e}-21$ | 0.033 | 316.0 | $7.77 \mathrm{e}-21$ | 0.029 |
| 317.0 | $6.99 \mathrm{e}-21$ | 0.026 | 318.0 | 6.08e-21 | 0.023 | 319.0 | $5.30 \mathrm{e}-21$ | 0.020 | 320.0 | $4.67 \mathrm{e}-21$ | 0.018 | 321.0 | $4.07 \mathrm{e}-21$ | 0.016 |
| 322.0 | $3.44 \mathrm{e}-21$ | 0.014 | 323.0 | $2.87 \mathrm{e}-21$ | 0.012 | 324.0 | $2.43 \mathrm{e}-21$ | 0.011 | 325.0 | $2.05 \mathrm{e}-21$ | 0.009 | 326.0 | $1.68 \mathrm{e}-21$ | 0.008 |
| 327.0 | $1.35 \mathrm{e}-21$ | 0.007 | 328.0 | $1.08 \mathrm{e}-21$ | 0.006 | 329.0 | $8.60 \mathrm{e}-22$ | 0.005 | 330.0 | $6.70 \mathrm{e}-22$ | 0.005 | 331.0 | 5.10e-22 | 0.004 |
| 332.0 | $4.00 \mathrm{e}-22$ | 0.003 | 333.0 | $3.10 \mathrm{e}-22$ | 0.003 | 334.0 | $2.60 \mathrm{e}-22$ | 0.002 | 335.0 | $1.70 \mathrm{e}-22$ | 0.002 | 336.0 | $1.40 \mathrm{e}-22$ | 0.002 |
| 337.0 | 1.10e-22 | 0.002 | 338.0 | $9.00 \mathrm{e}-23$ | 0.001 | 339.0 | $6.00 \mathrm{e}-23$ | 0.001 | 340.0 | $5.00 \mathrm{e}-23$ | 0.001 | 341.0 | $5.00 \mathrm{e}-23$ | 0.001 |
| 342.0 | $3.00 \mathrm{e}-23$ | 0.001 | 343.0 | $4.00 \mathrm{e}-23$ | 0.001 | 344.0 | $2.00 \mathrm{e}-23$ | 0.000 |  |  |  |  |  |  |
| KETONE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 198.5 | 3.95e-19 | 1.000 | 199.0 | $1.61 \mathrm{e}-19$ | 1.000 | 199.5 | $7.75 \mathrm{e}-20$ | 1.000 | 200.0 | 3.76e-20 | 1.000 | 200.5 | $2.51 \mathrm{e}-20$ | 1.000 |
| 201.0 | $1.83 \mathrm{e}-20$ | 1.000 | 201.5 | 1.36e-20 | 1.000 | 202.0 | $1.16 \mathrm{e}-20$ | 1.000 | 202.5 | $8.97 \mathrm{e}-21$ | 1.000 | 203.0 | $4.62 \mathrm{e}-21$ | 1.000 |
| 203.5 | $3.18 \mathrm{e}-21$ | 1.000 | 204.0 | $2.42 \mathrm{e}-21$ | 1.000 | 204.5 | $2.01 \mathrm{e}-21$ | 1.000 | 205.0 | $1.77 \mathrm{e}-21$ | 1.000 | 205.5 | $1.64 \mathrm{e}-21$ | 1.000 |
| 206.0 | $1.54 \mathrm{e}-21$ | 1.000 | 206.5 | $1.52 \mathrm{e}-21$ | 1.000 | 207.0 | $1.54 \mathrm{e}-21$ | 1.000 | 207.5 | $1.62 \mathrm{e}-21$ | 1.000 | 208.0 | $1.64 \mathrm{e}-21$ | 1.000 |
| 208.5 | $1.60 \mathrm{e}-21$ | 1.000 | 209.0 | $1.57 \mathrm{e}-21$ | 1.000 | 209.5 | $1.49 \mathrm{e}-21$ | 1.000 | 210.0 | $1.47 \mathrm{e}-21$ | 1.000 | 210.5 | $1.52 \mathrm{e}-21$ | 1.000 |
| 211.0 | $1.50 \mathrm{e}-21$ | 1.000 | 211.5 | $1.62 \mathrm{e}-21$ | 1.000 | 212.0 | $1.81 \mathrm{e}-21$ | 1.000 | 212.5 | 2.10e-21 | 1.000 | 213.0 | $2.23 \mathrm{e}-21$ | 1.000 |
| 213.5 | 2.06e-21 | 1.000 | 214.0 | $1.69 \mathrm{e}-21$ | 1.000 | 214.5 | $1.49 \mathrm{e}-21$ | 1.000 | 215.0 | $1.42 \mathrm{e}-21$ | 1.000 | 215.5 | $1.42 \mathrm{e}-21$ | 1.000 |
| 216.0 | $1.42 \mathrm{e}-21$ | 1.000 | 216.5 | $1.48 \mathrm{e}-21$ | 1.000 | 217.0 | $1.48 \mathrm{e}-21$ | 1.000 | 217.5 | $1.53 \mathrm{e}-21$ | 1.000 | 218.0 | $1.56 \mathrm{e}-21$ | 1.000 |
| 218.5 | $1.67 \mathrm{e}-21$ | 1.000 | 219.0 | $1.68 \mathrm{e}-21$ | 1.000 | 219.5 | $1.78 \mathrm{e}-21$ | 1.000 | 220.0 | $1.85 \mathrm{e}-21$ | 1.000 | 220.5 | $1.92 \mathrm{e}-21$ | 1.000 |
| 221.0 | $2.01 \mathrm{e}-21$ | 1.000 | 221.5 | $2.11 \mathrm{e}-21$ | 1.000 | 222.0 | $2.23 \mathrm{e}-21$ | 1.000 | 222.5 | $2.33 \mathrm{e}-21$ | 1.000 | 223.0 | $2.48 \mathrm{e}-21$ | 1.000 |
| 223.5 | $2.60 \mathrm{e}-21$ | 1.000 | 224.0 | $2.74 \mathrm{e}-21$ | 1.000 | 224.5 | $2.85 \mathrm{e}-21$ | 1.000 | 225.0 | $3.04 \mathrm{e}-21$ | 1.000 | 225.5 | $3.15 \mathrm{e}-21$ | 1.000 |
| 226.0 | $3.33 \mathrm{e}-21$ | 1.000 | 226.5 | $3.55 \mathrm{e}-21$ | 1.000 | 227.0 | $3.73 \mathrm{e}-21$ | 1.000 | 227.5 | $3.93 \mathrm{e}-21$ | 1.000 | 228.0 | $4.11 \mathrm{e}-21$ | 1.000 |
| 228.5 | $4.34 \mathrm{e}-21$ | 1.000 | 229.0 | 4.56e-21 | 1.000 | 229.5 | $4.75 \mathrm{e}-21$ | 1.000 | 230.0 | $5.01 \mathrm{e}-21$ | 1.000 | 230.5 | $5.27 \mathrm{e}-21$ | 1.000 |
| 231.0 | $5.53 \mathrm{e}-21$ | 1.000 | 231.5 | $5.83 \mathrm{e}-21$ | 1.000 | 232.0 | $6.15 \mathrm{e}-21$ | 1.000 | 232.5 | $6.45 \mathrm{e}-21$ | 1.000 | 233.0 | $6.73 \mathrm{e}-21$ | 1.000 |
| 233.5 | 7.02e-21 | 1.000 | 234.0 | 7.42e-21 | 1.000 | 234.5 | $7.83 \mathrm{e}-21$ | 1.000 | 235.0 | $8.11 \mathrm{e}-21$ | 1.000 | 235.5 | $8.45 \mathrm{e}-21$ | 1.000 |
| 236.0 | 8.82e-21 | 1.000 | 236.5 | $9.21 \mathrm{e}-21$ | 1.000 | 237.0 | $9.65 \mathrm{e}-21$ | 1.000 | 237.5 | $1.00 \mathrm{e}-20$ | 1.000 | 238.0 | $1.05 \mathrm{e}-20$ | 1.000 |
| 238.5 | $1.10 \mathrm{e}-20$ | 1.000 | 239.0 | $1.15 \mathrm{e}-20$ | 1.000 | 239.5 | $1.20 \mathrm{e}-20$ | 1.000 | 240.0 | $1.23 \mathrm{e}-20$ | 1.000 | 240.5 | $1.28 \mathrm{e}-20$ | 1.000 |
| 241.0 | $1.32 \mathrm{e}-20$ | 1.000 | 241.5 | $1.38 \mathrm{e}-20$ | 1.000 | 242.0 | $1.44 \mathrm{e}-20$ | 1.000 | 242.5 | $1.50 \mathrm{e}-20$ | 1.000 | 243.0 | $1.57 \mathrm{e}-20$ | 1.000 |
| 243.5 | $1.63 \mathrm{e}-20$ | 1.000 | 244.0 | $1.68 \mathrm{e}-20$ | 1.000 | 244.5 | $1.75 \mathrm{e}-20$ | 1.000 | 245.0 | 1.81e-20 | 1.000 | 245.5 | $1.88 \mathrm{e}-20$ | 1.000 |
| 246.0 | 1.96e-20 | 1.000 | 246.5 | $2.03 \mathrm{e}-20$ | 1.000 | 247.0 | $2.11 \mathrm{e}-20$ | 1.000 | 247.5 | $2.19 \mathrm{e}-20$ | 1.000 | 248.0 | $2.25 \mathrm{e}-20$ | 1.000 |
| 248.5 | $2.33 \mathrm{e}-20$ | 1.000 | 249.0 | $2.40 \mathrm{e}-20$ | 1.000 | 249.5 | $2.48 \mathrm{e}-20$ | 1.000 | 250.0 | 2.56e-20 | 1.000 | 250.5 | $2.64 \mathrm{e}-20$ | 1.000 |
| 251.0 | 2.73e-20 | 1.000 | 251.5 | $2.81 \mathrm{e}-20$ | 1.000 | 252.0 | $2.88 \mathrm{e}-20$ | 1.000 | 252.5 | $2.98 \mathrm{e}-20$ | 1.000 | 253.0 | $3.07 \mathrm{e}-20$ | 1.000 |
| 253.5 | 3.16e-20 | 1.000 | 254.0 | $3.25 \mathrm{e}-20$ | 1.000 | 254.5 | $3.34 \mathrm{e}-20$ | 1.000 | 255.0 | $3.43 \mathrm{e}-20$ | 1.000 | 255.5 | $3.51 \mathrm{e}-20$ | 1.000 |
| 256.0 | $3.59 \mathrm{e}-20$ | 1.000 | 256.5 | 3.67e-20 | 1.000 | 257.0 | $3.75 \mathrm{e}-20$ | 1.000 | 257.5 | $3.84 \mathrm{e}-20$ | 1.000 | 258.0 | $3.94 \mathrm{e}-20$ | 1.000 |
| 258.5 | $4.03 \mathrm{e}-20$ | 1.000 | 259.0 | $4.13 \mathrm{e}-20$ | 1.000 | 259.5 | $4.22 \mathrm{e}-20$ | 1.000 | 260.0 | $4.28 \mathrm{e}-20$ | 1.000 | 260.5 | $4.33 \mathrm{e}-20$ | 1.000 |
| 261.0 | $4.41 \mathrm{e}-20$ | 1.000 | 261.5 | $4.49 \mathrm{e}-20$ | 1.000 | 262.0 | $4.57 \mathrm{e}-20$ | 1.000 | 262.5 | $4.65 \mathrm{e}-20$ | 1.000 | 263.0 | $4.72 \mathrm{e}-20$ | 1.000 |
| 263.5 | $4.78 \mathrm{e}-20$ | 1.000 | 264.0 | $4.85 \mathrm{e}-20$ | 1.000 | 264.5 | $4.92 \mathrm{e}-20$ | 1.000 | 265.0 | $4.99 \mathrm{e}-20$ | 1.000 | 265.5 | $5.04 \mathrm{e}-20$ | 1.000 |
| 266.0 | 5.12e-20 | 1.000 | 266.5 | $5.22 \mathrm{e}-20$ | 1.000 | 267.0 | $5.28 \mathrm{e}-20$ | 1.000 | 267.5 | $5.34 \mathrm{e}-20$ | 1.000 | 268.0 | 5.41e-20 | 1.000 |

Table A-3 (continued)

| $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 268.5 | 5.46e-20 | 1.000 | 269.0 | 5.51e-20 | 1.000 | 269.5 | $5.55 \mathrm{e}-20$ | 1.000 | 270.0 | $5.59 \mathrm{e}-20$ | 1.000 | 270.5 | $5.63 \mathrm{e}-20$ | 1.000 |
| 271.0 | 5.66e-20 | 1.000 | 271.5 | 5.70e-20 | 1.000 | 272.0 | $5.74 \mathrm{e}-20$ | 1.000 | 272.5 | $5.78 \mathrm{e}-20$ | 1.000 | 273.0 | $5.81 \mathrm{e}-20$ | 1.000 |
| 273.5 | 5.86e-20 | 1.000 | 274.0 | $5.90 \mathrm{e}-20$ | 1.000 | 274.5 | $5.93 \mathrm{e}-20$ | 1.000 | 275.0 | $5.96 \mathrm{e}-20$ | 1.000 | 275.5 | $5.97 \mathrm{e}-20$ | 1.000 |
| 276.0 | 5.98e-20 | 1.000 | 276.5 | 5.98e-20 | 1.000 | 277.0 | $5.99 \mathrm{e}-20$ | 1.000 | 277.5 | $5.99 \mathrm{e}-20$ | 1.000 | 278.0 | $5.98 \mathrm{e}-20$ | 1.000 |
| 278.5 | 5.96e-20 | 1.000 | 279.0 | 5.96e-20 | 1.000 | 279.5 | $5.95 \mathrm{e}-20$ | 1.000 | 280.0 | $5.94 \mathrm{e}-20$ | 1.000 | 280.5 | $5.92 \mathrm{e}-20$ | 1.000 |
| 281.0 | 5.90e-20 | 1.000 | 281.5 | 5.88e-20 | 1.000 | 282.0 | $5.86 \mathrm{e}-20$ | 1.000 | 282.5 | $5.83 \mathrm{e}-20$ | 1.000 | 283.0 | $5.79 \mathrm{e}-20$ | 1.000 |
| 283.5 | $5.75 \mathrm{e}-20$ | 1.000 | 284.0 | 5.71e-20 | 1.000 | 284.5 | $5.67 \mathrm{e}-20$ | 1.000 | 285.0 | $5.61 \mathrm{e}-20$ | 1.000 | 285.5 | $5.56 \mathrm{e}-20$ | 1.000 |
| 286.0 | 5.51e-20 | 1.000 | 286.5 | $5.45 \mathrm{e}-20$ | 1.000 | 287.0 | $5.41 \mathrm{e}-20$ | 1.000 | 287.5 | $5.37 \mathrm{e}-20$ | 1.000 | 288.0 | $5.33 \mathrm{e}-20$ | 1.000 |
| 288.5 | 5.27e-20 | 1.000 | 289.0 | 5.21e-20 | 1.000 | 289.5 | $5.15 \mathrm{e}-20$ | 1.000 | 290.0 | $5.08 \mathrm{e}-20$ | 1.000 | 290.5 | $4.99 \mathrm{e}-20$ | 1.000 |
| 291.0 | $4.89 \mathrm{e}-20$ | 1.000 | 291.5 | $4.82 \mathrm{e}-20$ | 1.000 | 292.0 | $4.73 \mathrm{e}-20$ | 1.000 | 292.5 | $4.62 \mathrm{e}-20$ | 1.000 | 293.0 | $4.53 \mathrm{e}-20$ | 1.000 |
| 293.5 | 4.41e-20 | 1.000 | 294.0 | $4.32 \mathrm{e}-20$ | 1.000 | 294.5 | $4.23 \mathrm{e}-20$ | 1.000 | 295.0 | $4.15 \mathrm{e}-20$ | 1.000 | 295.5 | $4.11 \mathrm{e}-20$ | 1.000 |
| 296.0 | 4.01e-20 | 1.000 | 296.5 | 3.94e-20 | 1.000 | 297.0 | $3.88 \mathrm{e}-20$ | 1.000 | 297.5 | $3.77 \mathrm{e}-20$ | 1.000 | 298.0 | $3.69 \mathrm{e}-20$ | 1.000 |
| 298.5 | $3.63 \mathrm{e}-20$ | 1.000 | 299.0 | $3.54 \mathrm{e}-20$ | 1.000 | 299.5 | $3.46 \mathrm{e}-20$ | 1.000 | 300.0 | $3.36 \mathrm{e}-20$ | 1.000 | 300.5 | $3.24 \mathrm{e}-20$ | 1.000 |
| 301.0 | 3.16e-20 | 1.000 | 301.5 | 3.06e-20 | 1.000 | 302.0 | $2.95 \mathrm{e}-20$ | 1.000 | 302.5 | $2.82 \mathrm{e}-20$ | 1.000 | 303.0 | $2.70 \mathrm{e}-20$ | 1.000 |
| 303.5 | 2.59e-20 | 1.000 | 304.0 | 2.49e-20 | 1.000 | 304.5 | $2.42 \mathrm{e}-20$ | 1.000 | 305.0 | $2.34 \mathrm{e}-20$ | 1.000 | 305.5 | $2.28 \mathrm{e}-20$ | 1.000 |
| 306.0 | $2.19 \mathrm{e}-20$ | 1.000 | 306.5 | 2.11e-20 | 1.000 | 307.0 | $2.04 \mathrm{e}-20$ | 1.000 | 307.5 | $1.93 \mathrm{e}-20$ | 1.000 | 308.0 | $1.88 \mathrm{e}-20$ | 1.000 |
| 308.5 | $1.80 \mathrm{e}-20$ | 1.000 | 309.0 | $1.73 \mathrm{e}-20$ | 1.000 | 309.5 | $1.66 \mathrm{e}-20$ | 1.000 | 310.0 | $1.58 \mathrm{e}-20$ | 1.000 | 310.5 | $1.48 \mathrm{e}-20$ | 1.000 |
| 311.0 | $1.42 \mathrm{e}-20$ | 1.000 | 311.5 | $1.34 \mathrm{e}-20$ | 1.000 | 312.0 | $1.26 \mathrm{e}-20$ | 1.000 | 312.5 | $1.17 \mathrm{e}-20$ | 1.000 | 313.0 | $1.13 \mathrm{e}-20$ | 1.000 |
| 313.5 | $1.08 \mathrm{e}-20$ | 1.000 | 314.0 | $1.04 \mathrm{e}-20$ | 1.000 | 314.5 | $9.69 \mathrm{e}-21$ | 1.000 | 315.0 | 8.91e-21 | 1.000 | 315.5 | $8.61 \mathrm{e}-21$ | 1.000 |
| 316.0 | 7.88e-21 | 1.000 | 316.5 | $7.25 \mathrm{e}-21$ | 1.000 | 317.0 | $6.92 \mathrm{e}-21$ | 1.000 | 317.5 | $6.43 \mathrm{e}-21$ | 1.000 | 318.0 | $6.07 \mathrm{e}-21$ | 1.000 |
| 318.5 | $5.64 \mathrm{e}-21$ | 1.000 | 319.0 | $5.19 \mathrm{e}-21$ | 1.000 | 319.5 | $4.66 \mathrm{e}-21$ | 1.000 | 320.0 | $4.36 \mathrm{e}-21$ | 1.000 | 320.5 | $3.95 \mathrm{e}-21$ | 1.000 |
| 321.0 | $3.64 \mathrm{e}-21$ | 1.000 | 321.5 | $3.38 \mathrm{e}-21$ | 1.000 | 322.0 | $3.17 \mathrm{e}-21$ | 1.000 | 322.5 | $2.80 \mathrm{e}-21$ | 1.000 | 323.0 | $2.62 \mathrm{e}-21$ | 1.000 |
| 323.5 | $2.29 \mathrm{e}-21$ | 1.000 | 324.0 | 2.13e-21 | 1.000 | 324.5 | $1.93 \mathrm{e}-21$ | 1.000 | 325.0 | $1.70 \mathrm{e}-21$ | 1.000 | 325.5 | $1.58 \mathrm{e}-21$ | 1.000 |
| 326.0 | $1.48 \mathrm{e}-21$ | 1.000 | 326.5 | $1.24 \mathrm{e}-21$ | 1.000 | 327.0 | $1.20 \mathrm{e}-21$ | 1.000 | 327.5 | $1.04 \mathrm{e}-21$ | 1.000 | 328.0 | $9.51 \mathrm{e}-22$ | 1.000 |
| 328.5 | 8.44e-22 | 1.000 | 329.0 | 7.26e-22 | 1.000 | 329.5 | $6.70 \mathrm{e}-22$ | 1.000 | 330.0 | 6.08e-22 | 1.000 | 330.5 | $5.15 \mathrm{e}-22$ | 1.000 |
| 331.0 | 4.56e-22 | 1.000 | 331.5 | $4.13 \mathrm{e}-22$ | 1.000 | 332.0 | $3.56 \mathrm{e}-22$ | 1.000 | 332.5 | $3.30 \mathrm{e}-22$ | 1.000 | 333.0 | $2.97 \mathrm{e}-22$ | 1.000 |
| 333.5 | 2.67e-22 | 1.000 | 334.0 | 2.46e-22 | 1.000 | 334.5 | $2.21 \mathrm{e}-22$ | 1.000 | 335.0 | $1.93 \mathrm{e}-22$ | 1.000 | 335.5 | $1.56 \mathrm{e}-22$ | 1.000 |
| 336.0 | 1.47e-22 | 1.000 | 336.5 | $1.37 \mathrm{e}-22$ | 1.000 | 337.0 | $1.27 \mathrm{e}-22$ | 1.000 | 337.5 | $1.19 \mathrm{e}-22$ | 1.000 | 338.0 | $1.09 \mathrm{e}-22$ | 1.000 |
| 338.5 | $1.01 \mathrm{e}-22$ | 1.000 | 339.0 | $9.09 \mathrm{e}-23$ | 1.000 | 339.5 | $8.22 \mathrm{e}-23$ | 1.000 | 340.0 | 7.66e-23 | 1.000 | 340.5 | $7.43 \mathrm{e}-23$ | 1.000 |
| 341.0 | 6.83e-23 | 1.000 | 341.5 | 6.72e-23 | 1.000 | 342.0 | $6.04 \mathrm{e}-23$ | 1.000 | 342.5 | $4.78 \mathrm{e}-23$ | 1.000 | 343.0 | $0.00 \mathrm{e}+00$ | 1.000 |
| $\mathrm{COOH}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 210.0 | 3.12e-19 | 1.000 | 215.0 | 2.09e-19 | 1.000 | 220.0 | 1.54e-19 | 1.000 | 225.0 | 1.22e-19 | 1.000 | 230.0 | $9.62 \mathrm{e}-20$ | 1.000 |
| 235.0 | 7.61e-20 | 1.000 | 240.0 | $6.05 \mathrm{e}-20$ | 1.000 | 245.0 | $4.88 \mathrm{e}-20$ | 1.000 | 250.0 | $3.98 \mathrm{e}-20$ | 1.000 | 255.0 | $3.23 \mathrm{e}-20$ | 1.000 |
| 260.0 | 2.56e-20 | 1.000 | 265.0 | 2.11e-20 | 1.000 | 270.0 | $1.70 \mathrm{e}-20$ | 1.000 | 275.0 | $1.39 \mathrm{e}-20$ | 1.000 | 280.0 | $1.09 \mathrm{e}-20$ | 1.000 |
| 285.0 | $8.63 \mathrm{e}-21$ | 1.000 | 290.0 | 6.91e-21 | 1.000 | 295.0 | $5.51 \mathrm{e}-21$ | 1.000 | 300.0 | $4.13 \mathrm{e}-21$ | 1.000 | 305.0 | $3.13 \mathrm{e}-21$ | 1.000 |
| 310.0 | $2.39 \mathrm{e}-21$ | 1.000 | 315.0 | $1.82 \mathrm{e}-21$ | 1.000 | 320.0 | $1.37 \mathrm{e}-21$ | 1.000 | 325.0 | $1.05 \mathrm{e}-21$ | 1.000 | 330.0 | $7.90 \mathrm{e}-22$ | 1.000 |
| 335.0 | 6.10e-22 | 1.000 | 340.0 | 4.70e-22 | 1.000 | 345.0 | $3.50 \mathrm{e}-22$ | 1.000 | 350.0 | $2.70 \mathrm{e}-22$ | 1.000 | 355.0 | $2.10 \mathrm{e}-22$ | 1.000 |
| 360.0 | $1.60 \mathrm{e}-22$ | 1.000 | 365.0 | $1.20 \mathrm{e}-22$ | 1.000 | 370.0 | $0.00 \mathrm{e}+00$ | 1.000 |  |  |  |  |  |  |
| GLY_R |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 230.0 | 2.87e-21 | 1.000 | 235.0 | 2.87e-21 | 1.000 | 240.0 | 4.30e-21 | 1.000 | 245.0 | $5.73 \mathrm{e}-21$ | 1.000 | 250.0 | $8.60 \mathrm{e}-21$ | 1.000 |
| 255.0 | $1.15 \mathrm{e}-20$ | 1.000 | 260.0 | $1.43 \mathrm{e}-20$ | 1.000 | 265.0 | $1.86 \mathrm{e}-20$ | 1.000 | 270.0 | $2.29 \mathrm{e}-20$ | 1.000 | 275.0 | $2.58 \mathrm{e}-20$ | 1.000 |
| 280.0 | $2.87 \mathrm{e}-20$ | 1.000 | 285.0 | $3.30 \mathrm{e}-20$ | 1.000 | 290.0 | $3.15 \mathrm{e}-20$ | 1.000 | 295.0 | $3.30 \mathrm{e}-20$ | 1.000 | 300.0 | $3.58 \mathrm{e}-20$ | 1.000 |
| 305.0 | 2.72e-20 | 1.000 | 310.0 | 2.72e-20 | 1.000 | 312.5 | $2.87 \mathrm{e}-20$ | 1.000 | 315.0 | $2.29 \mathrm{e}-20$ | 1.000 | 320.0 | $1.43 \mathrm{e}-20$ | 1.000 |
| 325.0 | $1.15 \mathrm{e}-20$ | 1.000 | 327.5 | $1.43 \mathrm{e}-20$ | 1.000 | 330.0 | $1.15 \mathrm{e}-20$ | 1.000 | 335.0 | $2.87 \mathrm{e}-21$ | 1.000 | 340.0 | $0.00 \mathrm{e}+00$ | 1.000 |
| 345.0 | $0.00 \mathrm{e}+00$ | 1.000 | 350.0 | $0.00 \mathrm{e}+00$ | 1.000 | 355.0 | $0.00 \mathrm{e}+00$ | 1.000 | 360.0 | $2.29 \mathrm{e}-21$ | 1.000 | 365.0 | $2.87 \mathrm{e}-21$ | 1.000 |
| 370.0 | $8.03 \mathrm{e}-21$ | 1.000 | 375.0 | $1.00 \mathrm{e}-20$ | 1.000 | 380.0 | $1.72 \mathrm{e}-20$ | 0.972 | 382.0 | $1.58 \mathrm{e}-20$ | 0.855 | 384.0 | $1.49 \mathrm{e}-20$ | 0.737 |
| 386.0 | $1.49 \mathrm{e}-20$ | 0.619 | 388.0 | 2.87e-20 | 0.502 | 390.0 | $3.15 \mathrm{e}-20$ | 0.384 | 391.0 | $3.24 \mathrm{e}-20$ | 0.326 | 392.0 | $3.04 \mathrm{e}-20$ | 0.267 |
| 393.0 | 2.23e-20 | 0.208 | 394.0 | $2.63 \mathrm{e}-20$ | 0.149 | 395.0 | $3.04 \mathrm{e}-20$ | 0.090 | 396.0 | $2.63 \mathrm{e}-20$ | 0.032 | 397.0 | $2.43 \mathrm{e}-20$ | 0.000 |
| 398.0 | $3.24 \mathrm{e}-20$ | 0.000 | 399.0 | $3.04 \mathrm{e}-20$ | 0.000 | 400.0 | $2.84 \mathrm{e}-20$ | 0.000 | 401.0 | $3.24 \mathrm{e}-20$ | 0.000 | 402.0 | 4.46e-20 | 0.000 |
| 403.0 | 5.27e-20 | 0.000 | 404.0 | 4.26e-20 | 0.000 | 405.0 | $3.04 \mathrm{e}-20$ | 0.000 | 406.0 | $3.04 \mathrm{e}-20$ | 0.000 | 407.0 | $2.84 \mathrm{e}-20$ | 0.000 |
| 408.0 | $2.43 \mathrm{e}-20$ | 0.000 | 409.0 | 2.84e-20 | 0.000 | 410.0 | $6.08 \mathrm{e}-20$ | 0.000 | 411.0 | $5.07 \mathrm{e}-20$ | 0.000 | 411.5 | 6.08e-20 | 0.000 |
| 412.0 | 4.86e-20 | 0.000 | 413.0 | 8.31e-20 | 0.000 | 413.5 | $6.48 \mathrm{e}-20$ | 0.000 | 414.0 | $7.50 \mathrm{e}-20$ | 0.000 | 414.5 | 8.11e-20 | 0.000 |
| 415.0 | 8.11e-20 | 0.000 | 415.5 | 6.89e-20 | 0.000 | 416.0 | 4.26e-20 | 0.000 | 417.0 | $4.86 \mathrm{e}-20$ | 0.000 | 418.0 | $5.88 \mathrm{e}-20$ | 0.000 |
| GLY_ABS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 230.0 | 2.87e-21 | 1.000 | 235.0 | 2.87e-21 | 1.000 | 240.0 | $4.30 \mathrm{e}-21$ | 1.000 | 245.0 | $5.73 \mathrm{e}-21$ | 1.000 | 250.0 | $8.60 \mathrm{e}-21$ | 1.000 |
| 255.0 | $1.15 \mathrm{e}-20$ | 1.000 | 260.0 | $1.43 \mathrm{e}-20$ | 1.000 | 265.0 | $1.86 \mathrm{e}-20$ | 1.000 | 270.0 | $2.29 \mathrm{e}-20$ | 1.000 | 275.0 | $2.58 \mathrm{e}-20$ | 1.000 |
| 280.0 | $2.87 \mathrm{e}-20$ | 1.000 | 285.0 | $3.30 \mathrm{e}-20$ | 1.000 | 290.0 | $3.15 \mathrm{e}-20$ | 1.000 | 295.0 | $3.30 \mathrm{e}-20$ | 1.000 | 300.0 | $3.58 \mathrm{e}-20$ | 1.000 |
| 305.0 | 2.72e-20 | 1.000 | 310.0 | 2.72e-20 | 1.000 | 312.5 | $2.87 \mathrm{e}-20$ | 1.000 | 315.0 | $2.29 \mathrm{e}-20$ | 1.000 | 320.0 | $1.43 \mathrm{e}-20$ | 1.000 |
| 325.0 | $1.15 \mathrm{e}-20$ | 1.000 | 327.5 | $1.43 \mathrm{e}-20$ | 1.000 | 330.0 | $1.15 \mathrm{e}-20$ | 1.000 | 335.0 | 2.87e-21 | 1.000 | 340.0 | $0.00 \mathrm{e}+00$ | 1.000 |
| 355.0 | $0.00 \mathrm{e}+00$ | 1.000 | 360.0 | $2.29 \mathrm{e}-21$ | 1.000 | 365.0 | $2.87 \mathrm{e}-21$ | 1.000 | 370.0 | $8.03 \mathrm{e}-21$ | 1.000 | 375.0 | $1.00 \mathrm{e}-20$ | 1.000 |
| 380.0 | $1.72 \mathrm{e}-20$ | 1.000 | 382.0 | $1.58 \mathrm{e}-20$ | 1.000 | 384.0 | $1.49 \mathrm{e}-20$ | 1.000 | 386.0 | $1.49 \mathrm{e}-20$ | 1.000 | 388.0 | 2.87e-20 | 1.000 |
| 390.0 | $3.15 \mathrm{e}-20$ | 1.000 | 391.0 | $3.24 \mathrm{e}-20$ | 1.000 | 392.0 | $3.04 \mathrm{e}-20$ | 1.000 | 393.0 | $2.23 \mathrm{e}-20$ | 1.000 | 394.0 | $2.63 \mathrm{e}-20$ | 1.000 |
| 395.0 | $3.04 \mathrm{e}-20$ | 1.000 | 396.0 | $2.63 \mathrm{e}-20$ | 1.000 | 397.0 | $2.43 \mathrm{e}-20$ | 1.000 | 398.0 | $3.24 \mathrm{e}-20$ | 1.000 | 399.0 | $3.04 \mathrm{e}-20$ | 1.000 |
| 400.0 | $2.84 \mathrm{e}-20$ | 1.000 | 401.0 | $3.24 \mathrm{e}-20$ | 1.000 | 402.0 | $4.46 \mathrm{e}-20$ | 1.000 | 403.0 | $5.27 \mathrm{e}-20$ | 1.000 | 404.0 | 4.26e-20 | 1.000 |
| 405.0 | $3.04 \mathrm{e}-20$ | 1.000 | 406.0 | $3.04 \mathrm{e}-20$ | 1.000 | 407.0 | $2.84 \mathrm{e}-20$ | 1.000 | 408.0 | $2.43 \mathrm{e}-20$ | 1.000 | 409.0 | $2.84 \mathrm{e}-20$ | 1.000 |
| 410.0 | 6.08e-20 | 1.000 | 411.0 | $5.07 \mathrm{e}-20$ | 1.000 | 411.5 | $6.08 \mathrm{e}-20$ | 1.000 | 412.0 | $4.86 \mathrm{e}-20$ | 1.000 | 413.0 | 8.31e-20 | 1.000 |
| 413.5 | 6.48e-20 | 1.000 | 414.0 | 7.50e-20 | 1.000 | 414.5 | 8.11e-20 | 1.000 | 415.0 | $8.11 \mathrm{e}-20$ | 1.000 | 415.5 | $6.89 \mathrm{e}-20$ | 1.000 |
| 416.0 | 4.26e-20 | 1.000 | 417.0 | 4.86e-20 | 1.000 | 418.0 | $5.88 \mathrm{e}-20$ | 1.000 | 419.0 | $6.69 \mathrm{e}-20$ | 1.000 | 420.0 | $3.85 \mathrm{e}-20$ | 1.000 |
| 421.0 | 5.67e-20 | 1.000 | 421.5 | 4.46e-20 | 1.000 | 422.0 | $5.27 \mathrm{e}-20$ | 1.000 | 422.5 | $1.05 \mathrm{e}-19$ | 1.000 | 423.0 | $8.51 \mathrm{e}-20$ | 1.000 |
| 424.0 | 6.08e-20 | 1.000 | 425.0 | $7.29 \mathrm{e}-20$ | 1.000 | 426.0 | $1.18 \mathrm{e}-19$ | 1.000 | 426.5 | $1.30 \mathrm{e}-19$ | 1.000 | 427.0 | 1.07e-19 | 1.000 |
| 428.0 | 1.66e-19 | 1.000 | 429.0 | $4.05 \mathrm{e}-20$ | 1.000 | 430.0 | $5.07 \mathrm{e}-20$ | 1.000 | 431.0 | $4.86 \mathrm{e}-20$ | 1.000 | 432.0 | $4.05 \mathrm{e}-20$ | 1.000 |

Table A-3 (continued)

| $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{gathered} \mathrm{WL} \\ (\mathrm{~nm}) \end{gathered}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 433.0 | $3.65 \mathrm{e}-20$ | 1.000 | 434.0 | $4.05 \mathrm{e}-20$ | 1.000 | 434.5 | 6.08e-20 | 1.000 | 435.0 | 5.07e-20 | 1.000 | 436.0 | $8.11 \mathrm{e}-20$ | 1.000 |
| 436.5 | 1.13e-19 | 1.000 | 437.0 | 5.27e-20 | 1.000 | 438.0 | $1.01 \mathrm{e}-19$ | 1.000 | 438.5 | $1.38 \mathrm{e}-19$ | 1.000 | 439.0 | 7.70e-20 | 1.000 |
| 440.0 | 2.47e-19 | 1.000 | 441.0 | $8.11 \mathrm{e}-20$ | 1.000 | 442.0 | $6.08 \mathrm{e}-20$ | 1.000 | 443.0 | $7.50 \mathrm{e}-20$ | 1.000 | 444.0 | $9.32 \mathrm{e}-20$ | 1.000 |
| 445.0 | 1.13e-19 | 1.000 | 446.0 | 5.27e-20 | 1.000 | 447.0 | $2.43 \mathrm{e}-20$ | 1.000 | 448.0 | $2.84 \mathrm{e}-20$ | 1.000 | 449.0 | $3.85 \mathrm{e}-20$ | 1.000 |
| 450.0 | 6.08e-20 | 1.000 | 451.0 | $1.09 \mathrm{e}-19$ | 1.000 | 451.5 | $9.32 \mathrm{e}-20$ | 1.000 | 452.0 | 1.22e-19 | 1.000 | 453.0 | 2.39e-19 | 1.000 |
| 454.0 | 1.70e-19 | 1.000 | 455.0 | 3.40e-19 | 1.000 | 455.5 | $4.05 \mathrm{e}-19$ | 1.000 | 456.0 | 1.01e-19 | 1.000 | 457.0 | $1.62 \mathrm{e}-20$ | 1.000 |
| 458.0 | $1.22 \mathrm{e}-20$ | 1.000 | 458.5 | 1.42e-20 | 1.000 | 459.0 | $4.05 \mathrm{e}-21$ | 1.000 | 460.0 | $4.05 \mathrm{e}-21$ | 1.000 | 460.5 | 6.08e-21 | 1.000 |
| 461.0 | $2.03 \mathrm{e}-21$ | 1.000 | 462.0 | $0.00 \mathrm{e}+00$ | 1.000 |  |  |  |  |  |  |  |  |  |
| MGLY_ADJ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 219.0 | 9.84e-21 | 1.000 | 219.5 | $1.04 \mathrm{e}-20$ | 1.000 | 220.0 | 1.06e-20 | 1.000 | 220.5 | 1.11e-20 | 1.000 | 221.0 | $1.15 \mathrm{e}-20$ | 1.000 |
| 221.5 | $1.18 \mathrm{e}-20$ | 1.000 | 222.0 | $1.22 \mathrm{e}-20$ | 1.000 | 222.5 | $1.24 \mathrm{e}-20$ | 1.000 | 223.0 | 1.26e-20 | 1.000 | 223.5 | 1.26e-20 | 1.000 |
| 224.0 | $1.25 \mathrm{e}-20$ | 1.000 | 224.5 | $1.24 \mathrm{e}-20$ | 1.000 | 225.0 | $1.25 \mathrm{e}-20$ | 1.000 | 225.5 | 1.27e-20 | 1.000 | 226.0 | 1.27e-20 | 1.000 |
| 226.5 | $1.29 \mathrm{e}-20$ | 1.000 | 227.0 | $1.31 \mathrm{e}-20$ | 1.000 | 227.5 | $1.32 \mathrm{e}-20$ | 1.000 | 228.0 | $1.35 \mathrm{e}-20$ | 1.000 | 228.5 | 1.37e-20 | 1.000 |
| 229.0 | 1.40e-20 | 1.000 | 229.5 | $1.42 \mathrm{e}-20$ | 1.000 | 230.0 | $1.48 \mathrm{e}-20$ | 1.000 | 230.5 | $1.53 \mathrm{e}-20$ | 1.000 | 231.0 | $1.57 \mathrm{e}-20$ | 1.000 |
| 231.5 | $1.59 \mathrm{e}-20$ | 1.000 | 232.0 | $1.61 \mathrm{e}-20$ | 1.000 | 232.5 | $1.62 \mathrm{e}-20$ | 1.000 | 233.0 | 1.61e-20 | 1.000 | 233.5 | 1.68e-20 | 1.000 |
| 234.0 | $1.74 \mathrm{e}-20$ | 1.000 | 234.5 | $1.80 \mathrm{e}-20$ | 1.000 | 235.0 | $1.84 \mathrm{e}-20$ | 1.000 | 235.5 | $1.87 \mathrm{e}-20$ | 1.000 | 236.0 | $1.89 \mathrm{e}-20$ | 1.000 |
| 236.5 | 1.91e-20 | 1.000 | 237.0 | $1.93 \mathrm{e}-20$ | 1.000 | 237.5 | $1.94 \mathrm{e}-20$ | 1.000 | 238.0 | 1.96e-20 | 1.000 | 238.5 | 1.96e-20 | 1.000 |
| 239.0 | 2.01e-20 | 1.000 | 239.5 | 2.04e-20 | 1.000 | 240.0 | $2.08 \mathrm{e}-20$ | 1.000 | 240.5 | 2.10e-20 | 1.000 | 241.0 | 2.14e-20 | 1.000 |
| 241.5 | 2.16e-20 | 1.000 | 242.0 | 2.19e-20 | 1.000 | 242.5 | $2.20 \mathrm{e}-20$ | 1.000 | 243.0 | $2.23 \mathrm{e}-20$ | 1.000 | 243.5 | 2.26e-20 | 1.000 |
| 244.0 | $2.28 \mathrm{e}-20$ | 1.000 | 244.5 | 2.29e-20 | 1.000 | 245.0 | $2.30 \mathrm{e}-20$ | 1.000 | 245.5 | $2.32 \mathrm{e}-20$ | 1.000 | 246.0 | $2.33 \mathrm{e}-20$ | 1.000 |
| 246.5 | $2.35 \mathrm{e}-20$ | 1.000 | 247.0 | 2.38e-20 | 1.000 | 247.5 | $2.41 \mathrm{e}-20$ | 1.000 | 248.0 | 2.46e-20 | 1.000 | 248.5 | 2.51e-20 | 1.000 |
| 249.0 | $2.57 \mathrm{e}-20$ | 1.000 | 249.5 | $2.61 \mathrm{e}-20$ | 1.000 | 250.0 | $2.65 \mathrm{e}-20$ | 1.000 | 250.5 | 2.67e-20 | 1.000 | 251.0 | $2.69 \mathrm{e}-20$ | 1.000 |
| 251.5 | $2.69 \mathrm{e}-20$ | 1.000 | 252.0 | $2.71 \mathrm{e}-20$ | 1.000 | 252.5 | $2.72 \mathrm{e}-20$ | 1.000 | 253.0 | $2.73 \mathrm{e}-20$ | 1.000 | 253.5 | 2.74e-20 | 1.000 |
| 254.0 | 2.76e-20 | 1.000 | 254.5 | 2.78e-20 | 1.000 | 255.0 | $2.82 \mathrm{e}-20$ | 1.000 | 255.5 | $2.87 \mathrm{e}-20$ | 1.000 | 256.0 | $2.93 \mathrm{e}-20$ | 1.000 |
| 256.5 | $2.98 \mathrm{e}-20$ | 1.000 | 257.0 | $3.07 \mathrm{e}-20$ | 1.000 | 257.5 | $3.12 \mathrm{e}-20$ | 1.000 | 258.0 | $3.17 \mathrm{e}-20$ | 1.000 | 258.5 | $3.21 \mathrm{e}-20$ | 1.000 |
| 259.0 | $3.26 \mathrm{e}-20$ | 1.000 | 259.5 | $3.28 \mathrm{e}-20$ | 1.000 | 260.0 | $3.29 \mathrm{e}-20$ | 1.000 | 260.5 | $3.31 \mathrm{e}-20$ | 1.000 | 261.0 | $3.33 \mathrm{e}-20$ | 1.000 |
| 261.5 | $3.34 \mathrm{e}-20$ | 1.000 | 262.0 | $3.36 \mathrm{e}-20$ | 1.000 | 262.5 | $3.38 \mathrm{e}-20$ | 1.000 | 263.0 | $3.42 \mathrm{e}-20$ | 1.000 | 263.5 | $3.44 \mathrm{e}-20$ | 1.000 |
| 264.0 | $3.48 \mathrm{e}-20$ | 1.000 | 264.5 | $3.54 \mathrm{e}-20$ | 1.000 | 265.0 | $3.59 \mathrm{e}-20$ | 1.000 | 265.5 | $3.65 \mathrm{e}-20$ | 1.000 | 266.0 | $3.73 \mathrm{e}-20$ | 1.000 |
| 266.5 | $3.80 \mathrm{e}-20$ | 1.000 | 267.0 | $3.87 \mathrm{e}-20$ | 1.000 | 267.5 | $3.95 \mathrm{e}-20$ | 1.000 | 268.0 | $4.02 \mathrm{e}-20$ | 1.000 | 268.5 | $4.08 \mathrm{e}-20$ | 1.000 |
| 269.0 | $4.13 \mathrm{e}-20$ | 1.000 | 269.5 | 4.17e-20 | 1.000 | 270.0 | $4.20 \mathrm{e}-20$ | 1.000 | 270.5 | $4.22 \mathrm{e}-20$ | 1.000 | 271.0 | $4.22 \mathrm{e}-20$ | 1.000 |
| 271.5 | $4.22 \mathrm{e}-20$ | 1.000 | 272.0 | $4.23 \mathrm{e}-20$ | 1.000 | 272.5 | $4.24 \mathrm{e}-20$ | 1.000 | 273.0 | $4.27 \mathrm{e}-20$ | 1.000 | 273.5 | $4.29 \mathrm{e}-20$ | 1.000 |
| 274.0 | 4.31e-20 | 1.000 | 274.5 | $4.33 \mathrm{e}-20$ | 1.000 | 275.0 | $4.37 \mathrm{e}-20$ | 1.000 | 275.5 | $4.42 \mathrm{e}-20$ | 1.000 | 276.0 | $4.48 \mathrm{e}-20$ | 1.000 |
| 276.5 | $4.56 \mathrm{e}-20$ | 1.000 | 277.0 | $4.64 \mathrm{e}-20$ | 1.000 | 277.5 | $4.71 \mathrm{e}-20$ | 1.000 | 278.0 | $4.78 \mathrm{e}-20$ | 1.000 | 278.5 | $4.83 \mathrm{e}-20$ | 1.000 |
| 279.0 | $4.87 \mathrm{e}-20$ | 1.000 | 279.5 | 4.90e-20 | 1.000 | 280.0 | $4.92 \mathrm{e}-20$ | 1.000 | 280.5 | $4.93 \mathrm{e}-20$ | 1.000 | 281.0 | $4.94 \mathrm{e}-20$ | 1.000 |
| 281.5 | $4.92 \mathrm{e}-20$ | 1.000 | 282.0 | 4.90e-20 | 1.000 | 282.5 | $4.86 \mathrm{e}-20$ | 1.000 | 283.0 | $4.83 \mathrm{e}-20$ | 1.000 | 283.5 | $4.79 \mathrm{e}-20$ | 1.000 |
| 284.0 | 4.76e-20 | 1.000 | 284.5 | 4.72e-20 | 1.000 | 285.0 | $4.70 \mathrm{e}-20$ | 1.000 | 285.5 | $4.68 \mathrm{e}-20$ | 1.000 | 286.0 | 4.66e-20 | 1.000 |
| 286.5 | $4.65 \mathrm{e}-20$ | 1.000 | 287.0 | $4.65 \mathrm{e}-20$ | 1.000 | 287.5 | $4.68 \mathrm{e}-20$ | 1.000 | 288.0 | $4.73 \mathrm{e}-20$ | 1.000 | 288.5 | $4.78 \mathrm{e}-20$ | 1.000 |
| 289.0 | $4.84 \mathrm{e}-20$ | 1.000 | 289.5 | 4.89e-20 | 1.000 | 290.0 | $4.92 \mathrm{e}-20$ | 1.000 | 290.5 | $4.92 \mathrm{e}-20$ | 1.000 | 291.0 | $4.90 \mathrm{e}-20$ | 1.000 |
| 291.5 | $4.86 \mathrm{e}-20$ | 1.000 | 292.0 | $4.81 \mathrm{e}-20$ | 1.000 | 292.5 | $4.75 \mathrm{e}-20$ | 1.000 | 293.0 | 4.70e-20 | 1.000 | 293.5 | $4.65 \mathrm{e}-20$ | 1.000 |
| 294.0 | $4.58 \mathrm{e}-20$ | 1.000 | 294.5 | 4.48e-20 | 1.000 | 295.0 | $4.38 \mathrm{e}-20$ | 1.000 | 295.5 | $4.27 \mathrm{e}-20$ | 1.000 | 296.0 | $4.17 \mathrm{e}-20$ | 1.000 |
| 296.5 | $4.07 \mathrm{e}-20$ | 1.000 | 297.0 | $3.99 \mathrm{e}-20$ | 1.000 | 297.5 | $3.94 \mathrm{e}-20$ | 1.000 | 298.0 | 3.88e-20 | 1.000 | 298.5 | 3.82e-20 | 1.000 |
| 299.0 | $3.76 \mathrm{e}-20$ | 1.000 | 299.5 | $3.72 \mathrm{e}-20$ | 1.000 | 300.0 | $3.69 \mathrm{e}-20$ | 1.000 | 300.5 | $3.68 \mathrm{e}-20$ | 1.000 | 301.0 | $3.70 \mathrm{e}-20$ | 1.000 |
| 301.5 | $3.72 \mathrm{e}-20$ | 1.000 | 302.0 | $3.74 \mathrm{e}-20$ | 1.000 | 302.5 | $3.74 \mathrm{e}-20$ | 1.000 | 303.0 | $3.75 \mathrm{e}-20$ | 1.000 | 303.5 | $3.71 \mathrm{e}-20$ | 1.000 |
| 304.0 | $3.62 \mathrm{e}-20$ | 1.000 | 304.5 | $3.51 \mathrm{e}-20$ | 1.000 | 305.0 | $3.38 \mathrm{e}-20$ | 1.000 | 305.5 | $3.25 \mathrm{e}-20$ | 1.000 | 306.0 | $3.15 \mathrm{e}-20$ | 1.000 |
| 306.5 | $3.04 \mathrm{e}-20$ | 1.000 | 307.0 | 2.92e-20 | 1.000 | 307.5 | $2.80 \mathrm{e}-20$ | 1.000 | 308.0 | $2.71 \mathrm{e}-20$ | 1.000 | 308.5 | $2.63 \mathrm{e}-20$ | 1.000 |
| 309.0 | 2.52e-20 | 1.000 | 309.5 | 2.43e-20 | 1.000 | 310.0 | $2.34 \mathrm{e}-20$ | 1.000 | 310.5 | $2.25 \mathrm{e}-20$ | 1.000 | 311.0 | $2.19 \mathrm{e}-20$ | 1.000 |
| 311.5 | 2.12e-20 | 1.000 | 312.0 | 2.06e-20 | 1.000 | 312.5 | $2.02 \mathrm{e}-20$ | 1.000 | 313.0 | 1.96e-20 | 1.000 | 313.5 | 1.92e-20 | 1.000 |
| 314.0 | 1.91e-20 | 1.000 | 314.5 | $1.88 \mathrm{e}-20$ | 1.000 | 315.0 | $1.86 \mathrm{e}-20$ | 1.000 | 315.5 | $1.85 \mathrm{e}-20$ | 1.000 | 316.0 | 1.86e-20 | 1.000 |
| 316.5 | $1.87 \mathrm{e}-20$ | 1.000 | 317.0 | $1.87 \mathrm{e}-20$ | 1.000 | 317.5 | $1.87 \mathrm{e}-20$ | 1.000 | 318.0 | $1.83 \mathrm{e}-20$ | 1.000 | 318.5 | $1.75 \mathrm{e}-20$ | 1.000 |
| 319.0 | $1.69 \mathrm{e}-20$ | 1.000 | 319.5 | $1.60 \mathrm{e}-20$ | 1.000 | 320.0 | $1.50 \mathrm{e}-20$ | 1.000 | 320.5 | 1.41e-20 | 1.000 | 321.0 | $1.34 \mathrm{e}-20$ | 1.000 |
| 321.5 | $1.27 \mathrm{e}-20$ | 1.000 | 322.0 | $1.21 \mathrm{e}-20$ | 1.000 | 322.5 | $1.18 \mathrm{e}-20$ | 1.000 | 323.0 | $1.14 \mathrm{e}-20$ | 1.000 | 323.5 | 1.08e-20 | 1.000 |
| 324.0 | $1.01 \mathrm{e}-20$ | 1.000 | 324.5 | $9.62 \mathrm{e}-21$ | 1.000 | 325.0 | $9.28 \mathrm{e}-21$ | 1.000 | 325.5 | $8.75 \mathrm{e}-21$ | 1.000 | 326.0 | $8.49 \mathrm{e}-21$ | 1.000 |
| 326.5 | $8.21 \mathrm{e}-21$ | 1.000 | 327.0 | 7.71e-21 | 1.000 | 327.5 | $7.38 \mathrm{e}-21$ | 1.000 | 328.0 | 7.18e-21 | 1.000 | 328.5 | 6.86e-21 | 1.000 |
| 329.0 | $6.71 \mathrm{e}-21$ | 1.000 | 329.5 | $6.63 \mathrm{e}-21$ | 1.000 | 330.0 | $6.46 \mathrm{e}-21$ | 1.000 | 330.5 | $6.29 \mathrm{e}-21$ | 1.000 | 331.0 | 6.21e-21 | 1.000 |
| 331.5 | $6.18 \mathrm{e}-21$ | 1.000 | 332.0 | $6.20 \mathrm{e}-21$ | 1.000 | 332.5 | $5.49 \mathrm{e}-21$ | 1.000 | 333.0 | 5.21e-21 | 1.000 | 333.5 | $5.38 \mathrm{e}-21$ | 1.000 |
| 334.0 | $5.35 \mathrm{e}-21$ | 1.000 | 334.5 | 5.04e-21 | 1.000 | 335.0 | $4.94 \mathrm{e}-21$ | 1.000 | 335.5 | $4.90 \mathrm{e}-21$ | 1.000 | 336.0 | $4.52 \mathrm{e}-21$ | 1.000 |
| 336.5 | 4.26e-21 | 1.000 | 337.0 | 4.11e-21 | 1.000 | 337.5 | $3.76 \mathrm{e}-21$ | 1.000 | 338.0 | $3.61 \mathrm{e}-21$ | 1.000 | 338.5 | $3.58 \mathrm{e}-21$ | 1.000 |
| 339.0 | 3.47e-21 | 1.000 | 339.5 | 3.32e-21 | 1.000 | 340.0 | $3.22 \mathrm{e}-21$ | 1.000 | 340.5 | $3.10 \mathrm{e}-21$ | 1.000 | 341.0 | $3.00 \mathrm{e}-21$ | 1.000 |
| 341.5 | $2.94 \mathrm{e}-21$ | 1.000 | 342.0 | $2.89 \mathrm{e}-21$ | 1.000 | 342.5 | $2.86 \mathrm{e}-21$ | 1.000 | 343.0 | 2.88e-21 | 1.000 | 343.5 | 2.88e-21 | 1.000 |
| 344.0 | $2.89 \mathrm{e}-21$ | 0.992 | 344.5 | 2.91e-21 | 0.984 | 345.0 | $2.95 \mathrm{e}-21$ | 0.976 | 345.5 | $3.00 \mathrm{e}-21$ | 0.968 | 346.0 | 3.08e-21 | 0.960 |
| 346.5 | $3.18 \mathrm{e}-21$ | 0.953 | 347.0 | $3.25 \mathrm{e}-21$ | 0.945 | 347.5 | $3.30 \mathrm{e}-21$ | 0.937 | 348.0 | $3.39 \mathrm{e}-21$ | 0.929 | 348.5 | $3.51 \mathrm{e}-21$ | 0.921 |
| 349.0 | $3.63 \mathrm{e}-21$ | 0.913 | 349.5 | $3.73 \mathrm{e}-21$ | 0.905 | 350.0 | $3.85 \mathrm{e}-21$ | 0.897 | 350.5 | $3.99 \mathrm{e}-21$ | 0.889 | 351.0 | $4.27 \mathrm{e}-21$ | 0.881 |
| 351.5 | $4.47 \mathrm{e}-21$ | 0.873 | 352.0 | $4.63 \mathrm{e}-21$ | 0.865 | 352.5 | $4.78 \mathrm{e}-21$ | 0.858 | 353.0 | 4.92e-21 | 0.850 | 353.5 | 5.07e-21 | 0.842 |
| 354.0 | $5.23 \mathrm{e}-21$ | 0.834 | 354.5 | 5.39e-21 | 0.826 | 355.0 | $5.56 \mathrm{e}-21$ | 0.818 | 355.5 | 5.77e-21 | 0.810 | 356.0 | 5.97e-21 | 0.802 |
| 356.5 | 6.15e-21 | 0.794 | 357.0 | 6.35e-21 | 0.786 | 357.5 | $6.56 \mathrm{e}-21$ | 0.778 | 358.0 | 6.76e-21 | 0.770 | 358.5 | $6.95 \mathrm{e}-21$ | 0.763 |
| 359.0 | $7.20 \mathrm{e}-21$ | 0.755 | 359.5 | 7.44e-21 | 0.747 | 360.0 | $7.64 \mathrm{e}-21$ | 0.739 | 360.5 | 7.89e-21 | 0.731 | 361.0 | $8.15 \mathrm{e}-21$ | 0.723 |
| 361.5 | $8.43 \mathrm{e}-21$ | 0.715 | 362.0 | 8.71e-21 | 0.707 | 362.5 | $9.02 \mathrm{e}-21$ | 0.699 | 363.0 | $9.33 \mathrm{e}-21$ | 0.691 | 363.5 | $9.65 \mathrm{e}-21$ | 0.683 |
| 364.0 | $1.00 \mathrm{e}-20$ | 0.675 | 364.5 | 1.04e-20 | 0.668 | 365.0 | $1.08 \mathrm{e}-20$ | 0.660 | 365.5 | $1.11 \mathrm{e}-20$ | 0.652 | 366.0 | $1.15 \mathrm{e}-20$ | 0.644 |
| 366.5 | $1.19 \mathrm{e}-20$ | 0.636 | 367.0 | $1.23 \mathrm{e}-20$ | 0.628 | 367.5 | $1.27 \mathrm{e}-20$ | 0.620 | 368.0 | 1.31e-20 | 0.612 | 368.5 | $1.35 \mathrm{e}-20$ | 0.604 |
| 369.0 | 1.40e-20 | 0.596 | 369.5 | 1.44e-20 | 0.588 | 370.0 | $1.47 \mathrm{e}-20$ | 0.580 | 370.5 | $1.51 \mathrm{e}-20$ | 0.573 | 371.0 | $1.55 \mathrm{e}-20$ | 0.565 |
| 371.5 | $1.59 \mathrm{e}-20$ | 0.557 | 372.0 | $1.64 \mathrm{e}-20$ | 0.549 | 372.5 | $1.70 \mathrm{e}-20$ | 0.541 | 373.0 | $1.73 \mathrm{e}-20$ | 0.533 | 373.5 | 1.77e-20 | 0.525 |
| 374.0 | 1.81e-20 | 0.517 | 374.5 | 1.86e-20 | 0.509 | 375.0 | $1.90 \mathrm{e}-20$ | 0.501 | 375.5 | 1.96e-20 | 0.493 | 376.0 | 2.02e-20 | 0.486 |

Table A-3 (continued)

| $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 376.5 | $2.06 \mathrm{e}-20$ | 0.478 | 377.0 | $2.10 \mathrm{e}-20$ | 0.470 | 377.5 | $2.14 \mathrm{e}-20$ | 0.462 | 378.0 | $2.18 \mathrm{e}-20$ | 0.454 | 378.5 | $2.24 \mathrm{e}-20$ | 0.446 |
| 379.0 | $2.30 \mathrm{e}-20$ | 0.438 | 379.5 | $2.37 \mathrm{e}-20$ | 0.430 | 380.0 | $2.42 \mathrm{e}-20$ | 0.422 | 380.5 | $2.47 \mathrm{e}-20$ | 0.414 | 381.0 | $2.54 \mathrm{e}-20$ | 0.406 |
| 381.5 | $2.62 \mathrm{e}-20$ | 0.398 | 382.0 | $2.69 \mathrm{e}-20$ | 0.391 | 382.5 | $2.79 \mathrm{e}-20$ | 0.383 | 383.0 | 2.88e-20 | 0.375 | 383.5 | 2.96e-20 | 0.367 |
| 384.0 | $3.02 \mathrm{e}-20$ | 0.359 | 384.5 | $3.10 \mathrm{e}-20$ | 0.351 | 385.0 | $3.20 \mathrm{e}-20$ | 0.343 | 385.5 | $3.29 \mathrm{e}-20$ | 0.335 | 386.0 | $3.39 \mathrm{e}-20$ | 0.327 |
| 386.5 | $3.51 \mathrm{e}-20$ | 0.319 | 387.0 | $3.62 \mathrm{e}-20$ | 0.311 | 387.5 | $3.69 \mathrm{e}-20$ | 0.303 | 388.0 | $3.70 \mathrm{e}-20$ | 0.296 | 388.5 | $3.77 \mathrm{e}-20$ | 0.288 |
| 389.0 | $3.88 \mathrm{e}-20$ | 0.280 | 389.5 | $3.97 \mathrm{e}-20$ | 0.272 | 390.0 | $4.03 \mathrm{e}-20$ | 0.264 | 390.5 | $4.12 \mathrm{e}-20$ | 0.256 | 391.0 | $4.22 \mathrm{e}-20$ | 0.248 |
| 391.5 | $4.29 \mathrm{e}-20$ | 0.240 | 392.0 | $4.30 \mathrm{e}-20$ | 0.232 | 392.5 | $4.38 \mathrm{e}-20$ | 0.224 | 393.0 | $4.47 \mathrm{e}-20$ | 0.216 | 393.5 | $4.55 \mathrm{e}-20$ | 0.208 |
| 394.0 | 4.56e-20 | 0.201 | 394.5 | $4.59 \mathrm{e}-20$ | 0.193 | 395.0 | $4.67 \mathrm{e}-20$ | 0.185 | 395.5 | $4.80 \mathrm{e}-20$ | 0.177 | 396.0 | $4.87 \mathrm{e}-20$ | 0.169 |
| 396.5 | 4.96e-20 | 0.161 | 397.0 | 5.08e-20 | 0.153 | 397.5 | $5.19 \mathrm{e}-20$ | 0.145 | 398.0 | $5.23 \mathrm{e}-20$ | 0.137 | 398.5 | $5.39 \mathrm{e}-20$ | 0.129 |
| 399.0 | 5.46e-20 | 0.121 | 399.5 | $5.54 \mathrm{e}-20$ | 0.113 | 400.0 | $5.59 \mathrm{e}-20$ | 0.106 | 400.5 | $5.77 \mathrm{e}-20$ | 0.098 | 401.0 | 5.91e-20 | 0.090 |
| 401.5 | $5.99 \mathrm{e}-20$ | 0.082 | 402.0 | 6.06e-20 | 0.074 | 402.5 | $6.20 \mathrm{e}-20$ | 0.066 | 403.0 | $6.35 \mathrm{e}-20$ | 0.058 | 403.5 | $6.52 \mathrm{e}-20$ | 0.050 |
| 404.0 | $6.54 \mathrm{e}-20$ | 0.042 | 404.5 | $6.64 \mathrm{e}-20$ | 0.034 | 405.0 | $6.93 \mathrm{e}-20$ | 0.026 | 405.5 | $7.15 \mathrm{e}-20$ | 0.018 | 406.0 | $7.19 \mathrm{e}-20$ | 0.011 |
| 406.5 | $7.32 \mathrm{e}-20$ | 0.003 | 407.0 | $7.58 \mathrm{e}-20$ | 0.000 | 407.5 | $7.88 \mathrm{e}-20$ | 0.000 | 408.0 | $7.97 \mathrm{e}-20$ | 0.000 | 408.5 | $7.91 \mathrm{e}-20$ | 0.000 |
| 409.0 | 8.11e-20 | 0.000 | 409.5 | 8.41e-20 | 0.000 | 410.0 | $8.53 \mathrm{e}-20$ | 0.000 | 410.5 | $8.59 \mathrm{e}-20$ | 0.000 | 411.0 | $8.60 \mathrm{e}-20$ | 0.000 |
| 411.5 | 8.80e-20 | 0.000 | 412.0 | $9.04 \mathrm{e}-20$ | 0.000 | 412.5 | $9.45 \mathrm{e}-20$ | 0.000 | 413.0 | $9.34 \mathrm{e}-20$ | 0.000 | 413.5 | $9.37 \mathrm{e}-20$ | 0.000 |
| 414.0 | $9.63 \mathrm{e}-20$ | 0.000 | 414.5 | $9.71 \mathrm{e}-20$ | 0.000 | 415.0 | $9.70 \mathrm{e}-20$ | 0.000 | 415.5 | $9.65 \mathrm{e}-20$ | 0.000 | 416.0 | $9.69 \mathrm{e}-20$ | 0.000 |
| 416.5 | $9.89 \mathrm{e}-20$ | 0.000 | 417.0 | 1.00e-19 | 0.000 | 417.5 | $1.02 \mathrm{e}-19$ | 0.000 | 418.0 | $1.00 \mathrm{e}-19$ | 0.000 | 418.5 | 1.02e-19 | 0.000 |
| 419.0 | $1.01 \mathrm{e}-19$ | 0.000 | 419.5 | 1.01e-19 | 0.000 | 420.0 | 1.03e-19 | $0.000$ | 420.5 | 1.01e-19 | 0.000 | 421.0 | $1.04 \mathrm{e}-19$ | 0.000 |
| BACL_ADJ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 230.0 | $1.30 \mathrm{e}-20$ | 1.000 | 232.5 | 1.46e-20 | 1.000 | 235.0 | 1.68e-20 | 1.000 | 237.5 | $1.84 \mathrm{e}-20$ | 1.000 | 240.0 | 2.16e-20 | 1.000 |
| 242.5 | $2.49 \mathrm{e}-20$ | 1.000 | 245.0 | $2.65 \mathrm{e}-20$ | 1.000 | 247.5 | $2.71 \mathrm{e}-20$ | 1.000 | 250.0 | $3.03 \mathrm{e}-20$ | 1.000 | 252.5 | $3.46 \mathrm{e}-20$ | 1.000 |
| 255.0 | 3.46e-20 | 1.000 | 257.5 | $3.57 \mathrm{e}-20$ | 1.000 | 260.0 | $3.95 \mathrm{e}-20$ | 1.000 | 262.5 | $4.17 \mathrm{e}-20$ | 1.000 | 265.0 | $4.17 \mathrm{e}-20$ | 1.000 |
| 267.5 | $4.22 \mathrm{e}-20$ | 1.000 | 270.0 | $4.60 \mathrm{e}-20$ | 1.000 | 272.5 | $4.54 \mathrm{e}-20$ | 1.000 | 275.0 | $4.33 \mathrm{e}-20$ | 1.000 | 277.5 | $4.22 \mathrm{e}-20$ | 1.000 |
| 280.0 | 4.44e-20 | 1.000 | 282.5 | $4.33 \mathrm{e}-20$ | 1.000 | 285.0 | $3.90 \mathrm{e}-20$ | 1.000 | 287.5 | $3.57 \mathrm{e}-20$ | 1.000 | 290.0 | $3.25 \mathrm{e}-20$ | 1.000 |
| 292.5 | $2.92 \mathrm{e}-20$ | 1.000 | 295.0 | $2.60 \mathrm{e}-20$ | 1.000 | 297.5 | $2.16 \mathrm{e}-20$ | 1.000 | 300.0 | $1.79 \mathrm{e}-20$ | 1.000 | 302.5 | $1.73 \mathrm{e}-20$ | 1.000 |
| 305.0 | $1.46 \mathrm{e}-20$ | 1.000 | 307.5 | 1.08e-20 | 1.000 | 310.0 | $9.20 \mathrm{e}-21$ | 1.000 | 312.5 | $7.03 \mathrm{e}-21$ | 1.000 | 315.0 | $6.49 \mathrm{e}-21$ | 1.000 |
| 317.5 | 5.41e-21 | 1.000 | 320.0 | $5.41 \mathrm{e}-21$ | 1.000 | 322.5 | $5.41 \mathrm{e}-21$ | 1.000 | 325.0 | $4.33 \mathrm{e}-21$ | 1.000 | 327.5 | $3.25 \mathrm{e}-21$ | 1.000 |
| 330.0 | $3.79 \mathrm{e}-21$ | 1.000 | 332.5 | $3.79 \mathrm{e}-21$ | 1.000 | 335.0 | $4.33 \mathrm{e}-21$ | 1.000 | 337.5 | $4.87 \mathrm{e}-21$ | 1.000 | 340.0 | $5.41 \mathrm{e}-21$ | 1.000 |
| 342.5 | $5.95 \mathrm{e}-21$ | 1.000 | 345.0 | $6.49 \mathrm{e}-21$ | 1.000 | 347.5 | $7.03 \mathrm{e}-21$ | 1.000 | 350.0 | $8.12 \mathrm{e}-21$ | 0.995 | 352.5 | $7.57 \mathrm{e}-21$ | 0.960 |
| 355.0 | $9.20 \mathrm{e}-21$ | 0.925 | 357.5 | $9.74 \mathrm{e}-21$ | 0.890 | 360.0 | $1.08 \mathrm{e}-20$ | 0.855 | 362.5 | $1.19 \mathrm{e}-20$ | 0.820 | 365.0 | $1.41 \mathrm{e}-20$ | 0.785 |
| 367.5 | $1.51 \mathrm{e}-20$ | 0.750 | 370.0 | $1.79 \mathrm{e}-20$ | 0.715 | 372.5 | $2.00 \mathrm{e}-20$ | 0.680 | 375.0 | $2.11 \mathrm{e}-20$ | 0.645 | 377.5 | $2.33 \mathrm{e}-20$ | 0.610 |
| 380.0 | $2.60 \mathrm{e}-20$ | 0.575 | 382.5 | 2.81e-20 | 0.540 | 385.0 | $3.14 \mathrm{e}-20$ | 0.505 | 387.5 | $3.46 \mathrm{e}-20$ | 0.470 | 390.0 | $3.90 \mathrm{e}-20$ | 0.435 |
| 392.5 | $4.11 \mathrm{e}-20$ | 0.399 | 395.0 | $4.33 \mathrm{e}-20$ | 0.364 | 397.5 | $4.38 \mathrm{e}-20$ | 0.329 | 400.0 | $4.65 \mathrm{e}-20$ | 0.294 | 402.5 | $4.81 \mathrm{e}-20$ | 0.259 |
| 405.0 | $5.19 \mathrm{e}-20$ | 0.224 | 407.5 | $5.84 \mathrm{e}-20$ | 0.189 | 410.0 | $6.06 \mathrm{e}-20$ | 0.154 | 412.5 | $6.49 \mathrm{e}-20$ | 0.119 | 415.0 | $6.92 \mathrm{e}-20$ | 0.084 |
| 417.5 | 6.87e-20 | 0.049 | 420.0 | $6.82 \mathrm{e}-20$ | 0.014 | 422.5 | $6.71 \mathrm{e}-20$ | 0.000 | 425.0 | $6.49 \mathrm{e}-20$ | 0.000 | 427.5 | $5.95 \mathrm{e}-20$ | 0.000 |
| 430.0 | 5.73e-20 | 0.000 | 432.5 | $6.28 \mathrm{e}-20$ | 0.000 | 435.0 | $6.01 \mathrm{e}-20$ | 0.000 | 437.5 | $5.84 \mathrm{e}-20$ | 0.000 | 440.0 | $5.95 \mathrm{e}-20$ | 0.000 |
| 442.5 | $6.49 \mathrm{e}-20$ | 0.000 | 445.0 | $5.95 \mathrm{e}-20$ | 0.000 | 447.5 | $4.98 \mathrm{e}-20$ | 0.000 | 450.0 | $3.79 \mathrm{e}-20$ | 0.000 | 452.5 | 2.81e-20 | 0.000 |
| 455.0 | $1.73 \mathrm{e}-20$ | 0.000 | 457.5 | $1.08 \mathrm{e}-20$ | 0.000 | 460.0 | $5.41 \mathrm{e}-21$ | 0.000 | 462.5 | $3.79 \mathrm{e}-21$ | 0.000 | 465.0 | $2.16 \mathrm{e}-21$ | 0.000 |
| 467.5 | $1.08 \mathrm{e}-21$ | 0.000 | 470.0 | $1.08 \mathrm{e}-21$ | 0.000 | 472.5 |  | 0.000 |  |  |  |  |  |  |
| BZCHO |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 299.0 | $1.78 \mathrm{e}-19$ | 1.000 | 304.0 | $7.40 \mathrm{e}-20$ | 1.000 | 306.0 | 6.91e-20 | 1.000 | 309.0 | $6.41 \mathrm{e}-20$ | 1.000 | 313.0 | 6.91e-20 | 1.000 |
| 314.0 | 6.91e-20 | 1.000 | 318.0 | $6.41 \mathrm{e}-20$ | 1.000 | 325.0 | $8.39 \mathrm{e}-20$ | 1.000 | 332.0 | $7.65 \mathrm{e}-20$ | 1.000 | 338.0 | $8.88 \mathrm{e}-20$ | 1.000 |
| 342.0 | 8.88e-20 | 1.000 | 346.0 | $7.89 \mathrm{e}-20$ | 1.000 | 349.0 | $7.89 \mathrm{e}-20$ | 1.000 | 354.0 | $9.13 \mathrm{e}-20$ | 1.000 | 355.0 | $8.14 \mathrm{e}-20$ | 1.000 |
| 364.0 | 5.67e-20 | 1.000 | 368.0 | 6.66e-20 | 1.000 | 369.0 | $8.39 \mathrm{e}-20$ | 1.000 | 370.0 | $8.39 \mathrm{e}-20$ | 1.000 | 372.0 | $3.45 \mathrm{e}-20$ | 1.000 |
| 374.0 | $3.21 \mathrm{e}-20$ | 1.000 | 376.0 | $2.47 \mathrm{e}-20$ | 1.000 | 377.0 | $2.47 \mathrm{e}-20$ | 1.000 | 380.0 | $3.58 \mathrm{e}-20$ | 1.000 | 382.0 | $9.90 \mathrm{e}-21$ | 1.000 |
| 386.0 | $0.00 \mathrm{e}+00$ | 1.000 |  |  |  |  |  |  |  |  |  |  |  |  |
| ACROLEIN |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 250.0 | $1.80 \mathrm{e}-21$ | 1.000 | 252.0 | $2.05 \mathrm{e}-21$ | 1.000 | 253.0 | $2.20 \mathrm{e}-21$ | 1.000 | 254.0 | 2.32e-21 | 1.000 | 255.0 | $2.45 \mathrm{e}-21$ | 1.000 |
| 256.0 | $2.56 \mathrm{e}-21$ | 1.000 | 257.0 | $2.65 \mathrm{e}-21$ | 1.000 | 258.0 | $2.74 \mathrm{e}-21$ | 1.000 | 259.0 | $2.83 \mathrm{e}-21$ | 1.000 | 260.0 | $2.98 \mathrm{e}-21$ | 1.000 |
| 261.0 | $3.24 \mathrm{e}-21$ | 1.000 | 262.0 | 3.47e-21 | 1.000 | 263.0 | $3.58 \mathrm{e}-21$ | 1.000 | 264.0 | $3.93 \mathrm{e}-21$ | 1.000 | 265.0 | $4.67 \mathrm{e}-21$ | 1.000 |
| 266.0 | $5.10 \mathrm{e}-21$ | 1.000 | 267.0 | $5.38 \mathrm{e}-21$ | 1.000 | 268.0 | $5.73 \mathrm{e}-21$ | 1.000 | 269.0 | $6.13 \mathrm{e}-21$ | 1.000 | 270.0 | $6.64 \mathrm{e}-21$ | 1.000 |
| 271.0 | $7.20 \mathrm{e}-21$ | 1.000 | 272.0 | $7.77 \mathrm{e}-21$ | 1.000 | 273.0 | $8.37 \mathrm{e}-21$ | 1.000 | 274.0 | $8.94 \mathrm{e}-21$ | 1.000 | 275.0 | $9.55 \mathrm{e}-21$ | 1.000 |
| 276.0 | $1.04 \mathrm{e}-20$ | 1.000 | 277.0 | $1.12 \mathrm{e}-20$ | 1.000 | 278.0 | $1.19 \mathrm{e}-20$ | 1.000 | 279.0 | $1.27 \mathrm{e}-20$ | 1.000 | 280.0 | $1.27 \mathrm{e}-20$ | 1.000 |
| 281.0 | $1.26 \mathrm{e}-20$ | 1.000 | 282.0 | $1.26 \mathrm{e}-20$ | 1.000 | 283.0 | $1.28 \mathrm{e}-20$ | 1.000 | 284.0 | $1.33 \mathrm{e}-20$ | 1.000 | 285.0 | $1.38 \mathrm{e}-20$ | 1.000 |
| 286.0 | $1.44 \mathrm{e}-20$ | 1.000 | 287.0 | $1.50 \mathrm{e}-20$ | 1.000 | 288.0 | $1.57 \mathrm{e}-20$ | 1.000 | 289.0 | $1.63 \mathrm{e}-20$ | 1.000 | 290.0 | $1.71 \mathrm{e}-20$ | 1.000 |
| 291.0 | $1.78 \mathrm{e}-20$ | 1.000 | 292.0 | 1.86e-20 | 1.000 | 293.0 | $1.95 \mathrm{e}-20$ | 1.000 | 294.0 | $2.05 \mathrm{e}-20$ | 1.000 | 295.0 | $2.15 \mathrm{e}-20$ | 1.000 |
| 296.0 | $2.26 \mathrm{e}-20$ | 1.000 | 297.0 | $2.37 \mathrm{e}-20$ | 1.000 | 298.0 | $2.48 \mathrm{e}-20$ | 1.000 | 299.0 | $2.60 \mathrm{e}-20$ | 1.000 | 300.0 | $2.73 \mathrm{e}-20$ | 1.000 |
| 301.0 | $2.85 \mathrm{e}-20$ | 1.000 | 302.0 | $2.99 \mathrm{e}-20$ | 1.000 | 303.0 | $3.13 \mathrm{e}-20$ | 1.000 | 304.0 | $3.27 \mathrm{e}-20$ | 1.000 | 305.0 | $3.39 \mathrm{e}-20$ | 1.000 |
| 306.0 | $3.51 \mathrm{e}-20$ | 1.000 | 307.0 | $3.63 \mathrm{e}-20$ | 1.000 | 308.0 | $3.77 \mathrm{e}-20$ | 1.000 | 309.0 | $3.91 \mathrm{e}-20$ | 1.000 | 310.0 | $4.07 \mathrm{e}-20$ | 1.000 |
| 311.0 | $4.25 \mathrm{e}-20$ | 1.000 | 312.0 | $4.39 \mathrm{e}-20$ | 1.000 | 313.0 | $4.44 \mathrm{e}-20$ | 1.000 | 314.0 | $4.50 \mathrm{e}-20$ | 1.000 | 315.0 | $4.59 \mathrm{e}-20$ | 1.000 |
| 316.0 | $4.75 \mathrm{e}-20$ | 1.000 | 317.0 | $4.90 \mathrm{e}-20$ | 1.000 | 318.0 | $5.05 \mathrm{e}-20$ | 1.000 | 319.0 | $5.19 \mathrm{e}-20$ | 1.000 | 320.0 | $5.31 \mathrm{e}-20$ | 1.000 |
| 321.0 | 5.43e-20 | 1.000 | 322.0 | $5.52 \mathrm{e}-20$ | 1.000 | 323.0 | $5.60 \mathrm{e}-20$ | 1.000 | 324.0 | $5.67 \mathrm{e}-20$ | 1.000 | 325.0 | $5.67 \mathrm{e}-20$ | 1.000 |
| 326.0 | 5.62e-20 | 1.000 | 327.0 | $5.63 \mathrm{e}-20$ | 1.000 | 328.0 | $5.71 \mathrm{e}-20$ | 1.000 | 329.0 | $5.76 \mathrm{e}-20$ | 1.000 | 330.0 | $5.80 \mathrm{e}-20$ | 1.000 |
| 331.0 | $5.95 \mathrm{e}-20$ | 1.000 | 332.0 | $6.23 \mathrm{e}-20$ | 1.000 | 333.0 | $6.39 \mathrm{e}-20$ | 1.000 | 334.0 | $6.38 \mathrm{e}-20$ | 1.000 | 335.0 | $6.24 \mathrm{e}-20$ | 1.000 |
| 336.0 | 6.01e-20 | 1.000 | 337.0 | $5.79 \mathrm{e}-20$ | 1.000 | 338.0 | $5.63 \mathrm{e}-20$ | 1.000 | 339.0 | $5.56 \mathrm{e}-20$ | 1.000 | 340.0 | $5.52 \mathrm{e}-20$ | 1.000 |
| 341.0 | $5.54 \mathrm{e}-20$ | 1.000 | 342.0 | $5.53 \mathrm{e}-20$ | 1.000 | 343.0 | $5.47 \mathrm{e}-20$ | 1.000 | 344.0 | $5.41 \mathrm{e}-20$ | 1.000 | 345.0 | $5.40 \mathrm{e}-20$ | 1.000 |
| 346.0 | 5.48e-20 | 1.000 | 347.0 | $5.90 \mathrm{e}-20$ | 1.000 | 348.0 | $6.08 \mathrm{e}-20$ | 1.000 | 349.0 | $6.00 \mathrm{e}-20$ | 1.000 | 350.0 | $5.53 \mathrm{e}-20$ | 1.000 |
| 351.0 | 5.03e-20 | 1.000 | 352.0 | $4.50 \mathrm{e}-20$ | 1.000 | 353.0 | $4.03 \mathrm{e}-20$ | 1.000 | 354.0 | $3.75 \mathrm{e}-20$ | 1.000 | 355.0 | $3.55 \mathrm{e}-20$ | 1.000 |
| 356.0 | $3.45 \mathrm{e}-20$ | 1.000 | 357.0 | $3.46 \mathrm{e}-20$ | 1.000 | 358.0 | $3.49 \mathrm{e}-20$ | 1.000 | 359.0 | $3.41 \mathrm{e}-20$ | 1.000 | 360.0 | $3.23 \mathrm{e}-20$ | 1.000 |
| 361.0 | $2.95 \mathrm{e}-20$ | 1.000 | 362.0 | $2.81 \mathrm{e}-20$ | 1.000 | 363.0 | $2.91 \mathrm{e}-20$ | 1.000 | 364.0 | $3.25 \mathrm{e}-20$ | 1.000 | 365.0 | $3.54 \mathrm{e}-20$ | 1.000 |
| 366.0 | $3.30 \mathrm{e}-20$ | 1.000 | 367.0 | $2.78 \mathrm{e}-20$ | 1.000 | 368.0 | $2.15 \mathrm{e}-20$ | 1.000 | 369.0 | $1.59 \mathrm{e}-20$ | 1.000 | 370.0 | $1.19 \mathrm{e}-20$ | 1.000 |

Table A-3 (continued)

| $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 371.0 | $8.99 \mathrm{e}-21$ | 1.000 | 372.0 | 7.22e-21 | 1.000 | 373.0 | 5.86e-21 | 1.000 | 374.0 | $4.69 \mathrm{e}-21$ | 1.000 | 375.0 | $3.72 \mathrm{e}-21$ | 1.000 |
| 376.0 | $3.57 \mathrm{e}-21$ | 1.000 | 377.0 | $3.55 \mathrm{e}-21$ | 1.000 | 378.0 | $2.83 \mathrm{e}-21$ | 1.000 | 379.0 | $1.69 \mathrm{e}-21$ | 1.000 | 380.0 | $8.29 \mathrm{e}-24$ | 1.000 |
| 381.0 | $0.00 \mathrm{e}+00$ | 1.000 |  |  |  |  |  |  |  |  |  |  |  |  |
| IC3ONO2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 185.0 | $1.79 \mathrm{e}-17$ | 1.000 | 188.0 | $1.81 \mathrm{e}-17$ | 1.000 | 190.0 | $1.79 \mathrm{e}-17$ | 1.000 | 195.0 | $1.61 \mathrm{e}-17$ | 1.000 | 200.0 | 1.26e-17 | 1.000 |
| 205.0 | 8.67e-18 | 1.000 | 210.0 | $4.98 \mathrm{e}-18$ | 1.000 | 215.0 | 2.47e-18 | 1.000 | 220.0 | $1.17 \mathrm{e}-18$ | 1.000 | 225.0 | $5.80 \mathrm{e}-19$ | 1.000 |
| 230.0 | $3.10 \mathrm{e}-19$ | 1.000 | 235.0 | $1.80 \mathrm{e}-19$ | 1.000 | 240.0 | $1.10 \mathrm{e}-19$ | 1.000 | 245.0 | $7.00 \mathrm{e}-20$ | 1.000 | 250.0 | $5.70 \mathrm{e}-20$ | 1.000 |
| 255.0 | $5.20 \mathrm{e}-20$ | 1.000 | 260.0 | $4.90 \mathrm{e}-20$ | 1.000 | 265.0 | $4.60 \mathrm{e}-20$ | 1.000 | 270.0 | $4.10 \mathrm{e}-20$ | 1.000 | 275.0 | $3.60 \mathrm{e}-20$ | 1.000 |
| 280.0 | $2.90 \mathrm{e}-20$ | 1.000 | 285.0 | 2.30e-20 | 1.000 | 290.0 | $1.70 \mathrm{e}-20$ | 1.000 | 295.0 | $1.20 \mathrm{e}-20$ | 1.000 | 300.0 | $8.10 \mathrm{e}-21$ | 1.000 |
| 305.0 | $5.20 \mathrm{e}-21$ | 1.000 | 310.0 | $3.20 \mathrm{e}-21$ | 1.000 | 315.0 | $1.90 \mathrm{e}-21$ | 1.000 | 320.0 | $1.10 \mathrm{e}-21$ | 1.000 | 325.0 | $6.10 \mathrm{e}-22$ | 1.000 |
| 330.0 | $3.70 \mathrm{e}-22$ | 1.000 | 335.0 | $0.00 \mathrm{e}+00$ | 1.000 |  |  |  |  |  |  |  |  |  |
| MGLY_ABS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 219.0 | $9.84 \mathrm{e}-21$ | 1.000 | 219.5 | $1.04 \mathrm{e}-20$ | 1.000 | 220.0 | 1.06e-20 | 1.000 | 220.5 | $1.11 \mathrm{e}-20$ | 1.000 | 221.0 | $1.15 \mathrm{e}-20$ | 1.000 |
| 221.5 | $1.18 \mathrm{e}-20$ | 1.000 | 222.0 | $1.22 \mathrm{e}-20$ | 1.000 | 222.5 | $1.24 \mathrm{e}-20$ | 1.000 | 223.0 | $1.26 \mathrm{e}-20$ | 1.000 | 223.5 | 1.26e-20 | 1.000 |
| 224.0 | $1.25 \mathrm{e}-20$ | 1.000 | 224.5 | $1.24 \mathrm{e}-20$ | 1.000 | 225.0 | $1.25 \mathrm{e}-20$ | 1.000 | 225.5 | $1.27 \mathrm{e}-20$ | 1.000 | 226.0 | $1.27 \mathrm{e}-20$ | 1.000 |
| 226.5 | $1.29 \mathrm{e}-20$ | 1.000 | 227.0 | $1.31 \mathrm{e}-20$ | 1.000 | 227.5 | $1.32 \mathrm{e}-20$ | 1.000 | 228.0 | $1.35 \mathrm{e}-20$ | 1.000 | 228.5 | $1.37 \mathrm{e}-20$ | 1.000 |
| 229.0 | $1.40 \mathrm{e}-20$ | 1.000 | 229.5 | $1.42 \mathrm{e}-20$ | 1.000 | 230.0 | $1.48 \mathrm{e}-20$ | 1.000 | 230.5 | $1.53 \mathrm{e}-20$ | 1.000 | 231.0 | $1.57 \mathrm{e}-20$ | 1.000 |
| 231.5 | $1.59 \mathrm{e}-20$ | 1.000 | 232.0 | $1.61 \mathrm{e}-20$ | 1.000 | 232.5 | $1.62 \mathrm{e}-20$ | 1.000 | 233.0 | $1.61 \mathrm{e}-20$ | 1.000 | 233.5 | $1.68 \mathrm{e}-20$ | 1.000 |
| 234.0 | $1.74 \mathrm{e}-20$ | 1.000 | 234.5 | $1.80 \mathrm{e}-20$ | 1.000 | 235.0 | $1.84 \mathrm{e}-20$ | 1.000 | 235.5 | $1.87 \mathrm{e}-20$ | 1.000 | 236.0 | $1.89 \mathrm{e}-20$ | 1.000 |
| 236.5 | $1.91 \mathrm{e}-20$ | 1.000 | 237.0 | $1.93 \mathrm{e}-20$ | 1.000 | 237.5 | $1.94 \mathrm{e}-20$ | 1.000 | 238.0 | $1.96 \mathrm{e}-20$ | 1.000 | 238.5 | $1.96 \mathrm{e}-20$ | 1.000 |
| 239.0 | $2.01 \mathrm{e}-20$ | 1.000 | 239.5 | 2.04e-20 | 1.000 | 240.0 | 2.08e-20 | 1.000 | 240.5 | $2.10 \mathrm{e}-20$ | 1.000 | 241.0 | 2.14e-20 | 1.000 |
| 241.5 | $2.16 \mathrm{e}-20$ | 1.000 | 242.0 | $2.19 \mathrm{e}-20$ | 1.000 | 242.5 | 2.20e-20 | 1.000 | 243.0 | $2.23 \mathrm{e}-20$ | 1.000 | 243.5 | 2.26e-20 | 1.000 |
| 244.0 | $2.28 \mathrm{e}-20$ | 1.000 | 244.5 | $2.29 \mathrm{e}-20$ | 1.000 | 245.0 | $2.30 \mathrm{e}-20$ | 1.000 | 245.5 | $2.32 \mathrm{e}-20$ | 1.000 | 246.0 | $2.33 \mathrm{e}-20$ | 1.000 |
| 246.5 | $2.35 \mathrm{e}-20$ | 1.000 | 247.0 | 2.38e-20 | 1.000 | 247.5 | $2.41 \mathrm{e}-20$ | 1.000 | 248.0 | $2.46 \mathrm{e}-20$ | 1.000 | 248.5 | $2.51 \mathrm{e}-20$ | 1.000 |
| 249.0 | $2.57 \mathrm{e}-20$ | 1.000 | 249.5 | $2.61 \mathrm{e}-20$ | 1.000 | 250.0 | $2.65 \mathrm{e}-20$ | 1.000 | 250.5 | $2.67 \mathrm{e}-20$ | 1.000 | 251.0 | $2.69 \mathrm{e}-20$ | 1.000 |
| 251.5 | $2.69 \mathrm{e}-20$ | 1.000 | 252.0 | $2.71 \mathrm{e}-20$ | 1.000 | 252.5 | $2.72 \mathrm{e}-20$ | 1.000 | 253.0 | $2.73 \mathrm{e}-20$ | 1.000 | 253.5 | $2.74 \mathrm{e}-20$ | 1.000 |
| 254.0 | $2.76 \mathrm{e}-20$ | 1.000 | 254.5 | $2.78 \mathrm{e}-20$ | 1.000 | 255.0 | $2.82 \mathrm{e}-20$ | 1.000 | 255.5 | $2.87 \mathrm{e}-20$ | 1.000 | 256.0 | $2.93 \mathrm{e}-20$ | 1.000 |
| 256.5 | $2.98 \mathrm{e}-20$ | 1.000 | 257.0 | $3.07 \mathrm{e}-20$ | 1.000 | 257.5 | 3.12e-20 | 1.000 | 258.0 | $3.17 \mathrm{e}-20$ | 1.000 | 258.5 | $3.21 \mathrm{e}-20$ | 1.000 |
| 259.0 | $3.26 \mathrm{e}-20$ | 1.000 | 259.5 | 3.28e-20 | 1.000 | 260.0 | $3.29 \mathrm{e}-20$ | 1.000 | 260.5 | $3.31 \mathrm{e}-20$ | 1.000 | 261.0 | $3.33 \mathrm{e}-20$ | 1.000 |
| 261.5 | $3.34 \mathrm{e}-20$ | 1.000 | 262.0 | 3.36e-20 | 1.000 | 262.5 | $3.38 \mathrm{e}-20$ | 1.000 | 263.0 | $3.42 \mathrm{e}-20$ | 1.000 | 263.5 | $3.44 \mathrm{e}-20$ | 1.000 |
| 264.0 | $3.48 \mathrm{e}-20$ | 1.000 | 264.5 | $3.54 \mathrm{e}-20$ | 1.000 | 265.0 | $3.59 \mathrm{e}-20$ | 1.000 | 265.5 | $3.65 \mathrm{e}-20$ | 1.000 | 266.0 | $3.73 \mathrm{e}-20$ | 1.000 |
| 266.5 | $3.80 \mathrm{e}-20$ | 1.000 | 267.0 | 3.87e-20 | 1.000 | 267.5 | $3.95 \mathrm{e}-20$ | 1.000 | 268.0 | $4.02 \mathrm{e}-20$ | 1.000 | 268.5 | $4.08 \mathrm{e}-20$ | 1.000 |
| 269.0 | $4.13 \mathrm{e}-20$ | 1.000 | 269.5 | $4.17 \mathrm{e}-20$ | 1.000 | 270.0 | $4.20 \mathrm{e}-20$ | 1.000 | 270.5 | $4.22 \mathrm{e}-20$ | 1.000 | 271.0 | $4.22 \mathrm{e}-20$ | 1.000 |
| 271.5 | $4.22 \mathrm{e}-20$ | 1.000 | 272.0 | $4.23 \mathrm{e}-20$ | 1.000 | 272.5 | $4.24 \mathrm{e}-20$ | 1.000 | 273.0 | $4.27 \mathrm{e}-20$ | 1.000 | 273.5 | $4.29 \mathrm{e}-20$ | 1.000 |
| 274.0 | $4.31 \mathrm{e}-20$ | 1.000 | 274.5 | $4.33 \mathrm{e}-20$ | 1.000 | 275.0 | $4.37 \mathrm{e}-20$ | 1.000 | 275.5 | $4.42 \mathrm{e}-20$ | 1.000 | 276.0 | $4.48 \mathrm{e}-20$ | 1.000 |
| 276.5 | $4.56 \mathrm{e}-20$ | 1.000 | 277.0 | $4.64 \mathrm{e}-20$ | 1.000 | 277.5 | $4.71 \mathrm{e}-20$ | 1.000 | 278.0 | $4.78 \mathrm{e}-20$ | 1.000 | 278.5 | $4.83 \mathrm{e}-20$ | 1.000 |
| 279.0 | $4.87 \mathrm{e}-20$ | 1.000 | 279.5 | $4.90 \mathrm{e}-20$ | 1.000 | 280.0 | $4.92 \mathrm{e}-20$ | 1.000 | 280.5 | $4.93 \mathrm{e}-20$ | 1.000 | 281.0 | $4.94 \mathrm{e}-20$ | 1.000 |
| 281.5 | $4.92 \mathrm{e}-20$ | 1.000 | 282.0 | 4.90e-20 | 1.000 | 282.5 | 4.86e-20 | 1.000 | 283.0 | $4.83 \mathrm{e}-20$ | 1.000 | 283.5 | $4.79 \mathrm{e}-20$ | 1.000 |
| 284.0 | $4.76 \mathrm{e}-20$ | 1.000 | 284.5 | 4.72e-20 | 1.000 | 285.0 | $4.70 \mathrm{e}-20$ | 1.000 | 285.5 | $4.68 \mathrm{e}-20$ | 1.000 | 286.0 | $4.66 \mathrm{e}-20$ | 1.000 |
| 286.5 | $4.65 \mathrm{e}-20$ | 1.000 | 287.0 | $4.65 \mathrm{e}-20$ | 1.000 | 287.5 | $4.68 \mathrm{e}-20$ | 1.000 | 288.0 | $4.73 \mathrm{e}-20$ | 1.000 | 288.5 | $4.78 \mathrm{e}-20$ | 1.000 |
| 289.0 | $4.84 \mathrm{e}-20$ | 1.000 | 289.5 | $4.89 \mathrm{e}-20$ | 1.000 | 290.0 | $4.92 \mathrm{e}-20$ | 1.000 | 290.5 | $4.92 \mathrm{e}-20$ | 1.000 | 291.0 | $4.90 \mathrm{e}-20$ | 1.000 |
| 291.5 | $4.86 \mathrm{e}-20$ | 1.000 | 292.0 | 4.81e-20 | 1.000 | 292.5 | $4.75 \mathrm{e}-20$ | 1.000 | 293.0 | $4.70 \mathrm{e}-20$ | 1.000 | 293.5 | $4.65 \mathrm{e}-20$ | 1.000 |
| 294.0 | $4.58 \mathrm{e}-20$ | 1.000 | 294.5 | 4.48e-20 | 1.000 | 295.0 | $4.38 \mathrm{e}-20$ | 1.000 | 295.5 | $4.27 \mathrm{e}-20$ | 1.000 | 296.0 | $4.17 \mathrm{e}-20$ | 1.000 |
| 296.5 | $4.07 \mathrm{e}-20$ | 1.000 | 297.0 | $3.99 \mathrm{e}-20$ | 1.000 | 297.5 | $3.94 \mathrm{e}-20$ | 1.000 | 298.0 | $3.88 \mathrm{e}-20$ | 1.000 | 298.5 | $3.82 \mathrm{e}-20$ | 1.000 |
| 299.0 | $3.76 \mathrm{e}-20$ | 1.000 | 299.5 | 3.72e-20 | 1.000 | 300.0 | $3.69 \mathrm{e}-20$ | 1.000 | 300.5 | $3.68 \mathrm{e}-20$ | 1.000 | 301.0 | $3.70 \mathrm{e}-20$ | 1.000 |
| 301.5 | $3.72 \mathrm{e}-20$ | 1.000 | 302.0 | $3.74 \mathrm{e}-20$ | 1.000 | 302.5 | $3.74 \mathrm{e}-20$ | 1.000 | 303.0 | $3.75 \mathrm{e}-20$ | 1.000 | 303.5 | $3.71 \mathrm{e}-20$ | 1.000 |
| 304.0 | $3.62 \mathrm{e}-20$ | 1.000 | 304.5 | $3.51 \mathrm{e}-20$ | 1.000 | 305.0 | $3.38 \mathrm{e}-20$ | 1.000 | 305.5 | $3.25 \mathrm{e}-20$ | 1.000 | 306.0 | $3.15 \mathrm{e}-20$ | 1.000 |
| 306.5 | $3.04 \mathrm{e}-20$ | 1.000 | 307.0 | $2.92 \mathrm{e}-20$ | 1.000 | 307.5 | $2.80 \mathrm{e}-20$ | 1.000 | 308.0 | $2.71 \mathrm{e}-20$ | 1.000 | 308.5 | $2.63 \mathrm{e}-20$ | 1.000 |
| 309.0 | $2.52 \mathrm{e}-20$ | 1.000 | 309.5 | $2.43 \mathrm{e}-20$ | 1.000 | 310.0 | $2.34 \mathrm{e}-20$ | 1.000 | 310.5 | $2.25 \mathrm{e}-20$ | 1.000 | 311.0 | $2.19 \mathrm{e}-20$ | 1.000 |
| 311.5 | $2.12 \mathrm{e}-20$ | 1.000 | 312.0 | 2.06e-20 | 1.000 | 312.5 | $2.02 \mathrm{e}-20$ | 1.000 | 313.0 | $1.96 \mathrm{e}-20$ | 1.000 | 313.5 | $1.92 \mathrm{e}-20$ | 1.000 |
| 314.0 | $1.91 \mathrm{e}-20$ | 1.000 | 314.5 | $1.88 \mathrm{e}-20$ | 1.000 | 315.0 | 1.86e-20 | 1.000 | 315.5 | $1.85 \mathrm{e}-20$ | 1.000 | 316.0 | 1.86e-20 | 1.000 |
| 316.5 | $1.87 \mathrm{e}-20$ | 1.000 | 317.0 | $1.87 \mathrm{e}-20$ | 1.000 | 317.5 | $1.87 \mathrm{e}-20$ | 1.000 | 318.0 | $1.83 \mathrm{e}-20$ | 1.000 | 318.5 | $1.75 \mathrm{e}-20$ | 1.000 |
| 319.0 | $1.69 \mathrm{e}-20$ | 1.000 | 319.5 | $1.60 \mathrm{e}-20$ | 1.000 | 320.0 | $1.50 \mathrm{e}-20$ | 1.000 | 320.5 | $1.41 \mathrm{e}-20$ | 1.000 | 321.0 | $1.34 \mathrm{e}-20$ | 1.000 |
| 321.5 | $1.27 \mathrm{e}-20$ | 1.000 | 322.0 | $1.21 \mathrm{e}-20$ | 1.000 | 322.5 | $1.18 \mathrm{e}-20$ | 1.000 | 323.0 | $1.14 \mathrm{e}-20$ | 1.000 | 323.5 | $1.08 \mathrm{e}-20$ | 1.000 |
| 324.0 | $1.01 \mathrm{e}-20$ | 1.000 | 324.5 | $9.62 \mathrm{e}-21$ | 1.000 | 325.0 | $9.28 \mathrm{e}-21$ | 1.000 | 325.5 | $8.75 \mathrm{e}-21$ | 1.000 | 326.0 | $8.49 \mathrm{e}-21$ | 1.000 |
| 326.5 | $8.21 \mathrm{e}-21$ | 1.000 | 327.0 | 7.71e-21 | 1.000 | 327.5 | $7.38 \mathrm{e}-21$ | 1.000 | 328.0 | $7.18 \mathrm{e}-21$ | 1.000 | 328.5 | 6.86e-21 | 1.000 |
| 329.0 | $6.71 \mathrm{e}-21$ | 1.000 | 329.5 | $6.63 \mathrm{e}-21$ | 1.000 | 330.0 | 6.46e-21 | 1.000 | 330.5 | $6.29 \mathrm{e}-21$ | 1.000 | 331.0 | $6.21 \mathrm{e}-21$ | 1.000 |
| 331.5 | $6.18 \mathrm{e}-21$ | 1.000 | 332.0 | $6.20 \mathrm{e}-21$ | 1.000 | 332.5 | $5.49 \mathrm{e}-21$ | 1.000 | 333.0 | $5.21 \mathrm{e}-21$ | 1.000 | 333.5 | $5.38 \mathrm{e}-21$ | 1.000 |
| 334.0 | $5.35 \mathrm{e}-21$ | 1.000 | 334.5 | $5.04 \mathrm{e}-21$ | 1.000 | 335.0 | $4.94 \mathrm{e}-21$ | 1.000 | 335.5 | $4.90 \mathrm{e}-21$ | 1.000 | 336.0 | $4.52 \mathrm{e}-21$ | 1.000 |
| 336.5 | $4.26 \mathrm{e}-21$ | 1.000 | 337.0 | $4.11 \mathrm{e}-21$ | 1.000 | 337.5 | 3.76e-21 | 1.000 | 338.0 | $3.61 \mathrm{e}-21$ | 1.000 | 338.5 | $3.58 \mathrm{e}-21$ | 1.000 |
| 339.0 | $3.47 \mathrm{e}-21$ | 1.000 | 339.5 | $3.32 \mathrm{e}-21$ | 1.000 | 340.0 | $3.22 \mathrm{e}-21$ | 1.000 | 340.5 | $3.10 \mathrm{e}-21$ | 1.000 | 341.0 | $3.00 \mathrm{e}-21$ | 1.000 |
| 341.5 | $2.94 \mathrm{e}-21$ | 1.000 | 342.0 | $2.89 \mathrm{e}-21$ | 1.000 | 342.5 | 2.86e-21 | 1.000 | 343.0 | $2.88 \mathrm{e}-21$ | 1.000 | 343.5 | $2.88 \mathrm{e}-21$ | 1.000 |
| 344.0 | $2.89 \mathrm{e}-21$ | 1.000 | 344.5 | $2.91 \mathrm{e}-21$ | 1.000 | 345.0 | $2.95 \mathrm{e}-21$ | 1.000 | 345.5 | $3.00 \mathrm{e}-21$ | 1.000 | 346.0 | $3.08 \mathrm{e}-21$ | 1.000 |
| 346.5 | $3.18 \mathrm{e}-21$ | 1.000 | 347.0 | $3.25 \mathrm{e}-21$ | 1.000 | 347.5 | $3.30 \mathrm{e}-21$ | 1.000 | 348.0 | $3.39 \mathrm{e}-21$ | 1.000 | 348.5 | $3.51 \mathrm{e}-21$ | 1.000 |
| 349.0 | $3.63 \mathrm{e}-21$ | 1.000 | 349.5 | $3.73 \mathrm{e}-21$ | 1.000 | 350.0 | $3.85 \mathrm{e}-21$ | 1.000 | 350.5 | $3.99 \mathrm{e}-21$ | 1.000 | 351.0 | $4.27 \mathrm{e}-21$ | 1.000 |
| 351.5 | $4.47 \mathrm{e}-21$ | 1.000 | 352.0 | $4.63 \mathrm{e}-21$ | 1.000 | 352.5 | $4.78 \mathrm{e}-21$ | 1.000 | 353.0 | $4.92 \mathrm{e}-21$ | 1.000 | 353.5 | $5.07 \mathrm{e}-21$ | 1.000 |
| 354.0 | $5.23 \mathrm{e}-21$ | 1.000 | 354.5 | $5.39 \mathrm{e}-21$ | 1.000 | 355.0 | 5.56e-21 | 1.000 | 355.5 | $5.77 \mathrm{e}-21$ | 1.000 | 356.0 | $5.97 \mathrm{e}-21$ | 1.000 |
| 356.5 | $6.15 \mathrm{e}-21$ | 1.000 | 357.0 | $6.35 \mathrm{e}-21$ | 1.000 | 357.5 | $6.56 \mathrm{e}-21$ | 1.000 | 358.0 | $6.76 \mathrm{e}-21$ | 1.000 | 358.5 | $6.95 \mathrm{e}-21$ | 1.000 |
| 359.0 | $7.20 \mathrm{e}-21$ | 1.000 | 359.5 | $7.44 \mathrm{e}-21$ | 1.000 | 360.0 | $7.64 \mathrm{e}-21$ | 1.000 | 360.5 | $7.89 \mathrm{e}-21$ | 1.000 | 361.0 | $8.15 \mathrm{e}-21$ | 1.000 |
| 361.5 | $8.43 \mathrm{e}-21$ | 1.000 | 362.0 | $8.71 \mathrm{e}-21$ | 1.000 | 362.5 | $9.02 \mathrm{e}-21$ | 1.000 | 363.0 | $9.33 \mathrm{e}-21$ | 1.000 | 363.5 | $9.65 \mathrm{e}-21$ | 1.000 |
| 364.0 | $1.00 \mathrm{e}-20$ | 1.000 | 364.5 | $1.04 \mathrm{e}-20$ | 1.000 | 365.0 | 1.08e-20 | 1.000 | 365.5 | $1.11 \mathrm{e}-20$ | 1.000 | 366.0 | $1.15 \mathrm{e}-20$ | 1.000 |

Table A-3 (continued)

| $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{gathered} \mathrm{Abs} \\ \left(\mathrm{~cm}^{2}\right) \end{gathered}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY | $\begin{aligned} & \text { WL } \\ & (\mathrm{nm}) \end{aligned}$ | $\begin{aligned} & \mathrm{Abs} \\ & \left(\mathrm{~cm}^{2}\right) \end{aligned}$ | QY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 366.5 | $1.19 \mathrm{e}-20$ | 1.000 | 367.0 | $1.23 \mathrm{e}-20$ | 1.000 | 367.5 | 1.27e-20 | 1.000 | 368.0 | $1.31 \mathrm{e}-20$ | 1.000 | 368.5 | $1.35 \mathrm{e}-20$ | 1.000 |
| 369.0 | $1.40 \mathrm{e}-20$ | 1.000 | 369.5 | $1.44 \mathrm{e}-20$ | 1.000 | 370.0 | $1.47 \mathrm{e}-20$ | 1.000 | 370.5 | $1.51 \mathrm{e}-20$ | 1.000 | 371.0 | $1.55 \mathrm{e}-20$ | 1.000 |
| 371.5 | $1.59 \mathrm{e}-20$ | 1.000 | 372.0 | $1.64 \mathrm{e}-20$ | 1.000 | 372.5 | $1.70 \mathrm{e}-20$ | 1.000 | 373.0 | $1.73 \mathrm{e}-20$ | 1.000 | 373.5 | $1.77 \mathrm{e}-20$ | 1.000 |
| 374.0 | $1.81 \mathrm{e}-20$ | 1.000 | 374.5 | $1.86 \mathrm{e}-20$ | 1.000 | 375.0 | $1.90 \mathrm{e}-20$ | 1.000 | 375.5 | $1.96 \mathrm{e}-20$ | 1.000 | 376.0 | $2.02 \mathrm{e}-20$ | 1.000 |
| 376.5 | $2.06 \mathrm{e}-20$ | 1.000 | 377.0 | $2.10 \mathrm{e}-20$ | 1.000 | 377.5 | $2.14 \mathrm{e}-20$ | 1.000 | 378.0 | $2.18 \mathrm{e}-20$ | 1.000 | 378.5 | $2.24 \mathrm{e}-20$ | 1.000 |
| 379.0 | $2.30 \mathrm{e}-20$ | 1.000 | 379.5 | $2.37 \mathrm{e}-20$ | 1.000 | 380.0 | $2.42 \mathrm{e}-20$ | 1.000 | 380.5 | $2.47 \mathrm{e}-20$ | 1.000 | 381.0 | $2.54 \mathrm{e}-20$ | 1.000 |
| 381.5 | $2.62 \mathrm{e}-20$ | 1.000 | 382.0 | $2.69 \mathrm{e}-20$ | 1.000 | 382.5 | $2.79 \mathrm{e}-20$ | 1.000 | 383.0 | $2.88 \mathrm{e}-20$ | 1.000 | 383.5 | $2.96 \mathrm{e}-20$ | 1.000 |
| 384.0 | $3.02 \mathrm{e}-20$ | 1.000 | 384.5 | $3.10 \mathrm{e}-20$ | 1.000 | 385.0 | $3.20 \mathrm{e}-20$ | 1.000 | 385.5 | $3.29 \mathrm{e}-20$ | 1.000 | 386.0 | $3.39 \mathrm{e}-20$ | 1.000 |
| 386.5 | $3.51 \mathrm{e}-20$ | 1.000 | 387.0 | $3.62 \mathrm{e}-20$ | 1.000 | 387.5 | $3.69 \mathrm{e}-20$ | 1.000 | 388.0 | $3.70 \mathrm{e}-20$ | 1.000 | 388.5 | $3.77 \mathrm{e}-20$ | 1.000 |
| 389.0 | $3.88 \mathrm{e}-20$ | 1.000 | 389.5 | $3.97 \mathrm{e}-20$ | 1.000 | 390.0 | $4.03 \mathrm{e}-20$ | 1.000 | 390.5 | $4.12 \mathrm{e}-20$ | 1.000 | 391.0 | $4.22 \mathrm{e}-20$ | 1.000 |
| 391.5 | $4.29 \mathrm{e}-20$ | 1.000 | 392.0 | $4.30 \mathrm{e}-20$ | 1.000 | 392.5 | $4.38 \mathrm{e}-20$ | 1.000 | 393.0 | $4.47 \mathrm{e}-20$ | 1.000 | 393.5 | $4.55 \mathrm{e}-20$ | 1.000 |
| 394.0 | $4.56 \mathrm{e}-20$ | 1.000 | 394.5 | $4.59 \mathrm{e}-20$ | 1.000 | 395.0 | $4.67 \mathrm{e}-20$ | 1.000 | 395.5 | $4.80 \mathrm{e}-20$ | 1.000 | 396.0 | 4.87e-20 | 1.000 |
| 396.5 | 4.96e-20 | 1.000 | 397.0 | $5.08 \mathrm{e}-20$ | 1.000 | 397.5 | $5.19 \mathrm{e}-20$ | 1.000 | 398.0 | $5.23 \mathrm{e}-20$ | 1.000 | 398.5 | $5.39 \mathrm{e}-20$ | 1.000 |
| 399.0 | 5.46e-20 | 1.000 | 399.5 | $5.54 \mathrm{e}-20$ | 1.000 | 400.0 | $5.59 \mathrm{e}-20$ | 1.000 | 400.5 | $5.77 \mathrm{e}-20$ | 1.000 | 401.0 | 5.91e-20 | 1.000 |
| 401.5 | $5.99 \mathrm{e}-20$ | 1.000 | 402.0 | $6.06 \mathrm{e}-20$ | 1.000 | 402.5 | $6.20 \mathrm{e}-20$ | 1.000 | 403.0 | $6.35 \mathrm{e}-20$ | 1.000 | 403.5 | $6.52 \mathrm{e}-20$ | 1.000 |
| 404.0 | $6.54 \mathrm{e}-20$ | 1.000 | 404.5 | $6.64 \mathrm{e}-20$ | 1.000 | 405.0 | $6.93 \mathrm{e}-20$ | 1.000 | 405.5 | $7.15 \mathrm{e}-20$ | 1.000 | 406.0 | $7.19 \mathrm{e}-20$ | 1.000 |
| 406.5 | $7.32 \mathrm{e}-20$ | 1.000 | 407.0 | $7.58 \mathrm{e}-20$ | 1.000 | 407.5 | $7.88 \mathrm{e}-20$ | 1.000 | 408.0 | $7.97 \mathrm{e}-20$ | 1.000 | 408.5 | $7.91 \mathrm{e}-20$ | 1.000 |
| 409.0 | 8.11e-20 | 1.000 | 409.5 | 8.41e-20 | 1.000 | 410.0 | $8.53 \mathrm{e}-20$ | 1.000 | 410.5 | $8.59 \mathrm{e}-20$ | 1.000 | 411.0 | $8.60 \mathrm{e}-20$ | 1.000 |
| 411.5 | $8.80 \mathrm{e}-20$ | 1.000 | 412.0 | $9.04 \mathrm{e}-20$ | 1.000 | 412.5 | $9.45 \mathrm{e}-20$ | 1.000 | 413.0 | $9.34 \mathrm{e}-20$ | 1.000 | 413.5 | $9.37 \mathrm{e}-20$ | 1.000 |
| 414.0 | $9.63 \mathrm{e}-20$ | 1.000 | 414.5 | $9.71 \mathrm{e}-20$ | 1.000 | 415.0 | $9.70 \mathrm{e}-20$ | 1.000 | 415.5 | $9.65 \mathrm{e}-20$ | 1.000 | 416.0 | $9.69 \mathrm{e}-20$ | 1.000 |
| 416.5 | $9.89 \mathrm{e}-20$ | 1.000 | 417.0 | $1.00 \mathrm{e}-19$ | 1.000 | 417.5 | $1.02 \mathrm{e}-19$ | 1.000 | 418.0 | $1.00 \mathrm{e}-19$ | 1.000 | 418.5 | $1.02 \mathrm{e}-19$ | 1.000 |
| 419.0 | $1.01 \mathrm{e}-19$ | 1.000 | 419.5 | 1.01e-19 | 1.000 | 420.0 | $1.03 \mathrm{e}-19$ | 1.000 | 420.5 | $1.01 \mathrm{e}-19$ | 1.000 | 421.0 | $1.04 \mathrm{e}-19$ | 1.000 |
| 421.5 | $1.05 \mathrm{e}-19$ | 1.000 | 422.0 | 1.06e-19 | 1.000 | 422.5 | $1.04 \mathrm{e}-19$ | 1.000 | 423.0 | $1.05 \mathrm{e}-19$ | 1.000 | 423.5 | $1.05 \mathrm{e}-19$ | 1.000 |
| 424.0 | $1.01 \mathrm{e}-19$ | 1.000 | 424.5 | 1.01e-19 | 1.000 | 425.0 | $1.05 \mathrm{e}-19$ | 1.000 | 425.5 | $1.03 \mathrm{e}-19$ | 1.000 | 426.0 | $1.02 \mathrm{e}-19$ | 1.000 |
| 426.5 | $1.01 \mathrm{e}-19$ | 1.000 | 427.0 | $9.77 \mathrm{e}-20$ | 1.000 | 427.5 | $9.81 \mathrm{e}-20$ | 1.000 | 428.0 | $1.00 \mathrm{e}-19$ | 1.000 | 428.5 | $1.02 \mathrm{e}-19$ | 1.000 |
| 429.0 | $9.89 \mathrm{e}-20$ | 1.000 | 429.5 | $9.85 \mathrm{e}-20$ | 1.000 | 430.0 | $1.04 \mathrm{e}-19$ | 1.000 | 430.5 | $1.08 \mathrm{e}-19$ | 1.000 | 431.0 | $1.05 \mathrm{e}-19$ | 1.000 |
| 431.5 | $1.02 \mathrm{e}-19$ | 1.000 | 432.0 | $9.64 \mathrm{e}-20$ | 1.000 | 432.5 | 1.01e-19 | 1.000 | 433.0 | 1.06e-19 | 1.000 | 433.5 | $1.09 \mathrm{e}-19$ | 1.000 |
| 434.0 | $1.04 \mathrm{e}-19$ | 1.000 | 434.5 | $1.03 \mathrm{e}-19$ | 1.000 | 435.0 | $1.07 \mathrm{e}-19$ | 1.000 | 435.5 | 1.16e-19 | 1.000 | 436.0 | $1.09 \mathrm{e}-19$ | 1.000 |
| 436.5 | $1.11 \mathrm{e}-19$ | 1.000 | 437.0 | $9.81 \mathrm{e}-20$ | 1.000 | 437.5 | $9.71 \mathrm{e}-20$ | 1.000 | 438.0 | 1.06e-19 | 1.000 | 438.5 | $1.16 \mathrm{e}-19$ | 1.000 |
| 439.0 | $1.08 \mathrm{e}-19$ | 1.000 | 439.5 | $1.05 \mathrm{e}-19$ | 1.000 | 440.0 | $9.70 \mathrm{e}-20$ | 1.000 | 440.5 | 1.01e-19 | 1.000 | 441.0 | $1.04 \mathrm{e}-19$ | 1.000 |
| 441.5 | $1.07 \mathrm{e}-19$ | 1.000 | 442.0 | $1.02 \mathrm{e}-19$ | 1.000 | 442.5 | $9.68 \mathrm{e}-20$ | 1.000 | 443.0 | $1.00 \mathrm{e}-19$ | 1.000 | 443.5 | $1.14 \mathrm{e}-19$ | 1.000 |
| 444.0 | $1.13 \mathrm{e}-19$ | 1.000 | 444.5 | $1.03 \mathrm{e}-19$ | 1.000 | 445.0 | $9.74 \mathrm{e}-20$ | 1.000 | 445.5 | 8.46e-20 | 1.000 | 446.0 | $8.70 \mathrm{e}-20$ | 1.000 |
| 446.5 | $9.97 \mathrm{e}-20$ | 1.000 | 447.0 | 1.01e-19 | 1.000 | 447.5 | $9.15 \mathrm{e}-20$ | 1.000 | 448.0 | $9.41 \mathrm{e}-20$ | 1.000 | 448.5 | $8.99 \mathrm{e}-20$ | 1.000 |
| 449.0 | $1.10 \mathrm{e}-19$ | 1.000 | 449.5 | $9.12 \mathrm{e}-20$ | 1.000 | 450.0 | 8.56e-20 | 1.000 | 450.5 | $8.28 \mathrm{e}-20$ | 1.000 | 451.0 | $6.15 \mathrm{e}-20$ | 1.000 |
| 451.5 | 5.56e-20 | 1.000 | 452.0 | $6.47 \mathrm{e}-20$ | 1.000 | 452.5 | $7.27 \mathrm{e}-20$ | 1.000 | 453.0 | $5.75 \mathrm{e}-20$ | 1.000 | 453.5 | $5.08 \mathrm{e}-20$ | 1.000 |
| 454.0 | $4.38 \mathrm{e}-20$ | 1.000 | 454.5 | 3.81e-20 | 1.000 | 455.0 | 3.61e-20 | 1.000 | 455.5 | $3.61 \mathrm{e}-20$ | 1.000 | 456.0 | $3.13 \mathrm{e}-20$ | 1.000 |
| 456.5 | $2.72 \mathrm{e}-20$ | 1.000 | 457.0 | $2.44 \mathrm{e}-20$ | 1.000 | 457.5 | $2.22 \mathrm{e}-20$ | 1.000 | 458.0 | $1.82 \mathrm{e}-20$ | 1.000 | 458.5 | $1.43 \mathrm{e}-20$ | 1.000 |
| 459.0 | $1.32 \mathrm{e}-20$ | 1.000 | 459.5 | $1.05 \mathrm{e}-20$ | 1.000 | 460.0 | $8.95 \mathrm{e}-21$ | 1.000 | 460.5 | $8.90 \mathrm{e}-21$ | 1.000 | 461.0 | $7.94 \mathrm{e}-21$ | 1.000 |
| 461.5 | $7.04 \mathrm{e}-21$ | 1.000 | 462.0 | 6.46e-21 | 1.000 | 462.5 | $5.63 \mathrm{e}-21$ | 1.000 | 463.0 | $4.78 \mathrm{e}-21$ | 1.000 | 463.5 | $3.94 \mathrm{e}-21$ | 1.000 |
| 464.0 | $3.26 \mathrm{e}-21$ | 1.000 | 464.5 | $2.97 \mathrm{e}-21$ | 1.000 | 465.0 | $2.65 \mathrm{e}-21$ | 1.000 | 465.5 | $2.46 \mathrm{e}-21$ | 1.000 | 466.0 | $2.27 \mathrm{e}-21$ | 1.000 |
| 466.5 | $2.08 \mathrm{e}-21$ | 1.000 | 467.0 | $1.86 \mathrm{e}-21$ | 1.000 | 467.5 | $1.76 \mathrm{e}-21$ | 1.000 | 468.0 | $1.60 \mathrm{e}-21$ | 1.000 | 468.5 | $1.44 \mathrm{e}-21$ | 1.000 |
| 469.0 | $1.34 \mathrm{e}-21$ | 1.000 | 469.5 | $1.20 \mathrm{e}-21$ | 1.000 | 470.0 | $1.07 \mathrm{e}-21$ | 1.000 | 470.5 | $1.02 \mathrm{e}-21$ | 1.000 | 471.0 | $9.92 \mathrm{e}-22$ | 1.000 |
| 471.5 | $9.97 \mathrm{e}-22$ | 1.000 | 472.0 | 8.87e-22 | 1.000 | 472.5 | 8.27e-22 | 1.000 | 473.0 | $7.76 \mathrm{e}-22$ | 1.000 | 473.5 | $7.15 \mathrm{e}-22$ | 1.000 |
| 474.0 | 6.71e-22 | 1.000 | 474.5 | $6.67 \mathrm{e}-22$ | 1.000 | 475.0 | $6.10 \mathrm{e}-22$ | 1.000 | 475.5 | $6.17 \mathrm{e}-22$ | 1.000 | 476.0 | $5.54 \mathrm{e}-22$ | 1.000 |
| 476.5 | $5.22 \mathrm{e}-22$ | 1.000 | 477.0 | $5.10 \mathrm{e}-22$ | 1.000 | 477.5 | 5.17e-22 | 1.000 | 478.0 | $4.80 \mathrm{e}-22$ | 1.000 | 478.5 | $4.71 \mathrm{e}-22$ | 1.000 |
| 479.0 | $4.60 \mathrm{e}-22$ | 1.000 | 479.5 | $4.35 \mathrm{e}-22$ | 1.000 | 480.0 | $3.90 \mathrm{e}-22$ | 1.000 | 480.5 | $3.71 \mathrm{e}-22$ | 1.000 | 481.0 | $3.62 \mathrm{e}-22$ | 1.000 |
| 481.5 | $3.52 \mathrm{e}-22$ | 1.000 | 482.0 | $3.05 \mathrm{e}-22$ | 1.000 | 482.5 | $3.05 \mathrm{e}-22$ | 1.000 | 483.0 | 2.86e-22 | 1.000 | 483.5 | $2.53 \mathrm{e}-22$ | 1.000 |
| 484.0 | $2.75 \mathrm{e}-22$ | 1.000 | 484.5 | $2.59 \mathrm{e}-22$ | 1.000 | 485.0 | $2.47 \mathrm{e}-22$ | 1.000 | 485.5 | $2.36 \mathrm{e}-22$ | 1.000 | 486.0 | $2.12 \mathrm{e}-22$ | 1.000 |
| 486.5 | $1.89 \mathrm{e}-22$ | 1.000 | 487.0 | $1.93 \mathrm{e}-22$ | 1.000 | 487.5 | 1.86e-22 | 1.000 | 488.0 | $1.82 \mathrm{e}-22$ | 1.000 | 488.5 | $1.75 \mathrm{e}-22$ | 1.000 |
| 489.0 | $1.74 \mathrm{e}-22$ | 1.000 | 489.5 | $1.72 \mathrm{e}-22$ | 1.000 | 490.0 | $1.66 \mathrm{e}-22$ | 1.000 | 490.5 | $1.75 \mathrm{e}-22$ | 1.000 | 491.0 | $1.54 \mathrm{e}-22$ | 1.000 |
| 491.5 | $1.74 \mathrm{e}-22$ | 1.000 | 492.0 | $1.63 \mathrm{e}-22$ | 1.000 | 492.5 | $1.53 \mathrm{e}-22$ | 1.000 | 493.0 | $1.52 \mathrm{e}-22$ | 1.000 | 493.5 | $5.85 \mathrm{e}-23$ | 1.000 |
| 494.0 | $0.00 \mathrm{e}+00$ | 1.000 |  |  |  |  |  |  |  |  |  |  |  |  |

Table A-4. Chamber wall effect and background characterization parameters used in the environmental chamber model simulations for mechanism evaluation.

| Cham. Set [a] | Value | Discussion |
| :---: | :---: | :---: |
| RN-I (ppb) |  | Ratio of the rate of wall + h $\nu$-> HONO to the $\mathrm{NO}_{2}$ photolysis rate. |
| DTC 18 | 0.066 | Average of value of RS-I which gave best fits to n-butane - NOx chamber experiments carried out in this chamber. The initial HONO was optimized at the same time. If a temperature dependence is shown, it was derived from the temperature dependence of the RN-I values that best fit characterization data in outdoor chamber experiments, with the same activation energy used in all cases. If a temperature dependence is not shown, then the temperature variation for experiments in this set is small compared to the run-to-run variability in the best fit RN-I values. Note that the radical source in Sets 3, 12, 13, and 16 runs was anomalously high. Any dependence of apparent radical source on initial NOx levels in Teflon bag chambers was found to be much less than the run-to-run variability. |
| HONO-F (unitless) |  | Ratio of the initial HONO concentration to the measured initial NO2. [The initial NO2 in the experiment is reduced by a factor of $1-$ (HONO-F)]. Unless the characterization data indicate otherwise, it is assumed that the initial HONO is introduced with the NO2 injection, so is it is assumed to be proportional to the initial NO2 concentration. |
| DTC 18 | 0.8\% | Average of value of initial HONO to initial NO2 which gave best fits to nbutane - NOx chamber experiments carried out in this chamber. The RN-I parameter was optimized at the same time. |
| E-NO2/K1 (ppb) |  | Ratio of rate of NO 2 offgasing from the walls to the NO 2 photolysis rate. |
| All Teflon Bag Chambers | 0 | The NOx offgasing caused by representing the radical source by HONO offgasing appears to be sufficient for accounting for NOx offgasing effects in most cases. RN-I parameters adjusted to fit experiments sensitive to the radical source are consistent with NOx offgasing rates adjusted to fit pure air or aldehyde - air runs, to within the uncertainty and variability. |
| $\underline{\mathrm{k}(\mathrm{NO} 2 \mathrm{~W})\left(\mathrm{min}^{-1}\right)}$ |  | Rate of unimolecular loss (or hydrolysis) of NO 2 to the walls. |
| All Teflon Bag Chambers | 1.6e-4 | Based on dark NO2 decay and HONO formation measured in the ETC by Pitts et al. (1984). Assumed to be the same in all Teflon bag chambers, regardless of volume. |
| YHONO |  | Yield of HONO in the unimolecular reaction (hydrolysis) of NO2 on the walls. |
| All Teflon Bag Chambers | 0.2 | Based on dark NO2 decay and HONO formation measured in the ETC by Pitts et al. (1984). Assumed to be the same in all Teflon bag chambers, regardless of volume. |
| $\underline{\mathrm{k}(\mathrm{O} 3 \mathrm{~W})\left(\mathrm{min}^{-1}\right)}$ |  | Unimolecular loss rate of O3 to the walls. |
| DTC All | 1.5e-4 | Based on results of $\mathrm{O}_{3}$ decay in Teflon bag chambers experiments as discussed by Carter et al (1995d). |
| $\underline{\mathrm{k}(\mathrm{N} 26 \mathrm{I})\left(\mathrm{min}^{-1}\right)}$ |  | Rate constant for $\mathbf{N} 2 \mathrm{O5}$-> 2 Wall-NOx. This represents the humidityindependent portion of the wall loss of $\mathrm{N}_{2} \mathrm{O}_{5}$, or the intercept of plots of rates of $\mathrm{N}_{2} \mathrm{O}_{5}$ loss against humidity. |
| All Teflon Bag Chambers | $2.8 \mathrm{e}-3$ | Based on $\mathrm{N}_{2} \mathrm{O}_{5}$ decay rate measurements made by Tuazon et al (1983) for the ETC. Assumed to be independent of chamber size (Carter et al, 1995d). |

Table A-4 (continued)

| Cham. Set [a] | Value | Discussion |
| :---: | :---: | :---: |
| $\underline{\mathrm{k}(\mathrm{N} 26 \mathrm{~S})\left(\mathrm{ppm}^{-1} \mathrm{~min}^{-1}\right)}$ |  | Rate constant for $\mathbf{N 2 O 5}+\mathbf{H 2 O}$-> $\mathbf{2}$ Wall-NOx. This represents the humidity dependent portion of the wall loss of $\mathrm{N}_{2} \mathrm{O}_{5}$, or the slope of plots of rates of $\mathrm{N}_{2} \mathrm{O}_{5}$ loss against humidity. |
| All Teflon Bag Chambers | $1.1 \mathrm{e}-6$ | Based on $\mathrm{N}_{2} \mathrm{O}_{5}$ decay rate measurements made by Tuazon et al (1983) for the ETC. Assumed to be independent of chamber size (Carter et al, 1995d). |
| $\underline{\mathrm{k}(\mathrm{XSHC})\left(\mathrm{min}^{-1}\right)}$ |  | Rate constant for $\mathbf{O H}$-> HO2. This represents the effects of reaction of OH with reactive VOCs in the background air or offgased from the chamber walls. This parameter does not significantly affect model simulations of experiments other than pure air runs. |
| All Teflon Bag Chambers | 250 | Estimated from modeling several pure air in the ITC (Carter et al, 1996d), and also consistent with simulations of pure air runs in the ETC (Carter et al, 1997a). |
| $\underline{\mathrm{H} 2 \mathrm{O}}$ (ppm) |  | Default water vapor concentration for runs where no humidity data are available. |
| DTC all | $1.0 \mathrm{e}+3$ | Experiments in this chamber were carried out using dried purified air. The limited humidity data for such runs indicate that the humidity was less than $5 \%$, probably no more than $\sim 2.5 \%$, and possibly much less than that. The default value corresponds to $\sim 2.5-3 \% \mathrm{RH}$ for the conditions of most experiments. |

[a] Set refers to the characterization set, which refers to the group of experiments assumed to have the same run conditions and represented using the same chamber-dependent parameters. See Carter et al (1995) for more discussion. All experiments in this program were in DTC characterization set 18.


[^0]:    ${ }^{1}$ Note that this differs from how the term "incremental reactivity" is used in the context of chamber experiments. In that case, the incremental reactivity refers to the relative change observed in the individual experiments, which in general depends on the amount added.

